# FINAL REPORT

Performance Assessment of Past Bioremediation Approaches for Chlorinated Solvent Source Zones

# ESTCP Project ER-201427



#### DECEMBER 2020

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# ACRONYMS AND ABBREVIATIONS

%0	per mil; part per thousand
%	percent
AFB	Air Force base
ANOVA	analysis of variance
AOC	area of concern
APTIM	Aptim Federal Services, LLC
ATL	APTIM's Analytical and Treatability Laboratory
Bgs	below ground surface
C	carbon
°C	degrees Celsius
CAFB	Charleston Air Force Base
CB&I	Chicago Bridge & Iron Company Federal Services, LLC
CE	chlorinated ethenes
CFR	Code of Federal Regulations
CI	chlorine
COC	chain of custody
CSIA	compound-specific isotope analysis
DCE or <i>cis</i> -DCE	<i>cis</i> -1,2-dichloroethene
1,1-DCE	1,1-dichloroethene
1,2-DCA	1,2-dichloroethane
δ <sup>13</sup> C	delta carbon 13, a measure of the ratio of carbon 13 to carbon 12
DHC	<i>Dehalococcoides mccartyi</i>
DNA	deoxyribonucleic acid
DNAPL	dense non-aqueous phase liquid
DO	dissolved oxygen
DOC	dissolved organic carbon
DPT	direct push technology
DoD	Department of Defense
dPCR	digital polymerase chain reaction
EAB EDD EMD EPA ERIS ERPIMS ESTCP EVO Fe	enhanced anaerobic bioremediation electronic data deliverables Environmental Molecular Diagnostics Environmental Protection Agency Environmental Restoration Information System Environmental Resources Program Information Management System Environmental Security Technology Certification Program emulsified vegetable oil
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GC	gas chromatography
H2	hydrogen
H3PO4	phosphoric acid
HCl	hydrochloric acid
HNO3	nitric acid
HPLC	high-performance liquid chromatography
IPR	in progress review
ITRC	Interstate Technology & Regulatory Council
IW	injection well
KB-1	<i>Dehalococcoides</i> containing bioaugmentation culture
KW	Kruskal-Wallace test
L	liter
µg/L	micrograms per liter
M	mass
MCL	Maximum Contaminant Level
MBT	molecular biological tool
MEE	methane, ethane, ethene
Mg	milligrams
Mg/L	milligrams per liter
mL	milliliter
mL/min	milliliters per minute
Mn	manganese
NAPL	non-aqueous phase liquid
NB	naval base
NIRIS	Navy Installation Restoration Information Solution
nL	nanoliter(s)
nM	nanomoles
NO <sup>3-</sup>	nitrate
NRC	National Research Council
NS	Naval Station
NWS	Naval Weapons Station
O2	oxygen
O&M	Operation & Maintenance
ORP	oxidation-reduction potential
PCE	tetrachloroethene
PFM	passive flux meter
pH	hydrogen potential
PI	principal investigator
PM	project manager

QA	quality assurance
QC	quality control
qPCR	quantitative polymerase chain reaction
recirc	recirculation of groundwater
RD-qChip	real-time polymerase chain reaction chip
RPM	Remedial Project Manager
rRNA	ribosomal ribonucleic acid
SDC-9	<i>Dehalococcoides</i> containing bioaugmentation culture
SERDP	Strategic Environmental Research and Development Program
SO4 <sup>2-</sup>	sulfate
SOP	standard operating procedure
SQL	Structured Query Language
SRO	Spearman Rank Order
T	temperature
1,1,1-TCA	1,1,1-trichloroethane
1,1,2-TCE	1,1,2-trichloroethane
TCE	trichloroethene
tDCE	trans-1,2-dichloroethene
TDS	total dissolved solids
TOC	total organic carbon
USAF	United States Air Force
USEPA	United States Environmental Protection Agency
UST	underground storage tank
VC	vinyl chloride
VOC	chlorinated volatile organic compounds
VFA	volatile fatty acid
VOA	volatile organic analysis
VOC	volatile organic compound
WRS	Wilcoxon Rank Sum test

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# **EXECUTIVE SUMMARY**

#### **INTRODUCTION**

Source areas of chlorinated volatile organic compounds (VOCs) in groundwater create and perpetuate dilute groundwater plumes that subsequently pose risks to downgradient receptors for decades or centuries. Although bioremediation has been applied to treat many contaminant source areas over the past two decades, the overall success of this treatment and factors that differentiate successful from unsuccessful treatment applications have not been thoroughly assessed. We used two separate approaches to evaluate the success or failure of different bioremediation applications and to evaluate factors that may have contributed to those outcomes.

Initially, we selected fifteen (15) VOC sites with chlorinated ethenes as the primary contaminant that have been treated using common bioremediation techniques and with data that allow statistical evaluation of remedial performance over time. Sites with extensive pre- and post-treatment data were given priority, and overburden aquifers were selected over bedrock aquifers. A large database was developed from these sites that includes 1) site location; 2) VOC concentrations over time; 3) hydrogeology; 4) geochemistry; 5) water chemistry; 6) abundances of relevant microbial biomarkers where available; 7) treatment approach; and 8) other relevant site data. A large database was developed and statistical analyses were then performed to identify factors that may promote or prevent successful application of different bioremediation strategies.

From the 15 selected sites for historical data compilation and statistical analysis, 5 sites were chosen for sampling of select wells and application of current assessment tools in order to quantify (1) contaminant mass flux, (2) presence of key dehalogenating organisms and genes, and (3) whether biodegradation is still occurring using parent and daughter product VOC concentrations and compound-specific isotope analysis (CSIA). Pre-treatment data and post-treatment data were compared with information on site conditions (geologic, hydrogeologic, and geochemical) and treatment methods. The historical and current data were used to draw conclusions about the long-term effectiveness of VOC bioremediation at the 5 sites.

#### **OBJECTIVE**

The primary project objective was to assess the long-term effectiveness of past biological treatment approaches for remediating source areas of chlorinated ethene-contaminated aquifers. In support of this objective the project team:

- Developed a large database and statistically evaluated the effectiveness of source area bioremediation to identify factors that resulted in, or prevented, successful application of different bioremediation strategies;
- Conducted sampling at select sites to obtain a current snapshot of the effectiveness of past treatment approaches using traditional measures (e.g., contaminant concentrations) and new tools, including analysis of key dehalogenating organisms/genes; passive flux meters to assess current contaminant flux; and CSIA to quantify whether degradation is ongoing

#### **TECHNOLOGY DESCRIPTION**

#### Database and Statistical Analyses

The multi-site evaluation included gathering of historical analytical data and information on site conditions and the treatments applied. Electronic data were obtained for each of the evaluation sites whenever possible. Data was also gathered from hard copy sources when electronic data were not available. Key sources of electronic data included the following databases:

- The Environmental Restoration Information System (ERIS) is the Army's web-based database system for the storage of Army environmental restoration field data. It serves as a central repository for the Army installation chemical, geological, and geographical data.
- The Environmental Resources Program Information Management System (ERPIMS) is the Air Force system for management of data from environmental projects at Air Force installations. This data contains analytical chemistry samples, tests, and results, as well as hydrogeological information, site/location descriptions, and monitoring well characteristics.
- The Navy Installation Restoration Information Solution (NIRIS) provides analytical data along with data management and visualization tools. NIRIS also includes a collection of site documents.

Documents generated as part of site remediation activities were obtained from web-based document-sharing sites, where available, and through direct requests to installation Remedial Project Managers (RPMs). Site reports were reviewed to obtain details on hydrogeologic conditions and the treatments applied at each site. Typical documents with site information included, for example, Remedial Designs, Remedial Action Workplans, Remedial Action Completion Reports and Monitoring and Progress Reports.

Because of the large amount of existing information and data collected and analyzed during this project, a database was created to organize and evaluate the data. Data were housed in a SQL Server database. The SQL server provided a back-end data repository for analytical data and was managed by a front-end Microsoft Access database. Data added to the Microsoft Access database included: 1) site location; 2) VOC concentrations; 3) hydrogeology (e.g., groundwater velocity, groundwater gradient, porosity, etc.); 4) geochemistry; 5) water chemistry (pH, Eh, anions, cations, metals, VFAs, co-contaminants, DOC, TOC, methane, ethene, ethane, acetylene, etc.); 6) abundances of relevant microbial biomarkers (e.g., *Dehalococcoides* numbers, reductive dehalogenase genes); 7) treatment approach (bioaugmentation, biostimulation, active, passive); and 8) other relevant site information.

The data analysis included a statistical evaluation comparing measures of remedial effectiveness to a number of possible factors. The evaluation was conducted on a well-by-well basis for each of the treatment zones and treatment phases at the 15 evaluation sites. Measures of effectiveness were developed to provide quantitative metrics for the statistical evaluation. The measures were as follows: *Mass Reduction, Dechlorination* (mols of chlorine removed from initial chlorinated ethenes), *Risk Reduction* (MCLs of initial and residual VOCs), *Rebound, DCE Accumulation, and VC Accumulation.* These measures were compared to a number of potential factors for success including *Hydrogeologic factors* (percent fines, heterogeneity, hydraulic conductivity, hydraulic

gradient and seepage velocity); *Treatment Approach factors* (substrate used, substrate loading, substrate dosing, nutrient amendments, treatment approach and spacing of injection points); *Geochemical factors* (DO, ORP, pH, alkalinity, iron , manganese, sulfate, sulfide, TOC, temperature); and *Biological factors* (abundance of *Dehalococcoides*, abundance of reductase genetic markers).

Correlations between selected pairs of numerical parameters were evaluated using the Spearman Rank Order (SRO) procedure, which is a nonparametric method to evaluate statistical dependence between two variables. It assesses how well the relationship between two variables can be described using a monotonic function. The coefficient R ranges from -1.0 to +1.0. The SRO procedure also returns a p-value, which allows the results to be interpreted in terms of the statistical confidence in any apparent correlation. The nonparametric SRO approach was selected here because there is no expectation that any of the parameters considered in this evaluation will be linearly related to each other, but at least some monotonic relationships were expected.

Some of the parameters in the database are categorical (non-numeric) so their effects on performance cannot be evaluated with the SRO method. The relationships between these parameters (e.g., presence of DNAPL: Y/N) versus performance parameters were evaluated using box plots for visual comparisons of the distributions within each individual category versus a performance parameter, and the quantitative Wilcoxon Rank Sum (WRS) test or the Kruskal-Wallis (KW) test. Both nonparametric procedures test the null hypothesis that the sub-groups of results being compared are drawn from the same population. Both tests also return a p-value that indicates the probability of the null hypothesis (no difference between sub-groups) being true.

The software package *Statistica* (version 12) was used to evaluate the correlations between selected parameters in the database using SRO correlations, KW tests, and WRS tests; and was also used to generate the box plots.

#### **Bioremediation Performance Evaluation at Five Sites**

Five field sites were selected for additional field investigations to assess current post-treatment conditions. The investigations included groundwater sampling from existing monitoring wells and analysis of a variety of parameters. The five field sites were as follows:

- Former Raritan Arsenal, AOC 2
- Dover Air Force Base, Area 6
- Charleston Air Force Base, Zone 4
- Seal Beach Naval Weapons Station, Site 70
- Treasure Island Naval Station, Site 24

The objective of the groundwater sampling was to assess the long-term impacts of the previously performed biostimulation and/or bioaugmentation treatments on contaminant concentrations, groundwater quality, and biogeochemistry. Field activities at each site included the collection of groundwater samples and installation of passive flux meters (PFMs) in existing monitoring wells at each of the five field sites.

Groundwater samples were collected at monitoring wells variously located within treatment zones and at downgradient locations. Wells were selected based on treatment effectiveness and include sites with apparent successful and unsuccessful bioremediation applications at different times in the past. The analysis included the following parameters:

#### <u>VOCs</u>

VOCs analysis by EPA Method 8260 at APTIM's Analytical and Treatability Lab (ATL) was used to assess the extent of treatment and/or rebound, and the extent to which reductive dechlorination has occurred/continued to occur since treatment implementation.

#### Reduced Gases

Reduced gases, including methane, ethane, and ethene (MEE) measured using EPA Method 3810, RSK-175 were used to assess the extent to which complete reductive dechlorination is still occurring within the treatment area. Methane data also provide insight into the redox conditions.

#### <u>Anions</u>

Anions, including chloride, nitrate, and sulfate, measured by EPA Method 300.0, provide insight into the current redox conditions and the dominant electron acceptors. Total sulfate plays a significant role in rates/extents of reductive dechlorination at specific sites or locations at a given site. Chloride levels provide a measure of salinity at each site.

#### Dissolved Iron and Manganese

Dissolved iron and manganese levels were measured using EPA Method 601.0. These data provide insight into the current redox conditions and dominant electron acceptors and may also provide useful information about potential abiotic reactions.

#### Volatile Fatty Acids

Volatile fatty acids (VFAs), measured by EPA Method 300m, are commonly present during active bioremediation treatment. Groundwater samples were analyzed to determine if residual levels persist, possibly serving as electron donors.

#### Dissolved Hydrogen

Dissolved hydrogen levels, measured using EPA Method 3810, are used to assess whether hydrogen is present at detectable levels at any site months to several years after biological treatment, as an indication of ongoing fermentation.

#### Total Organic Carbon

Total Organic Carbon (TOC) levels, measured using EPA Method 415.1, indicate available organic carbon that could support reductive dechlorination.

#### Field Parameters

Dissolved oxygen (DO), pH, turbidity, and ORP were measured as indicators of geochemical conditions using a standard multi-parameter field meter. Values were recorded during low-flow sampling after each well had stabilized.

#### Microbial Community Analysis by qPCR

A targeted microbial community analysis was conducted by qPCR to determine the presence and abundance of organohalide-respiring bacteria including *Dehalococcoides*, *Dehalobacter*, and *Dehalogenimonas*. Key functional genes involved in reductive dechlorination of chlorinated solvents were also quantified. The presence of dehalogenating organisms and genes was evaluated as a measure of current day biodegradation potential. Analyses were conducted by Dr. Frank Löffler at the University of Tennessee.

#### <u>CSIA</u>

CSIA analysis was performed for carbon in TCE, cis-DCE, and VC as a means to assess whether VOC degradation is occurring or has occurred in the recent past based on current values of  $\delta^{13}$ C. The measured  $\delta^{13}$ C value for TCE were evaluated as a function of (1) historical values of  $\delta^{13}$ C in manufactured TCE (generally -30 ± 5 ‰); (2) previous CSIA analyses conducted at the site when available and/or (3) distance from the source area. The analysis conducted was dependent on the current and historical data available. The analysis was done by Pace Analytical, Inc. (Pittsburgh, PA).

#### Mass Flux

Contaminant mass flux was estimated at the different sites using passive flux meters (PFMs). PFMs are self-contained, permeable units that are inserted into a well or subsurface boring such that they intercept, but do not retain, groundwater flow. PFMs allow measurements of both groundwater flow and contaminant mass flux. The mass of contaminant accumulated on a permeable sorbent (typically granular activated carbon) is used to calculate a time-averaged contaminant mass flux, while residual resident tracer mass loss is used to calculate cumulative water flux. VOC concentrations also can be estimated by this procedure. Mass flux measurements were collected and analyzed by Drs Mike Annable and Alexander Haluska at the University of Florida.

#### PERFORMANCE ASSESSMENT

#### Statistical Analysis of Database

The parameter *Concentration Reduction*, equal to the greatest reduction in total chlorinated ethene concentration for each well during the treatment period prior to any rebound, was selected as a key performance parameter for the majority of the statistical evaluations. Rebound parameters (*Rebound* and *cis-DCE Accumulation*) were also evaluated.

Numerical site characteristics that showed statistically significant ( $\alpha$ = 0.05) positive (+) or negative (-) correlations with *Concentration Reduction* include *Min Sulfate* (-), *Sulfate Depletion* (+), *Initial ORP* (-), *Max Iron* (+), *Fines* (-), *Initial Sulfate* (-), *Initial CE* (+), *Min ORP* (-), and *DNAPL 1/0* (+) in decreasing order. The strongest correlations were with *Min Sulfate* (-) and *Sulfate Depletion* (+), suggesting that creating strongly reducing redox conditions that reduce sulfate concentrations as low as possible will contribute to a more successful remediation. The third and fourth most significant correlations were with *Initial ORP* (-) and *Max Iron* (+) demonstrating that either initial or minimum redox potentials are strong predictors of successful concentrations. These first four strongest correlations provide independent evidence that creating redox conditions corresponding to sulfate-reducing (or lower) redox potentials is beneficial for reducing chlorinated ethene concentrations.

*Initial Sulfate* had a significant negative correlation with *Concentration Reduction*, indicating that high initial sulfate is a predictor of poor performance. High initial sulfate may be present at coastal sites or at arid sites where gypsum layers are present in the treatment zone. *Fines* also had a significant negative correlation with *Concentration Reduction*. This observation suggests that it may be more difficult to lower chlorinated ethene concentrations in treatment zones that have a high proportion of finer-grained sediments. Diffusion of chlorinated ethenes into the fine-grained units over time within the treatment zone may extend the time required for treatment and may also increase the likelihood of rebound as the compounds slowly diffuse back out of the fine-grained units after treatment has stopped. The presence of fine-grained layers can also prevent adequate distribution of amendments in the subsurface.

Cases where DNAPL was present show significantly greater concentration reduction than cases where DNAPL was not present. This is likely caused by higher initial chlorinated ethene concentrations were DNAPL was present. The significant correlations between *Concentration Reduction* versus *Initial CE* (+) and *DNAPL 1/0* (+) likely reflect the fact that chlorinated ethene concentrations can only show strong decreases if they are initially high.

Numerical parameters that contribute to *Rebound* include *Min Sulfate* (+) and *Initial ORP* (+), suggesting that high initial ORP is a predictor of rebound. In addition, failure to adequately reduce sulfate to low concentrations in the groundwater during treatment may also lead to rebound. A positive correlation with the categorical *Fines Rank* parameter indicates that a high proportion of fines in the treatment zone is a predictor of rebound, as well as poor reduction in chlorinated ethene concentrations, as noted above. In addition, sites that have undergone previous remediation show significantly less rebound than sites that have not had previous remediation.

Numerical site characteristics that have significant correlations with *cis-DCE Accumulation* are *Minimum Sulfate* (+), *Initial Iron* (-), *Initial Sulfate* (+), *Heterogeneity* (-), and *Initial ORP* (+) in decreasing order. These results suggest that high initial redox conditions, as indicated by high initial ORP and low initial iron, are predictors of cis-DCE accumulation. The use of recirculation loops was observed to correlate negatively with accumulation of cis-DCE, most likely by allowing more effective mixing of amendments in the subsurface. Initial pH below 5.0 is a predictor of cis-DCE accumulation presumably by causing inhibition of complete dehalogenation.

In summary, the key parameters that control chlorinated ethene reduction are similar to the parameters that control rebound. These parameters are as follows:

<u>Redox conditions</u> – The strongest predictors of chlorinated ethene concentration reduction and rebound are *Initial Sulfate, Min Sulfate*, and *Sulfate Depletion*. These results demonstrate that it is essential to establish sulfate-reducing (or lower) redox potentials for successful remediation. Correlations of these performance measures with other redox-related parameters such as *Initial ORP, Min ORP*, and *Initial Iron* provide independent evidence for these predictors. High initial sulfate or lack of sufficient sulfate depletion may prevent successful remediation if there is an ongoing source of sulfate during the treatment period that prevents redox conditions from falling below sulfate-reducing potentials. Sources of sulfate may include seawater intrusion, arid climate (where sulfate groundwater concentrations are typically high), or the presence of gypsum in the treatment zone.

<u>Fines</u> – The presence of fines in the treatment zone is correlated with poor concentration reduction, as well as rebound and cis-DCE accumulation. These results provide independent evidence that the presence of fine-grained layers in the treatment zone interfere with performance, likely by allowing diffusion of chlorinated ethenes into the fine layers over time. During remediation when the chlorinated ethenes concentrations decrease in groundwater along interconnected flow paths, the local concentration gradients reverse, and chlorinated ethenes that had diffused into the fine layers slowly diffuses back out, thus contributing to rebound. In addition, the presence of fine-grained layers can interfere with the distribution of amendments in the subsurface.

<u>Initial pH</u> – Initial pH below 5.0 is a predictor of poor chlorinated ethenes concentration reduction and is also a predictor of cis-DCE accumulation. These effects are likely caused by inhibition of microbial activity.

<u>Previous remediation</u> – Cases that have undergone previous remediation have significantly higher CE reduction and significantly less rebound and cis-DCE accumulation. These results highlight the need for multiple rounds of remediation at some locations.

<u>Presence of DNAPL</u> – One surprising result was that cases where DNAPL was present had significantly greater reductions in chlorinated ethenes concentrations than cases where DNAPL was absent. This result is likely due to the fact that sites with DNAPL usually have higher initial CE concentrations relative to sites without DNAPL, so the sites with DNAPL tend to show a greater decrease during remediation. (The presence of DNAPL as the parameter *DNAPL 1/0* is positively correlated with *Initial CE, Min CE, Max CE* and *Final CE*.) These results also suggest that the presence of DNAPL at a site does not interfere with reductions in CE concentrations during remediation. The degree of rebound and cis-DCE accumulation were insensitive to the presence or absence of DNAPL.

#### Current Assessment

Summary results from the five different sites are provided below:

#### Former Raritan Arsenal, AOC 2

Site AOC 2 at the former Raritan Arsenal previously had a TCE plume that extended approximately 3,000 ft downgradient. Excavation of contaminated soil occurred in 1998 and 2002. Bioremediation was implemented in the source area in 2009 to reduce contaminant concentrations. The system included recirculation with sodium lactate and nutrients along with the SDC-9 bioaugmentation culture in a deep aquifer zone. Sodium and potassium hydroxide were used to raise the pH. The recirculation system operated for a 10-month period. Near the end of the recirculation period, an emulsified oil substrate was introduced to provide a longer-term, slow release of electron donor to promote continuing biodegradation. A shallower contaminated zone was treated with emulsified oil and bioaugmentation culture SDC-9 using direct push. The site was sampled ~ 5.5 years after treatment ended.

Groundwater monitoring, CSIA and PFM data collected during this project suggest that the size of the plume at Raritan has decreased and biodegradation continues 5.5 years after active treatment ended. Source zone groundwater and PFM data show that cis-DCE and VC dominate the source area but dramatic declines in the concentration and flux of both VOCs are apparent downgradient.

CSIA data provides further evidence of dechlorination of TCE to cis-DCE and then cis-DCE to VC. Based on the simple calculations using  $\delta^{13}$ C values for TCE and cis-DCE along the groundwater flow path, half-lives for the two compounds were estimated at ~ 80 days for TCE and 30 days for cis-DCE. The microbial community analysis indicated the presence of several different dehalogenation- associated organisms/genes in water samples from the Raritan Site 5.5 years after cessation of active treatment. Dehalococcoides was detected in all of the aforementioned wells, albeit at rather low concentrations ( $\leq 1 \times 10^2$  cells/mL). The vcrA gene and the bvcA gene (both of which encode enzymes that dehalogenate cis-DCE) also were detected as was the cobS gene, which encodes a vitamin B12 pathway in Dehalococcoides. The data suggest that dehalogenating organisms/genes are present in the deep aquifer zone more than 5.5 years after the last treatment, which included bioaugmentation with SDC-9. This zone also contains the highest concentration of residual VOCs at the site. Overall, site data suggest that removal of organic contaminants was highly effective and is still occurring where some residual persists in the deep zone (shallow zone is clean) and that the potential for anaerobic reductive dechlorination persists, although electron donor is likely limiting at his time. Target treatments with an electron donor, buffer (to ensure optimal pH) and nutrients may be sufficient to promote conditions necessary for complete reductive dechlorination in the deep zone at the site.

#### Dover Air Force Base, Area 6

Area 6 at Dover Air Force Base was previously characterized by a large plume of chlorinated ethenes and ethanes, approximately 2,500 ft wide and 6,500 ft long. Several source areas were identified. Maximum TCE concentrations historically exceeded 20,000  $\mu$ g/L. PCE, cis-DCE, and VC were present at lesser concentrations. A recirculation system for bioremediation was actively operated at Building 719 between 2002 and 2006. Sodium lactate and diammonium phosphate (DAP) were added as biostimulation amendments. The system was converted to a passive mode in 2007 with recirculation temporarily restarted to introduce additional amendments. The system was operated for this purpose on three occasions: September 2008, June 2011 and February 2013.

The data collected from the site by our project team indicate highly effective bioremediation of chlorinated ethenes across the site. Low concentrations of chlorinated ethenes persist in a few of the wells, possibly as a result of diffusion from low permeability materials, but concentrations are orders of magnitude lower than pre-treatment. In wells MW605S, MW605D, and MW608S, *Dehalococcoides* was detected at concentrations ranging from 3 x 10<sup>3</sup> to 8 x 10<sup>4</sup> cells/mL, methane concentrations were relatively high (ranging from 19.1 to 22.5 mg/L) and the presence of ethene and low ORP suggest that conditions are conducive to ongoing reductive dehalogenation. CSIA data supported this conclusion in MW608S (with VOC concentrations too low to measure in the other two wells). Conversely, in wells MW102S and MW102D, *Dehalococcoides* were not detected, ORP values were positive, methane concentrations were low, and no ethane or ethene were detected, suggesting the conditions were not conducive to reductive dechlorination at these well. However, MW102S had no residual VOCs. MW102D had very low concentrations of PCE, TCE, and cis-DCE. Overall, the data from this site suggest that treatment has been highly effective and that reductive dechlorination at the site continues in many areas 5 years after cessation of active treatment.

#### Charleston Air Force Base, Zone 4

Zone 4 at Charleston Air Force Base (CAFB) is an industrial area near the CAFB flight line. Trichloroethene (TCE) was formerly used for aircraft parts cleaning. Two distinct plumes of TCE contamination, designated Plume 1 and Plume 2, are present at Zone 4. TCE was historically present in Plume 1 at concentrations exceeding 100,000  $\mu$ g/L, and the plume extended 600 ft. Historical TCE concentrations at Plume 2 exceeded 10,000  $\mu$ g/L. Plume 2 is also approximately 600 ft in length.

Multiple phases of bioremediation have been implemented at Zone 4. The first phase of treatment began in 2003 with the introduction of molasses and bicarbonate buffer using injection wells at Plume 1. Follow-up treatments were initiated in 2004 using molasses, whey, and bicarbonate. This phase of treatment was expanded to include both Plume 1 and Plume 2 and used direct-push technology (DPT) in addition to injection wells. A third phase was implemented in 2008 with the injection of emulsified vegetable oil (EVO) in Plume 1, and a fourth phase was implemented in 2010 with additional EVO injections at both Plume 1 and Plume 2. The EVO injections included both grids and transects of injection wells.

Data from Charleston collected during this project indicate that biogeochemical conditions remain generally favorable for reductive dechlorination of chlorinated ethenes. The presence of daughter products as well as ethane and/or ethene in most wells based on both sampling and PFM data indicates ongoing dehalogenation, although significant mass was still likely present in the Plume 1 source area. The microbial community analysis showed the presence of numerous different dehalogenation-associated organisms/genes in water samples from across the site with *Dehalococcoides* concentrations ranging from 9 x  $10^1$  cells/mL to 3.3 x  $10^6$  cells/mL. The overall CSIA data did not support degradation in the Plume 1 source area well (MW 89-7), but this may reflect the presence of DNAPL, and the continued resupply of unenriched TCE as biodegradation occurs. All other measures indicated ongoing degradation in this area. Moreover, CSIA clearly indicated ongoing biodegradation downgradient of this source. The same scenario was found in Plume 2: ambiguous data in the most upgradient well (MW29-30), with clear isotopic enrichment in daughter products (TCE is largely gone), downgradient of this well, indicative of ongoing biodegradation. Groundwater and PFM data showed that free-phase DNAPL was still present at the site in and or around well MW 89-7 (Plume 1 source area). Overall, the results suggest that biological degradation processes have persisted ~ 4 years after cessation of active bioremediation.

#### Seal Beach Naval Weapons Station, Site 70

Seal Beach Site 70 is also known as the Research, Testing, and Evaluation Area. TCE was used during various research and development activities that occurred at the site between 1962 and 1985. Groundwater contamination is present in the source area where dense non-aqueous phase liquid (DNAPL) is suspected to occur based on the groundwater concentrations; however, DNAPL has not been observed in the various subsurface investigations. A dissolved-phase plume currently extends at least 4,000 ft downgradient to a depth of approximately 160 ft below ground surface (bgs). TCE was the primary contaminant in the plume before bioremediation. High sulfate concentrations (>1,000 mg/L) are present throughout much of the site.

An ESTCP demonstration was implemented in the source area from 2008 to 2010 (ESTCP Project ER-200513). The primary objective of the demonstration was to compare the ability to distribute *Dehalococcoides* via passive versus active approaches. Two test cells were established in the source area, designated the Passive Cell and the Active Cell. Sodium lactate and the SDC-9 bioaugmentation culture were introduced into both cells. The amendments were allowed to migrate with the natural groundwater gradient in the Passive Cell. In the Active Cell, the amendments were distributed by recirculation established between two injection wells and two extraction wells.

The treatment time for the ESTCP demonstration extended over a 10-month period, and significant dechlorination occurred in both test cells as a result.

A full-scale bioremediation program was implemented shortly after the ESTCP demonstration, with a combination of a grid of injection wells in the source area and a series of biobarriers perpendicular to the dissolved-phase plume further downgradient. EVO and the KB-1 bioaugmentation culture were the amendments used in the full-scale design. The full-scale system was implemented in areas of the plume with TCE exceeding 250  $\mu$ g/L. The initial injections in the biobarriers occurred in 2009 and in the source area grid in 2010. The second round of injections occurred in 2013 in the source area and all but one of the biobarriers.

Based on our project data, the long-term results from the Seal Beach site varied by treatment approach and, in some instances, by well in a given treatment area. VOC concentrations in an upgradient well declined since the conclusion of the ESTCP test project completed at the site (evaluating passive vs active approaches for adding electron donors) and daughter products were present, but CSIA data provided no indication of ongoing degradation, and relevant dehalogenating cultures and genes were absent. In the former ESTCP active treatment plot, one well had high concentrations of TCE and daughter products, and the second well had non-detect concentrations. CSIA data showed no clear indication of ongoing degradation of TCE or daughter products. In the former ESTCP passive treatment area, 2/3 wells appeared to have ongoing degradation based on CSIA data. Microbial community analysis indicated the presence of dehalogenation-associated organisms/genes in several but not all locations at the Seal Beach Site. Wells sampled in the passive treatment plots contained *Dehalococcoides* at  $\sim 9 \times 10^2$  to  $4 \times 10^3$ cells/mL, whereas Dehalococcoides was not detected in the active treatment cell. Groundwater data in both the active and passive cell suggest that TCE rebound had occurred in some areas, but below the historically high concentration. In the full-scale treatment area (most recently treated via bioaugmentation), all indications suggest ongoing VOC biodegradation. Taken in sum, the data suggest that biodegradation has persisted in the former ESTCP passive treatment plot ~3 years after injection of cultures and electron donor, whereas dechlorination has largely ceased in the former ESTCP cell that underwent active treatment. The full-scale area that received EVO and bioaugmentation culture also appears to have ongoing VOC biodegradation based on all relevant measures.

#### Treasure Island Naval Station, Site 24

Site 24 at Naval Station Treasure Island is contaminated with PCE, and a plume of contamination extends over 1,100 ft to the San Francisco Bay. The main source area, Building 99, contained dry cleaning facilities that operated between 1942 and 1977. Total VOC concentrations above 1 mg/L occurred throughout much of the plume before remediation. A freshwater aquifer is present in the upper part of the fill material, but the groundwater becomes saline at depths of greater than 30 ft. The aquifer is generally anaerobic, has a neutral pH, and has relatively high sulfate concentrations.

An initial pilot study was conducted at the Building 99 source area from 2003 to 2004. The pilot study system included recirculation with three injection wells and three extraction wells arranged to provide three test loops. Sodium lactate, hydrogen gas, and SDC-9 were applied to various degrees in each of the loops. The pilot study results data showed a substantial decrease in contamination due to the treatment.

Following the successful pilot study, the full-scale treatment was implemented in three phases. Phase 1, conducted from November 2004 to May 2007, consisted of injecting and recirculating a lactic acid solution, hydrogen gas, and SDC-9 across the plume downgradient of the pilot study area. Phase 2, conducted from June 2008 to October 2010, consisted of a combination of recirculating a sodium lactate solution in some areas and injections of emulsified vegetable oil substrate in other areas. The biotreatment was successful throughout much of the plume, but several pockets of contamination remained. Phase 3 was conducted from 2011 to 2012 to address the remaining hot spots of contamination and included additional source area recirculation with lactic acid and SDC-9, along with direct-push applications of Lactoil<sup>TM</sup> in other parts of the plume. Three areas there were targeted for additional treatment included the South Source Area Treatment Area, the EW12 Treatment Area, and the EW30 Treatment Area.

Results from this study showed that biogeochemical conditions remain generally favorable for reductive dechlorination of chlorinated ethenes. Residual PCE was detected in 24-TW-11, in the South Source Treatment Area at 1,190 µg/L. This well also had 779 µg/L of TCE, 7,840 µg/L of cis-DCE and 937 µg/L of VC, respectively. This well showed rebound of PCE after various treatment phases with levels as high as 16,000 µg/L in 2011, but minimal rebound was observed after treatment in 2012 until this sampling event approximately 5 years later. The other well in this same treatment area, 24-EW04 had much lower levels of VOCs than 24-TW-11, with PCE at 1.7 µg/L, TCE at 9.0 µg/L, and VC at 204 µg/L. This well had concentrations as high as 36,000 µg/L in 2005. The other 4 wells that were sampled at Site 24 had residual PCE < 7 µg/L, cis-DCE < 16 µg/L, and VC < 3 µg/L. Thus, overall treatment effectiveness in the EW-12 and EW-30 areas showed little to no rebound after treatment. These wells were primarily contaminated with *cis*-DCE and VC during previous years.

Both groundwater sampling and the limited amount of PFM data detected the presence of ethene and ethane at the site, suggesting that complete biological reductive chlorination was still occurring ~5 years after the cessation of active treatment. The microbial community analysis indicated the presence of quantifiable dehalogenation-associated organisms/genes only in well 24-TW-11, which is the well with highest residual VOCs. It is likely that dehalogenators are not present in the vicinity of other wells due to the exceedingly low residual VOC concentrations. Overall, the CSIA data from this site suggest a rather complex scenario. Wells in the EW30 area, where concentrations of VOCs are very low, each show isotopic evidence of ongoing *cis*-DCE degradation. One of the two wells in the South Source Zone Treatment Area (24-EW04) showed clear evidence of both cis-DCE and VC degradation, but no measurable numbers of relevant dehalogenating organisms/genes. The other well in this region (24-TW-11) showed some evidence of ongoing cis-DCE biodegradation, and very light VC ( $\ddot{o}^{13}C = -38.9\%$ ) indicating that VC was being formed from cis-DCE (i.e., very light daughter product is expected initially as a parent VOC degrades), but probably not biodegrading further. This well had detectable dehalogenating organisms/genes and the highest residual VOC concentrations.

#### COST ASSESSMENT

The nature of this project does not allow for a traditional ESTCP cost assessment, where one remediation or other technology is evaluated against traditional alternatives. However, a reasonable cost estimate for the different assessment technologies utilized during this demonstration is provided. The cost includes labor and per diem for field sampling and PFM

installation, materials for sample collection, CSIA, molecular analysis, PFM analysis, basic geochemistry and VOC analysis. The assessment does not include monies spent on the development of the site database and statistical analysis of the database results.

For the cost assessment, we assume that a total of 8 groundwater wells would be sampled by a single field technician using low-flow sampling, and that the technician could sample 4 wells per day. For each well, basic field parameters (dissolved oxygen, oxidation-reduction potential, pH, conductivity) would be determined using a field meter, and samples would subsequently be collected for (1) VOC concentrations (EPA Method 8260); (2) anions (EPA Method 300); (3) C stable isotope analysis of TCE, cis-DCE and VC and (4) molecular analysis of important dehalogenating organisms and genes. For the PFM analysis, it was assumed that pricing included installation, removal and data analysis on a per PFM basis. Based on all assumptions provided above, the estimated cost of sampling and analysis of 8 wells in support of a complete biodegradation evaluation of PCE/TCE and daughter products was \$37,883. A complete breakdown is provided in the report.

#### **IMPLEMENTATION ISSUES**

The primary end-users of these technologies (MBT, PFM, CSIA) are expected to be DoD site managers and their contractors, consultants and engineers. The general concerns of these end users are likely to include the following: (1) technology availability and cost; (2) appropriate application of the technology at DoD sites; and (3) interpretation of CSIA, MBT and PFM data. These implementation issues are addressed in the following sections. The database developed during this project will also be made available via the ESTCP website.

<u>Availability:</u> The C and Cl stable isotope analyses of VOCs described herein as well as the general qPCR analysis of important organisms and genes responsible for VOC biodegradation are commercially available and conducted in multiple university laboratories. Commercial laboratories include Microbial Insights (Knoxville, TN) and Pace Analytical (Pittsburgh, PA). PFM installation, sampling and analysis is also commercially available from EnviroFlux (Gainsville, Fl). Thus, the key technologies used in this ESTCP project are commercially available. The database developed during this project will also be made available via the ESTCP website.

<u>Technology Application</u>: Appropriate application of the technologies used in this project will vary by site depending on specific conditions and the questions to be answered. During this project, our primary question was the long-term effectiveness of bioremediation and the different tools were combined toward this end and to assess whether biodegradation was still occurring at select sites a few to several years after active treatment.

<u>Data Interpretation: CSIA.</u> CSIA data gathered on environmental pollutants has been utilized to (1) document biological and abiotic contaminant degradation, (2) estimate or constrain rates of contaminant degradation; (3) identify dominant degradation mechanisms; and (4) forensically determine dominant sources of a specific contaminant in the environment, as well as various other specific applications for individual contaminants. The application and interpretation of CSIA data for the above purposes have been thoroughly reviewed in a US Environmental Protection Agency (EPA) document entitled "A Guide for Assessing Biodegradation and Source Identification of Organic Groundwater Contaminants Using Compound Specific Isotope Analysis (CSIA)."

This document is available online through the EPA NEPIS Site. The readers of this ESTCP report are referred to Chapter 4 in this document entitled "Interpretation of Stable Isotope Data from Field Sites" which clearly describes and provides examples of how CSIA data can be utilized to document and quantify the biodegradation of organic contaminants in groundwater aquifers. The Interstate Technology and Regulatory Council (ITRC) Environmental Molecular Diagnostics (EMD) team has also developed online guidance and instruction on CSIA. The C isotope data for TCE, cis-DCE and VC gathered during this project provide information on ongoing degradation of these contaminants at 5 different sites and, for the Raritan Site where samples were collected along a flowpath, an ability to estimate field rates.

<u>Data interpretation: MBT.</u> The molecular analysis conducted during this project (qPCR of key dehalogenating organisms/genes) is now conducted routinely at VOC sites. Guidance concerning the application of this technique and interpretation of results is available at the ITRC EMD website and via Microbial Insights, who provide information on the relative abundance of different gene markers at groundwater sites.

<u>Data interpretation: PFMs.</u> PFM installation, sampling, and analysis is a specialty service that can be provided commercially by EnviroFlux (Enviroflux.com), Dr. Mike Annable at the University of Florida can also provide data analysis and interpretation.

## **1.0 INTRODUCTION**

#### 1.1 BACKGROUND

Aptim Federal Services, LLC (APTIM), along with its project team members at the University of Florida and the University of Tennessee, have prepared this Final Report to describe the methods and results for ESTCP Project ER-201427. We also wish to acknowledge Pace Analytical for conducting the relevant compound specific isotope analysis (CSIA) measurements of chlorinated volatile organic compounds (VOCs).

Source areas of chlorinated VOCs in groundwater create and perpetuate dilute groundwater plumes that subsequently pose risks to downgradient receptors for decades or centuries. Although bioremediation has been applied to treat many source areas with chlorinated ethenes over the past two decades, the overall success of this treatment and factors that differentiate successful from unsuccessful treatment applications have not been thoroughly assessed. We used two separate approaches to evaluate the success or failure of different bioremediation applications and to evaluate factors that may have contributed to those outcomes.

Initially, we selected fifteen (15) VOC sites contaminated primarily with chlorinated ethenes that have been treated using common bioremediation techniques and with data that allow statistical evaluation of remedial performance over time. Sites with extensive pre- and post-treatment data were given priority, and overburden aquifers were selected over bedrock aquifers. A large database was developed from these sites that includes 1) site location; 2) VOC concentrations over time; 3) hydrogeology (e.g., groundwater velocity, depth, groundwater gradient, porosity, percent fines etc.); 4) geochemistry (mineralogy, etc.); 5) water chemistry (pH, Eh, anions, cations, metals, volatile fatty acids (VFAs), co-contaminants, dissolved organic carbon (DOC), total organic carbon (TOC), dissolved gases (methane, ethene, ethane, etc.)); 6) abundances of relevant microbial biomarkers where available (e.g., *Dehalococcoides* (DHC) and relevant reductive dehalogenase genes where available); 7) treatment approach (bioaugmentation, biostimulation, active, passive); and 8) other relevant site data. Statistical analyses were then performed to identify factors that resulted in, or prevented, successful application of the different bioremediation strategies.

From the 15 selected sites for historical data compilation and statistical analysis, 5 sites were chosen to conduct current-day sampling of select wells and used current tools to assess (1) contaminant mass flux, (2) presence of key dehalogenating organisms and genes, and (3) extent of biodegradation that has occurred via CSIA. Pre-treatment data and post-treatment data were compared along with information on site conditions (geologic, hydrogeologic, and geochemical) and treatment methods.

A more thorough understanding of the effectiveness of historical bioremediation applications for treating contaminant source areas will provide the Department of Defense (DoD) with an improved ability to optimize future remediation efforts and to understand the extent of remaining liability at sites where biological treatment has been performed to remediate source areas.

#### **1.2 OBJECTIVE OF THE DEMONSTRATION**

The overall objective of ESTCP Project ER-201427 was to assess the long-term effectiveness of past biological treatment approaches for remediating source areas of chlorinated solvent-contaminated aquifers. In support of this objective, the project team:

- evaluated the effectiveness of source area bioremediation as it has been applied over the past decade;
- determined the distribution and persistence of dechlorinating organisms and genes following bioremediation treatment at select sites;
- determined impacts of biotreatment on contaminant flux from source areas;
- identified important factors that resulted in successful, or failed, bioremediation of chlorinated solvent source areas; and
- identified approaches and characterization/monitoring tools that were most effective.

This multi-site effort identified field-scale aquifer effects of *in situ* bioremediation technologies and provided data for identifying factors affecting the success of bioremediation for source area treatment. Several sites where bioremediation treatment had been implemented were evaluated by using a variety of methods, including:

- historical data review of VOC and geochemical parameters;
- a statistical evaluation of factors contributing to treatment effectiveness;
- field measurements to estimate contaminant flux using Passive Flux Meters (PFMs);
- compound-specific stable isotope analysis (CSIA) to determine the extent and mechanisms of degradation; and
- real-time quantitative polymerase chain reaction (qPCR) to monitor the relevant microbes and indicator genes.

The evaluation contained two major components: (1) a multi-site review and statistical evaluation of historical data for fifteen sites; and (2) additional PFM, CSIA, and molecular biology (qPCR) site characterization at five field sites.

#### **1.3 REGULATORY DRIVERS**

The most common chlorinated solvents formerly used at DoD facilities include tetrachloroethene (PCE), trichloroethene (TCE), and 1,1,1-trichloroethane (1,1,1-TCA). Degradation of these parent compounds results in several chlorinated daughter products. Many of the above compounds are regulated by both the U.S. Environmental Protection Agency (USEPA) and various states. The federal Maximum Contaminant Levels (MCLs) are commonly used as remediation standards, however many states have established more stringent standards. The MCLs for the commonly occurring parent chlorinated compounds and their daughter products are listed in **Table 1-1**.

Constituents	USEPA MCL (µg/L)
Tetrachloroethene (PCE)	5
Trichloroethene (TCE)	5
cis-1,2-Dichlorethene (cis-DCE)	70
trans-1,2-Dichloroethene (tDCE)	100
Vinyl Chloride (VC)	2
1,1,1-Trichloroethane (1,1,1-TCA)	200
1,1,2-Trichloroethane (1,1,2-TCA)	5
1,1-Dichloroethene (1,1-DCE)	7
1,2-Dichloroethane (1,2-DCA)	5

Table 1-1. Federal Maximum Contaminant Levels.

Code of Federal Regulations 40 CFR Part 141

#### 2.0 TECHNOLOGY

#### 2.1 TECHNOLOGY DESCRIPTION AND DEVELOPMENT

#### 2.1.1 Background – Long-term Performance of Enhanced Anaerobic Bioremediation

Chlorinated solvents, such as PCE and TCE, were widely used as degreasing agents at both military installations and in industrial processes. Poor disposal processes resulted in widespread contamination of groundwater sources in the United States. Dense non-aqueous phase liquid (DNAPL) dissolution can serve as a long-term source of contamination, creating plumes that can stretch for miles. In the past three decades, several remediation technologies have been developed for in situ treatment of source zone groundwater, which includes chemical oxidation or reduction, thermal treatment, enhanced DNAPL dissolution and enhanced anaerobic bioremediation (EAB) (Alvarez and Illman, 2006). Among these technologies, EAB has emerged as a promising and cost-effective treatment for the removal of high dissolved concentrations of chloroethenes (Aulenta et al., 2007; Cope and Hughes, 2001; Da Silva et al., 2006; Harkness et al., 1999; Hood et al., 2008; Lendvay et al., 2003; Rodriguez et al., 2004; Major et al., 2002; Pérez-de-Mora et al., 2014; Schaefer et al., 2009; Schaefer et al., 2010; Song et al., 2002; Yu and Semprini, 2009).

For chlorinated ethenes, EAB typically involves the subsurface injection of organic substrates and sometimes *Dehalococcoides* sp. (DHC) that are capable of completely dechlorinating PCE and TCE. Hydrogen (H<sub>2</sub>) is produced as a result of microbial degradation of a organic substrates via fermentation (e.g., lactate, acetate, emulsified vegetable oils, etc.) and is the only electron donor *Dehalococcoides* can use to reduce TCE to ethene (Grady et al., 2011; Schink, 1997). VFAs, a byproduct of fermentation, can be further fermented to hydrogen (Grady et al., 2011; Schink, 1997). Thus, a steady supply of organic carbon can support dechlorination over the long-term (Rectanus et al., 2007; Thomas et al., 2013).

During DNAPL source zone EAB, microorganisms drive down the dissolved phase contamination concentration and increase the mass transfer of contaminants from DNAPL to the aqueous phase (Chen et al., 2013). Enhanced solubilization of DNAPL sources creates the potential for subsequent complete dechlorination downgradient of the source zone. Despite the promise of this technology, questions about the long-term sustainability remain.

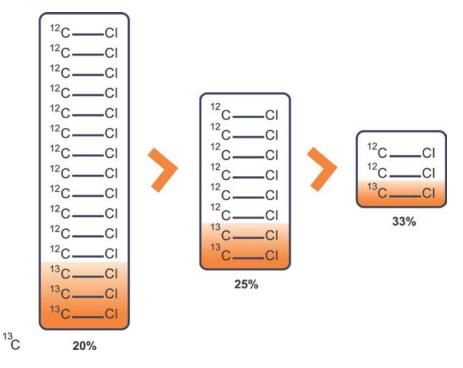
High aqueous phase concentrations near saturation have been shown to be toxic to dechlorinating microbes, leading to inhibition of reductive dechlorination of chloroethenes (Adamson et al., 2003, 2004; Amos et al., 2008; Yang and McCarty, 2000). Adamson et al. (2003) noted the accumulation of TCE and cis-DCE until PCE concentrations decreased to approximately 10  $\mu$ M (~ 1.7 mg/L). Amos et al. (2008) demonstrated that dissolved phase PCE concentrations above 0.54 mM (~ 90 mg/L) inhibited dechlorinating microorganisms but also noted that some PCE dechlorinators produced cis-DCE in the presence of PCE DNAPL. In such cases, PCE-to-DCE dechlorination rates exceeded the rates of PCE dissolution, resulting in non-inhibitory aqueous phase PCE concentrations. Bioremediation has been shown to enhance the rate of PCE DNAPL dissolution in sand columns and flow cells by factors ranging from 1.1 to 21 (Amos et al., 2008; Amos et al., 2009; Christ and Abriola, 2007; Glover et al., 2007; Haest et al., 2012). Bioremediation has resulted in enhanced dissolution to cis-DCE occurs or contaminant rebound occurs (Bondehagen 2010; ITRC, 2007).

Short-term data sets suggest that EAB is an effective treatment strategy for source zones, but the long-term sustainability of this technology is unproven (Hood et al., 2008; Lookman et al. 2007; Puigserver et al., 2016; Scheurtz et al. 2008; Semkiw and Barcelona, 2011; Suthersan et al., 2011). Schaefer et al. (2018) and Haluska et al. (2019) completed a field study showing reductive dechlorination was still occurring ~3.7 years after active treatment in an aquifer containing two hydrological units with different permeabilities. DNAPL entrapped in the lower permeability zone was likely inaccessible to microorganisms, but dechlorinating microorganisms still persisted as mass flux emanating from the source zone decreased and ethene production was sustained long after active treatment (Haluska et al., 2019; Schaefer et al., 2018). McGuire et al. (2016) conducted a multi-site evaluation of post-source zone monitoring data showing a median concentration reduction of 90% at 34 sites, but rebound occurred at 35% of these sites, suggesting remediation is sustainable in the long-term but the rate may not be sufficient to prevent rebound. Tilloston and Borden (2017) also conducted a multisite performance assessment of 37 EAB sites and found that PCE and TCE were removed at ~90% of the sites, but cis-DCE and VC often persisted after active treatment. However, the demonstration of sustained biological activity after treatment requires documented loss of contaminants at field scale, biogeochemical indicator trends, and confirmation of microbial species/genes and degradative activity. Mass flux measurements have also become valuable tools in the assessment of sustained biological activity. We used traditional approaches (biogeochemical measures, contaminant concentration trends, daughter products, etc) and modern tools including compound specific isotope analysis (CSIA) to document ongoing dehalogenation of chlorinated ethenes, quantitative polymerase chain reaction (qPCR) to quantify key dehalogenating organisms and genes, and passive flux meters (PFMs) to quantify contaminant mass flux at several sites in order to assess past performance and post-treatment acitvity of EAB.

#### 2.1.2 Modern Tools for Monitoring In Situ Bioremediation Processes

Several modern tools are now available for evaluating the performance of bioremediation approaches for remediating VOC source zones. Examples of these tools include CSIA, which allows practitioners to evaluate the extent of contaminant degradation at a given location in a plume over time, and to calculate or constrain degradation rates by evaluating the extent of isotopic enrichment at multiple points along a contaminant plume. Likewise, contaminant flux measurements using PFMs allow measurements of the contaminant mass leaving a source area, as well as a highly discrete evaluation of contaminant flow paths in a vertical cross-section of the plume. Together, these tools allow for a more precise evaluation of contaminant mass flux following treatment. Molecular biological tools (MBTs) also are available to evaluate the presence and abundance of degradative microorganisms and to assist in estimating the rate of ongoing biodegradation at a site. Combined with reactive transport modeling and statistical analyses, these tools allow more detailed assessments of treatment effectiveness, factors affecting successful treatment, predictions of on-going treatment, a site's future trajectory, and overall risk reduction. Many of these modern tools have been developed and verified with funding from SERDP and ESTCP. Their application is described in more detail below.

CSIA, especially for PCE and TCE and their degradation daughter products, has become a wellestablished and useful tool for evaluating the extent of transformation of these compounds in plumes, and for estimating their rates of transformation *in situ* (e.g., USEPA, 2008a; Braeckevelt et al., 2012; Hatzinger et al., 2013). The CSIA technique relies on the observation that biological degradation of these compounds results in a change in the ratio of  ${}^{13}C/{}^{12}C$  in the remaining parent compound. This ratio is commonly measured as "delta 13 C" ( $\delta^{13}$ C) (Meckenstock et al., 2004; Sherwood Lollar et al., 2001; Song et al., 2002) (**Figure 2-1**). Thus, by measuring the ratio of  ${}^{13}$ C/ ${}^{12}$ C in the chlorinated solvent parent compound and its daughter products at a given sample location and comparing those ratios to the same ratios of contaminants in the source area, the extent of biodegradation can be estimated. By comparing the extent of degradation along a contaminant plume at different distances from the source, and by using a groundwater reactive transport model to estimate travel times between the sampling points, the rate of contaminant transformation also can be calculated (USEPA, 2008a; Meckenstock et al., 2004; Braeckevelt et al., 2012; Thullner et al., 2012; Song et al., 2002). When combined with flux measurements and discrete groundwater flow estimates, the method can be applied to evaluate the effectiveness of a source area treatment approach over extended time periods (i.e., years following technology implementation). CSIA analysis for this demonstration was conducted by our subcontracted team member Pace Analytical (Pittsburg, PA).

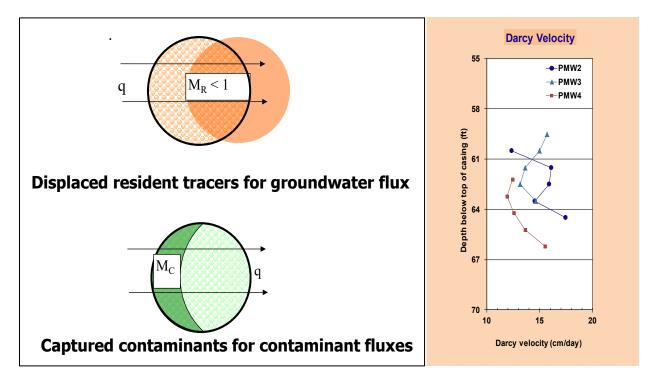


# Figure 2-1. Illustration of 13C Enrichment During Degradation of a Contaminant with a C-Cl Bond.

(Source: Microseeps, Inc., used with permission.)

A key measure of the success of remedial actions is a reduction of risk to downgradient receptors. This can be reflected in a reduction of the quantity of contaminant mass, leaving a source area following treatment over a period of time or a flux reduction. This contaminant flux can be estimated by using PFMs. The PFMs are self-contained, permeable units that are inserted into a well or subsurface boring such that it intercepts, but does not retain, groundwater flow (University of Florida, 2006; Annable et al., 2006). PFMs allow measurements of both groundwater flow and contaminant mass flux (University of Florida, 2006). The mass of contaminant accumulated on a permeable sorbent (typically granular activated carbon) is used to calculate a time-averaged contaminant mass flux, while residual resident tracer mass loss is used to calculate cumulative water flux.

An advantage of the PFM technology is that the medium in the PFM can be sectioned vertically to obtain fine detail profiles (i.e., discrete sampling) of groundwater flow and contaminant mass flux through the subsurface porous media (**Figure 2-2**). Thus, this important information can be ascertained without the high cost of additional drilling and the use of mechanical devices and extensive sample collection to measure discrete groundwater flow and, ultimately, contaminant flux.



#### Figure 2-2. Graphic Showing the Dual Use of Mass Flux Meters for Both Groundwater and Contaminant Mass Flux (Left Panel) and an Example of Data Generated Using Passive Flux Meters to Quantify Fluid Flux in a Groundwater Aquifer in Rancho Cordova, CA (Right Panel).

In the right panel, the three lines show groundwater velocity in different wells with each symbol representing a data point from a different depth in the aquifer (data from ESTCP ER-200828, final report).

The qPCR technique has emerged as the method of choice for monitoring organisms- and processspecific genes in environmental samples, including groundwater (Bustin et al., 2009; Ritalahti et al., 2006, Capiro et al. 2014). A wide number reductive dechlorination and VOC degradation biomarker genes have been identified and validated qPCR assays are available. Until recently, the number of potentially relevant target genes that could be analyzed was limited, primarily because each gene required a separate assay, and the cost associated with the analysis of many biomarkers has been problematic. The more recent advances in chip technology allows the simultaneous analysis of numerous biomarker genes in a single instrument run. The reductive dechlorination quantitative real-time PCR chip (RD-qChip), for example, uses a nanoliter fluidics platform (QuantStudio 12K Flex Real-Time PCR System, Life Technologies, Grand Island, NY) that offers

flexibility, scalability, and high-throughput sample processing to analyze over 12,000 individual qPCR assays in a single instrument run. The flexibility in array construction allows appropriate scaling to achieve stepwise progression to the current limit of 224 biomarker targets on a single array plate (with up to four plates per run). Each array plate can accommodate from 56 assays with 48 samples or up to 224 assays with 12 different samples in parallel. In contrast to manual plate setup in conventional qPCR, the QuantStudio platform uses simple, automated workflows with integrated data analysis software and built-in quality controls, thus effectively mitigating inherent human bias, and generating data that are comparable between samples (i.e., sites) and analytical laboratories. The majority of variation would arise from the sample collection and nucleic acid extraction procedures, but those errors can be accounted for by adding appropriate internal standards (Hatt, et al., 2013). This technology has been adapted for monitoring organismsand process-specific biomarker genes for chlorinated solvent degradation by our collaborator Dr. Frank Löffler as part of SERDP project ER-2312 (Kara-Murdoch et al., 2020; in preparation). The qPCR chip technology was tested at some locations during this project along with a conventional 384-well plate approach and shown to be very comparable. Data from the conventional approach are reported herein for consistency.

With the data collected on contaminant mass flux, degradation (e.g., through CSIA analysis and traditional measures) and microbial biomarkers, an analysis of ongoing biodegradation was performed at the study sites as a "snapshot in time". Several analytical tools are available to compare VOC degradation activities at different sites, and in some cases, factors influencing remedial performance can be understood. For example, bivariate and multivariate statistical methods can be used to identify correlations between these measures of effectiveness versus site characteristics, such as contaminant type(s), pre-remediation VOC concentrations and DHC abundance, presence or absence of DNAPL, the extent of source control, background pH and redox conditions, dynamics of microbial populations present, water quality parameters, etc. Measures of effectiveness (e.g., mass reduction, dechlorination, risk reduction) versus remediation parameters such as remediation methodology (e.g., electron donor used, bioaugmentation versus biostimulation, injection versus recirculation, etc.) were evaluated in conjunction with the analysis of post-remediation redox conditions, post-remediation microbial populations, etc. to recognize patterns and establish correlations.

In terms of technology maturity, CSIA analysis, especially for chlorinated solvents, has become well established as a useful tool for evaluating the extent of transformation of these compounds in plumes and for estimating their rates of transformation *in situ* (USEPA, 2008b; Morrill et al., 2006; Mundle et al., 2012). In addition, contaminant mass flux characterization is an accepted method for evaluating the success of remedial processes, especially in relation to source areas (Brooks at al., 2008), and MBTs, especially for monitoring chlorinated solvent-degrading bacteria, are becoming widely used for monitoring the performance of biological treatment approaches. The ITRC Environmental Molecular Diagnostics (EMD) Team, of which two of the performers on this project were contributing members, published an extensive online Guidance Document concerning the use of CSIA and new molecular tools for documenting *in situ* bioremediation of VOCs and other contaminants (online: http://www.itrcweb.org/emd-2/). Statistical evaluations are an essential component of many fields of research, and the statistical techniques used in this document are well established.

Dechlorination step	RDase gene	Organism	Organism-specific assay		Functional gene assay
			Genus <sup>a</sup>	Species	
$PCE/TCE \rightarrow cis-DCE$	pceA	Dehalococcoides mccartyi	+	+	_ <sup>b</sup>
$TCE \rightarrow VC$	tceA	Dehalococcoides mccartyi	+	+	+
$cis-DCE \rightarrow ethene$	vcrA	Dehalococcoides mccartyi	+	+	+
$cis-DCE \rightarrow ethene$	bvcA	Dehalococcoides mccartyi	+	+	+
$tDCE \rightarrow VC$	-	Dehalogenimonas sp.	+		-
$1,1,1$ -TCA $\rightarrow 1,1$ -DCA	cfrA	Dehalobacter sp.	+		+
$1,1\text{-DCA} \rightarrow \text{CA}$	<i>dcrA</i>	Dehalobacter sp.	+		+
$1,2\text{-DCA} \rightarrow \text{ethene}$	bvcA	Dehalococcoides mccartyi	+	+	+
1,2-DCA → ethene	-	Dehalogenimonas spp.	+		-
$1,1,2$ -TCA $\rightarrow$ VC	-	Dehalogenimonas spp.	+		-

 Table 2-1. qPCR Gene Assays Chosen to Characterize the Microbial Community

 Associated with VOC Transformation.

PCE, tetrachloroethene; TCE, trichloroethene; cis-DCE, *cis*-1,2-dichloroethene; tDCE, *trans*-1,2-dichloroethene; VC, vinyl chloride;1,1-TCA, 1,1,1-trichlororethane; 1,1-DCA, 1,1-dichloroethane; 1,2-DCA, 1,2-dichloroethane;

- <sup>*a*</sup> Genus-specific MBTs are available; however, not all members of the genus are capable of catalyzing the respective dechlorination step.
- <sup>b</sup> A specific assay has not been designed but is possible with the available information.

# 2.2 ADVANTAGES AND LIMITATIONS OF THE TECHNOLOGY

The technologies used in this Demonstration included the CSIA, PFMs, MBTs described in **Section 2.1** along with a multi-site statistical evaluation. The advantages and limitations of these technologies are summarized below.

# CSIA

### Advantages

Advantages of CSIA technology include:

- CSIA does not rely on concentration trends or the observation of daughter products.
- Because it isolates the contaminant (i.e., is compound-specific), CSIA is relevant only to the compound of interest.
- CSIA can detect minute changes in the isotopic ratio with a high level of precision, which allows for careful assessment across a site to reveal subtle but essential differences in contaminant behavior.
- CSIA is a versatile monitoring tool and can be used for many contaminants in a wide range of applications.

# **Limitations**

Limitations of the CSIA technology include:

- A limited number of laboratories provide CSIA services.
- There is the potential for interference when many contaminants are present. However, modifications to CSIA methods can be used to overcome interferences, such as dual gas chromatograph (GC) column separation methods.
- Isotopic fractionation may be so minimal that little or no isotopic enrichment is detected.
- The initial fractionation factors for non-degraded chlorinated solvents area variable and introduce uncertainties into the evaluation.
- Isotopic shifts in daughter products can be challenging to interpret.
- Enrichment factors for some degradation processes may not be available.

# PFM

# Advantages

Advantages of PFM technology include:

- The PFM technique improves the ability to identify high-concentration source areas.
- The PFM technique provides a direct measurement of contaminant loading to receptors.
- The technique enables the cumulative measurement of the flow system and averages the variability obtained in grab sampling.

# Limitations

Limitations of the PFM technology include:

- Subsurface heterogeneities introduce uncertainty into the results.
- Long time periods may be needed for fluxes to reach equilibrium following remediation.

# MBT

# Advantages

Advantages of the MBT technologies include:

- qPCR has emerged as a robust technology for obtaining a quantitative understanding of the abundance of microorganisms of interest in environmental samples. Refined protocols for sample collection and handling, DNA extraction, PCR analysis, and data interpretation are available. In addition, extensive know-how of dealing with potential issues such as the presence of PCR inhibitors (e.g., heavy metals, humic acids) has been generated, and appropriate solutions are available.
- The key microbes and functional genes leading to the detoxification of chlorinated ethenes have been identified, and a comprehensive suite of qPCR assays is available.

# **Limitations**

Limitations of the MBT technologies include:

- qPCR will only provide information about known biomarker genes, and no information was obtained about unknown genes that have not been characterized. Many relevant microorganisms and functional biomarker genes (e.g., genes encoding for dehalogenases) have been identified; however, the sequence variability and diversity of existing biomarker genes is not fully understood. Thus, false-negative results are possible, or, more likely, the analysis does not capture the full diversity of target genes, and the results represent a conservative estimate of the true target gene abundance.
- qPCR analysis is generally performed using biomass collected from groundwater samples. Since aquifer microbes, including the dechlorinators, also attach to aquifer solids, qPCR analysis performed on groundwater samples misses the attached fraction. Consequently, the groundwater analysis represents a conservative estimate of the true target gene abundance.
- The presence of a gene of interest does not necessarily correlate to enzymatic activity.

# **Multi-Site Statistical Evaluation**

## <u>Advantages</u>

Advantages of the statistical evaluation technologies include:

- The multi-site approach allows for a comparison of factors that span a wide range of hydrogeologic and geochemical conditions.
- A consistent evaluation approach is used, thereby eliminating site-by-site bias.

# **Limitations**

Limitations of the statistical evaluation technologies include:

- Historical data are not available for PFM, CSIA, and MBT at most sites because these methods are relatively new and not widely used historically.
- Historical data used for the evaluation was obtained from external databases with generally unknown data quality. The data were assumed to be useable from a data quality perspective; independent data validation was not performed.
- Historical data was not consistently available for many analytes of interest. For example, VOC concentrations are used extensively for performance monitoring at bioremediation sites; but other analytes such as anions and metals were analyzed on an inconsistent basis or not at all.
- A general principle of statistical evaluations is the caveat that correlations do not prove causation. Any observed correlation must be carefully evaluated to determine if a valid conclusion can be drawn from the observation.

# **3.0 PERFORMANCE OBJECTIVES**

The performance objectives for the study, along with data requirements and success criteria, are described below and summarized in **Table 3-1**. The following subsections provide a brief description of each objective, along with data requirements and success criteria.

# **3.1** EVALUATE THE EFFECTIVENESS OF SOURCE AREA BIOREMEDIATION – HISTORICAL DATA

One of the primary objectives of this demonstration is to assess the long-term effectiveness of applying bioremediation at chlorinated solvent source areas. Understanding the extent to which source area concentrations have been reduced, and the subsequent impact on groundwater quality downgradient is critical in this evaluation

## Data Requirements

Several factors were considered as part of the statistical evaluation of historical data, including information on remedial design and substrate loading along with historical site data on pre-and post-treatment conditions. The following site data was gathered for the evaluation:

- 1. Chlorinated VOC concentrations before, during, and after treatment are the primary means to determine changes due to treatment.
- 2. Hydrogeologic factors include hydraulic conductivity, hydraulic gradient, porosity, and seepage velocity. Geologic characteristics include aquifer heterogeneity and the grain-size distribution of soils.
- 3. Geochemical indicator parameters include pH and oxidation-reduction potential (ORP). Commonly occurring native electron acceptors include dissolved oxygen (DO), nitrate, sulfate, iron, and manganese.
- 4. Treatment approaches include a variety of factors, including:
  - the method of injection such as recirculation wells, direct-push injections, standard injection wells, and temporary injection points installed by direct-push methods;
  - the substrate used, for example, sodium lactate or emulsified vegetable oil;
  - the dosing of the substrate expressed as the concentration of the substrate as a percentage;
  - the substrate loading (total pounds injected per treatment);
  - the spacing of the injection points;
  - whether bioaugmentation was employed, and if so, the specific microbial culture; and
  - whether or not nutrients were added.
- 5. The presence, nature, and distribution of microbial populations were evaluated in comparison to treatment effectiveness.

Performance Objective	Data Requirements	Success Criteria	Criteria Met					
Quantitative Performance Objectives								
Identify important factors that result in successful, or failed, bioremediation (15 sites total) Develop a database to correlate, remedial approach, remedial success, and relevant site factors	<ul> <li>Treatment approach</li> <li>Historical contaminant data (VOC; μg/L)</li> <li>Historical geochemical data (ORP, anions, metals, pH)</li> <li>Current and/or past microbial community – key organisms and genes</li> <li>Site geology</li> </ul>	Statistical correlations between treatment approach, remedial success, and site factors	Yes					
Evaluate the long-term effectiveness of source area bioremediation (5 sites)	<ul> <li>Groundwater samples from multiple wells per site – multiple remediation approaches in some cases.</li> <li>Current VOC data (μg/L),</li> <li>Mass flux data (M/L<sup>2</sup>/T),</li> <li>CSIA data (CSIA, δ<sup>13</sup>C)</li> <li>Microbial community analysis (arrays; cells/genes/mL)</li> </ul>	Source reduction based on parent VOC concentrations at or below historical levels Estimation of current occurrence/rate of bioremediation <i>CSIA</i> <i>VOC flux measurement</i> <i>Daughter products</i> <i>Microbial communities</i>	Yes					
Qualitative Performance Objectives								
Develop a useful database and tools for evaluating long-term bioremediation/quantifying key factors	Project results and data output	Expandable database Useful correlations observed	Yes					

# Table 3-1. Performance Objectives.

# Success Criteria

Measures of effectiveness are related to various aspects of VOC concentration reduction and the extent of dechlorination. Metrics were developed for each of the following measures of success and failure:

- Mass Reduction
- Dechlorination
- Risk Reduction (based on MCLs)
- Rebound

- DCE Accumulation
- VC Accumulation

The development of the metrics for the above measures is described in Section 5.

# **3.2 EVALUATE LONG-TERM EFFECTIVENESS OF SOURCE AREA BIOREMEDIATION**

The primary objective of this task is to evaluate the past effectiveness of source area bioremediation at five different sites where different remedial approaches were implemented. VOC analysis, mass flux analysis, CSIA analysis, and microbial community analysis were conducted on groundwater from several different source areas and downgradient locations. The data requirements and success criteria for each of these technologies are described in the following subsections.

# 3.2.1 Microbial Community/Gene Analysis

Dechlorination activity following the active phase of bioremediation treatment (i.e., electron donor additions) can occur due to both biotic and abiotic processes. Both processes can provide a remedial benefit and attenuate contaminants in post-remediation aquifers. Improved understanding of these long-term processes leads to improved treatment design and site management.

## Data Requirements

The data requirements for the microbial/gene analysis include the results of the qPCR analysis of key target organisms/genes from the University of Tennessee.

# Success Criteria

A key success criterion of the microbial analysis is the ability to quantify important dehalogenating organisms/genes in the site groundwater samples at current conditions. Ideally, these values can then be correlated with other site measures (VOC concentrations and/or mass flux at the same locations, daughter products, CSIA data) to provide an assessment of the important organisms/genes and/or their critical densities for long-term degradative activity at a site.

# **3.2.2** Contaminant Mass Flux from Source Areas

Decreases in groundwater VOC concentrations were the primary metric used to assess potential downgradient impacts. However, migration of DHC and changes in biogeochemical conditions also are useful. Decreases in mass flux emanating from the source area also were used as a metric.

### Data Requirements

Data requirements include flux measurements using PFMs combined with flux modeling based on pre- and post-treatment data.

### Success Criteria

Quantification of site-specific flux measurements was conducted at various sites. The mass flux measurements were evaluated along with CSIA and qPCR analyses to evaluate long-term degradative activity at a site.

# 3.2.3 CSIA Analysis

CSIA analysis was conducted for TCE and daughter products (cis-DCE and/or VC) to provide evidence of ongoing biodegradation of these VOCs at the field sites.

## Data Requirements

Measurement of  $\delta^{13}$ C of TCE, cis-DCE and/or VC (based on site conditions and concentrations) was conducted using groundwater samples from source area and downgradient wells and/or areas where different approaches were used. The data were used to obtain evidence of ongoing biodegradation of these VOCs based on (1) historical  $\delta^{13}$ C values of TCE and/or (2) increasing  $\delta^{13}$ C values along a contaminant flow path where wells/samples were available for such analysis.

## Success Criteria

The success criterion is the ability to document ongoing contaminant biodegradation based on isotopic enrichment compared to relevant source  $\delta^{13}C$  of TCE. If a flow path and groundwater velocity can be estimated, then rates of degradation can be estimated based on known fractionation factors and enrichment in <sup>13</sup>C in TCE, (cis-DCE or VC) as a function of travel time in the aquifer.

# **3.3 IDENTIFY IMPORTANT FACTORS THAT RESULT IN SUCCESSFUL OR FAILED BIOREMEDIATION**

The identification of important factors results from the statistical evaluation of effectiveness.

### Data requirements

As previously described, important data requirements include (1) historical site data showing preand post-treatment contaminant concentrations, geochemical indicators and microbial composition (2) geochemical and hydrogeologic site characteristics, and (3) treatment approach information including amendments used and delivery methods.

### Success Criteria

Success is evaluated using statistical methods to compare conditions to measures of effectiveness.

# **3.4 IDENTIFY APPROACHES, CHARACTERIZATION AND MONITORING TOOLS THAT WERE MOST EFFECTIVE**

The identification of approaches and tools that were most effective were obtained from examining the most successful sites.

### Data requirements

Data requirements are site information gathered from site reports along with historical analytical data.

### Success criteria

The results of the multivariate statistical analysis are used to identify approaches and tools that were most effective.

# 4.0 SITE DESCRIPTIONS

A Site Selection Memorandum was previously prepared that described the site selection process. The goal of the site selection effort was to have a group of sites that were representative of widely used bioremediation techniques and that allow statistical evaluation of remedial performance. Sites with extensive pre- and post-treatment data available and accessible for collecting additional samples were given priority, and overburden aquifers were selected rather than bedrock aquifers. Some sites were also selected that have applied bioaugmentation using various commercially available dehalogenating cultures.

The site selection process was initiated by first identifying DoD sites where bioremediation has been applied. A list of approximately 200 sites was identified by searching readily available public information. Sources of information included ESTCP case studies, other publications, administrative records, and general internet searches.

Sites were selected for this study based on: 1) the remedial approach used (e.g., bioaugmentation vs. biostimulation; active vs. passive); 2) electron donor used (e.g., vegetable oil, lactate, polylactate, molasses, lactate oil); 3) hydrogeology; 4) geographical location; 5) contaminant type and concentration(s); 6) background redox and pH conditions; and 7) time since remedial activities commenced.

Many of the selected sites have had multiple phases of implementation over time with differing amendments and approaches for delivery, thereby allowing the evaluation of different treatments at the same sites. Likewise, many of the sites also have multiple depth horizons with differing geologic characteristics, thereby allowing an evaluation of the same treatment approaches under varying geologic conditions at the same site.

Fifteen sites were selected for the statistical evaluation based on the criteria described above and preliminary interviews with installation Remedial Project Managers (RPMs). These sites, which are also summarized in **Table 4-1**, were the following:

- Alameda Naval Air Station, Site 4
- Former Charleston Air Force Base, Zone 4
- Dover Air Force Base, Area 6
- Moody Air Force Base, FT-07
- Former Myrtle Beach Air Force Base, FT-11
- North Island Naval Air Station, OU 24
- Orlando National Training Center, SA 17
- Point Mugu Ventura County Naval Base, Site 24
- Former Raritan Arsenal AOC 2
- Seal Beach Naval Weapons Station, Site 70
- St Juliens Creek Annex, Site 21
- Treasure Island Naval Station, Site 24
- Vandenberg Air Force Base, Site 15A,15B
- Vandenberg Air Force Base, Site 32C/35
- Vandenberg Air Force Base, Site 19

Sites	Area	Contaminant	Added Carbon	Bioaugmentation	Implemented	Geology	Delivery
Alameda NAS	Site 4	TCE	Lactate	SDC-9	2012	Sand	Recirc
Former Charleston AFB	Zone 4	TCE	EVO, Molasses	None	2003, 2004, 2008, 2010	Interbedded fine sand, silt, clay	IW, DPT
Dover AFB	Area 6	TCE	Lactate, DAP	None	2002, 2007, 2008, 2011	Medium sand	Recirc, DPT
Moody AFB	FT-07	TCE	HRC, EOS	SDC-9	2002, 2005 2008	alluvial clay, silt, sand	IW
Former Myrtle Beach AFB	FT-11	TCE	Lactate, Lactoil <sup>TM</sup> ,	SDC-9	2005, 2006, 2007, 2008	interbedded sand, silty sand	Recirc
North Island NAS	OU 24	TCE	Lactate	KB-1	2007, 2012	silty sand, silt, clay	Recirc
Orlando National Training Center	SA 17	TCE	EOS	None	2006, 2008, 2012	interbedded sand, silty sand	Recirc, DPT
Point Mugu Ventura County NB	Site 24	TCE	ESO, Lactic acid	BAC-9	2002, 2006, 2013	mixed fill, sand and clay	DPT
Former Raritan Arsenal	AOC 2	TCE	Lactate, Lactoil <sup>TM</sup> ,	SDC-9	2009	interbedded, sand, silty sand	Recirc
Seal Beach NWS	Site 70	TCE	EVO, Lactate	KB-1, SDC-9	2008, 2009, 2010, 2011	layered silt and sand	IW
St Juliens Creek Annex	Site 21	TCE	EVO	None	2008, 2010	silty sand	DPT, IW
Treasure Island NS	Site 24	PCE	Lactate, Lactic acid	SDC-9	2004, 2008, 2009, 2012	sand and silt	Recirc, IW
Vandenberg AFB	Site 15A	TCE	Lactate, Hydrogen	SDC-9	2010, 2011	sand and clayed sand	Recirc, DPT
Vandenberg AFB	Site 15B	TCE	Vegetable Oil	SDC-9	2010, 2011	sand and clayed sand	Recirc, DPT
Vandenberg AFB	Site 32C/35	TCE	Molasses, Lactoil <sup>TM</sup> , Lactate	SDC-9	2000, 2009	Unknown	IW
Vandenberg AFB	Site 19	РСЕ	HRC, Lactoil <sup>™</sup>	BDI, SDC-9	2006, 2008, 2010, 2011	interbedded, sand, clay	DPT

Table 4-1. Sites Selected for Evaluation.

\*NAS = Naval Air Station; \*AFB = Air Force Base; \*NB = Naval Base; \*NS = Naval Station; \*NWS = Naval Weapons Station; \*EVO = Emulsified vegetable oil

A subset of five of the above sites was selected for additional field investigations. The field investigation included standard laboratory analysis, including VOCs, anions, reduced gases, and alternate electron acceptors, along with PFM contaminant flux analysis, CSIA analysis, and MBT analysis. The suite of analytes is described further in **Section 5.2**. The five field sites are as follows:

- Former Charleston Air Force Base, Zone 4
- Dover Air Force Base, Area 6
- Former Raritan Arsenal, AOC 2
- Seal Beach Naval Weapons Station, Site 70
- Treasure Island Naval Station, Site 24

Brief descriptions of each of the five field sites are provided below. Brief descriptions of the remaining ten sites used in the statistical analysis are provided in **Table 4-1** and **Appendix B**.

# 4.1 FORMER CHARLESTON AIR FORCE BASE ZONE 4

Zone 4 at Charleston Air Force Base (CAFB) is an industrial area near the CAFB flight line. Trichloroethene (TCE) was formerly used for aircraft parts cleaning. Two distinct plumes of TCE contamination, designated Plume 1 and Plume 2, are present at Zone 4. Plume 1 is associated with an oil/water separator at Building 543. TCE was historically present in Plume 1 at concentrations exceeding 100,000  $\mu$ g/L, and the plume extended to a distance of 600 ft. The origin of Plume 2 is a solvent spill that occurred in 1978 at Building 532. Historical TCE concentrations at Plume 2 exceeded 10,000  $\mu$ g/L. Plume 2 is also approximately 600 ft in length. Additional details are provided in **Section 5.3.1**.

# 4.2 DOVER AIR FORCE BASE AREA 6

Area 6 at Dover Air Force Base contains a large plume of chlorinated ethenes and ethanes that is approximately 2,500 ft wide and 6,500 ft long. Several source areas have been identified with the most important source area being Site OT14, also known as Building 719. Building 719 is a jet engine maintenance facility, and TCE was used to degrease engine parts in the 1960s. Dip tanks and associated drain lines are the suspected sources of TCE contamination at this site (ORNL 2004-10). Significant contamination is present beneath the engine cleaning rooms. Two underground storage tanks (USTs) were formerly present at Building 719 that may have been contributing sources. This site contains a mixed solvent plume with PCE, TCE, and 1,1,1-TCA contributing to the source. Maximum TCE concentrations historically exceeded 20,000  $\mu$ g/L. PCE, cis-DCE, and VC are present at lower concentrations. Baseline concentrations of 1,1,1-TCA and its breakdown product 1,1-DCA) ranged between 500 and 1,000  $\mu$ g/L. The breakdown product 1,1-DCE was present at lower concentrations. Additional details are provided in **Section 5.3.2.** 

# 4.3 FORMER RARITAN ARSENAL AOC 2

AOC 2 at the former Raritan Arsenal contains a TCE plume that extends approximately 3,000 ft. Multiple sources contribute to the plume; however, the source at Building 256 is apparently the most significant contributor to contamination. Concentrations of VOCs exceeded 1 mg/L before treatment with cis-DCE having the highest concentration. The primary source of the plume is

located at Area 18C, the Building 256 Ramp Area. Building 256 was formerly used for vehicle maintenance. The specific operations that produced the contamination are uncertain, but TCE likely was discharged to the ground surface at the Ramp Area. Additional details are provided in **Section 5.3.3**.

# 4.4 SEAL BEACH NAVAL WEAPONS STATION SITE 70

Seal Beach Naval Weapons Station is located south of Los Angeles. Seal Beach Site 70 is also known as the Research, Testing, and Evaluation Area. TCE was used during various research and development activities that occurred at the site between 1962 and 1985. From 1962 to 1973, the area was used for the design and manufacture of Saturn II launch vehicles for the Apollo space program. Additional research and development activities were subsequently conducted at the site (Geosyntec 2008). Additional details are provided in **Section 5.3.4**.

# 4.5 TREASURE ISLAND NAVAL STATION SITE 24

Site 24 at Naval Station Treasure Island is contaminated with PCE, and a plume of contamination extends over 1,100 ft to the San Francisco Bay. The main source area, Building 99, contained dry cleaning facilities that operated between 1942 and 1977. Total VOC concentrations above 1 mg/L occurred throughout much of the plume before remediation. Additional details are provided in **Section 5.3.5**.

# 5.0 TEST DESIGN

The following subsections describe the system design and testing conducted to address the performance objectives described in **Section 3.0**. The primary phase of the project included database development, field investigations at the 5 select sites, data analysis, and statistical evaluations as described in the following subsections.

# 5.1 DATABASE DEVELOPMENT

The multi-site evaluation included the gathering of historical analytical data and information on site conditions and the treatments applied. Electronic data were obtained for each of the evaluation sites whenever possible. Data was also gathered from hard copy sources when electronic data were not available. Critical sources of electronic data include the following databases maintained by the government:

- The Environmental Restoration Information System (ERIS) is the Army's web-based database system for the storage of Army environmental restoration field data. It serves as a central repository for the Army installation chemical, geological, and geographical data.
- The Environmental Resources Program Information Management System (ERPIMS) is the Air Force system for the management of data from environmental projects at Air Force installations. This data contains analytical chemistry samples, tests, and results, as well as hydrogeological information, site/location descriptions, and monitoring well characteristics.
- The Navy Installation Restoration Information Solution (NIRIS) provides analytical data along with data management and visualization tools. NIRIS also includes a collection of site documents.

Documents generated as part of site remediation activities were obtained from web-based document-sharing sites, where available, and through direct requests to installation Remedial Project Managers (RPMs). Site reports were reviewed to obtain details on hydrogeologic conditions and the treatments applied at each site. Typical documents with site information included, for example, Remedial Designs, Remedial Action Workplans, Remedial Action Completion Reports and Monitoring, and Progress Reports.

Because of the large amount of existing information and data collected and analyzed during this project, a database was created to organize and evaluate the data. Data were housed in a SQL Server database. The SQL server provided a back-end data repository for analytical data and was managed by a front-end Microsoft Access database. Data added to the Microsoft Access database included: 1) site location; 2) VOC concentrations; 3) hydrogeology (e.g., groundwater velocity, groundwater gradient, porosity, etc.); 4) geochemistry; 5) water chemistry (pH, Eh, anions, cations, metals, VFAs, co-contaminants, DOC, TOC, methane, ethene, ethane, acetylene, etc.); 6) abundances of relevant microbial biomarkers (e.g., DHC, reductive dehalogenase genes); 7) treatment approach (bioaugmentation, biostimulation, active, passive); and 8) other relevant site information. A chart showing the flow of data is provided in **Figure 5-1**.

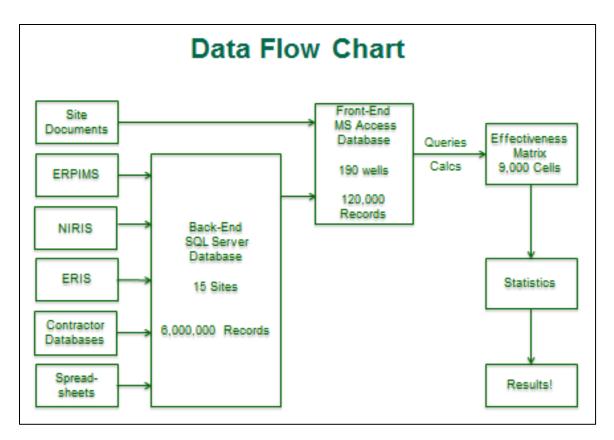


Figure 5-1. Data Flow Chart.

# Database Data Quality Assessment

The analytical data obtained from the various sources (e.g., government databases, contractor databases, and data summary spreadsheet) was assumed to be accurate. An independent data validation of the original laboratory reports was not conducted as part of this evaluation. A limited data quality assessment was conducted, as described below.

Analytical results outside the normal range for some analytes were eliminated as outliers and not included in the evaluation. Criteria for elimination as outliers included results outside the following limits:

- pH < 0;
- pH > 14;
- DO < 0 mg/L;
- DO > 10 mg/L;
- ORP < 600mV; and
- ORP > + 1,000 mV

The above analytes were all from measurements with field instruments. The results outside the ranges noted above likely resulted from instrument error, operator error, or transcription error. Regardless of the cause, the results outside the limits noted above were rejected as unusable.

Duplicate results included field duplicates, laboratory replicates, and data entry duplicates. Field duplicates were not used in the evaluation in preference to the non-field-duplicate result. Laboratory replicates occur as a result of multiple runs on the same sample. These results often differ by a factor of 15% or less. Often the laboratory designates a single result with a "Reportable Result" flag. However, this designation was not available in many cases. In the case of laboratory replicates without a Reportable Result flag, the maximum result of the laboratory replicates was selected as a conservative measure.

# 5.2 FIELD INVESTIGATIONS

# 5.2.1 Sampling Rationale

Five field sites were selected for additional field investigations to assess current post-treatment conditions. The investigations included groundwater sampling from existing monitoring wells and analysis of a variety of parameters. As previously noted, the five field sites were as follows:

- Former Raritan Arsenal, AOC 2
- Dover Air Force Base, Area 6
- Charleston Air Force Base, Zone 4
- Seal Beach Naval Weapons Station, Site 70
- Treasure Island Naval Station, Site 24

The objective of the groundwater sampling was to assess the long-term impacts of the previously performed biostimulation and/or bioaugmentation treatments on contaminant concentrations, groundwater quality, and biogeochemistry. Field activities at each site included the collection of groundwater samples and installation of passive flux meters (PFMs) in existing monitoring wells at each of the five field sites.

Groundwater samples were collected at monitoring wells variously located within treatment zones and at downgradient locations. Wells were selected based on treatment effectiveness and include sites with apparent successful and unsuccessful bioremediation applications at different times in the past. A description of each of the field sites, sampling rationale for each site, the monitoring wells selected for sampling, and the number of groundwater samples at each site is described in subsequent subsections.

The analysis included the following parameters:

# <u>VOCs</u>

VOCs analysis by EPA Method 8260 at APTIM's Analytical and Treatability Lab (ATL) was used to assess the extent of treatment and/or rebound, and the extent to which reductive dechlorination has occurred/continued to occur since treatment implementation.

## Reduced Gases

Reduced gases, including methane, ethane, and ethene (MEE) measured using EPA Method 3810, RSK-175 were used to assess the extent to which complete reductive dechlorination is still occurring within the treatment area. Methane data also provide insight into the redox conditions.

## <u>Anions</u>

Anions, including chloride, nitrate, and sulfate, measured by EPA Method 300.0, provide insight into the current redox conditions and the dominant electron acceptors. Total sulfate plays a significant role in rates/extents of reductive dechlorination at specific sites or locations at a given site. Chloride levels provide a measure of salinity at each site.

### Dissolved Iron and Manganese

Dissolved iron and manganese levels were measured using EPA Method 601.0. These data provide insight into the current redox conditions and dominant electron acceptors and may also provide useful information about potential abiotic reactions.

## Volatile Fatty Acids

Volatile fatty acids (VFAs), measured by EPA Method 300m, are commonly present during active bioremediation treatment. Groundwater samples were analyzed to determine if residual levels persist, possibly serving as electron donors.

## Dissolved Hydrogen

Dissolved hydrogen levels, measured using EPA Method 3810, are used to assess whether hydrogen is present at detectable levels at any site months to several years after biological treatment, as an indication of ongoing fermentation.

### Total Organic Carbon

Total Organic Carbon (TOC) levels, measured using EPA Method 415.1, indicate available organic carbon that could support reductive dechlorination.

### Field Parameters

Dissolved oxygen (DO), pH, turbidity, and ORP were measured as indicators of geochemical conditions using a standard multi-parameter field meter. Values were recorded during low-flow sampling after each well had stabilized.

### Microbial Community

Dr. Frank Löffler at the University of Tennessee performed a targeted microbial community analysis (qPCR) to determine the presence and abundance of organohalide-respiring bacteria including DHC, *Dehalobacter*, and *Dehalogenimonas*. In addition, the University of Tennessee quantified functional genes involved in reductive dechlorination of chlorinated solvents.

# CSIA

CSIA analysis was performed for carbon in TCE, cis-DCE, and VC as a means to assess whether VOC degradation is occurring or has occurred in the recent past based on current values of  $\delta^{13}$ C.

The measured  $\delta^{13}C$  value for TCE are evaluated as a function of (1) historical values of  $\delta^{13}C$  in manufactured TCE (generally -30 ± 5 ‰); (2) previous CSIA analyses conducted at the site when available and/or (3) distance from the source area. The analysis conducted will depend on the current and historical data that are available. The analysis was done by Pace Analytical, Inc. (Pittsburgh, PA).

# Mass Flux

The mass flux of VOCs was evaluated by Dr. Mike Annable at the University of Florida using the PFM procedures summarized in **Appendix D**.

# 5.2.2 Groundwater Sampling Procedures

Groundwater wells were sampled in accordance with the procedures described in this section. The sampling method used was dependent on well construction, with standard parameter measurements taken in conventional large-diameter monitoring, extraction, and injection wells. Modified sampling procedures were applied to multi-level and small-diameter wells. The wells were purged, by micro-purge technique, prior to sampling using a pump and dedicated tubing.

Groundwater samples were collected from the monitoring wells using a portable bladder or peristaltic pump and low flow techniques. Dedicated polyethylene tubing was maintained for each monitoring well. Specific considerations for individual well types are as follows:

- Large Well (2-inch or greater diameter) Sampling was conducted using a standard 1.5inch (nominal) diameter down-hole bladder pump. Field chemistry parameters were measured during the purge process, and sampling was initiated once parameters had stabilized, as discussed below.
- Narrow Diameter Well (less than 2-inch diameter) Sampling was conducted using a surface-based peristaltic pump. Field chemistry parameters were measured during the purge process, and sampling was initiated once parameters had stabilized as discussed below.
- Sampling of multilevel wells (if present) was conducted using a surface-based peristaltic pump. Sampling was initiated after purging a minimum of 1 liter of groundwater. No field chemistry parameters were measured during the purge process.

When purging required the measurement of field parameters, the field sampler monitored purged water for turbidity, pH, temperature, DO, ORP, and specific conductivity using a hand-held instrument (Horiba U-10 or equivalent) and a flow-through cell. After the monitoring wells had been purged, and the field parameters had stabilized, groundwater samples were collected directly from the end of the pump tubing.

The following standard procedures were followed when sampling.

- 1. Confirm the well identification at each well.
- 2. Calibrate field instruments following the manufacturer's directions. Record all calibration documentation in the field logbook.

- 3. Measure the depth to water at each well using an electronic water level indicator probe. Record the water level measurement to the nearest 0.01 inch in the field logbook. Decontaminate the water level indicator before each measurement.
- 4. Carefully lower the pump into the well. Use dedicated tubing for sampling. Place the intake of the pump to the middle of the screen interval. Ideally, the pump speed was set, so the water column in the well did not drop more than 1 ft below the initial water level reading, although sampling multi-level wells precludes simultaneous water level measurement. Using a micro-purge technique, the flow rates generally ranged between 300 and 500 milliliters per minute (mL/min).
- 5. Where required, water quality parameters (i.e., DO, conductivity, pH, ORP, turbidity, and temperature) were measured every 3 to 5 minutes during purging. In-line monitoring equipment was used to minimize exposure to the atmosphere to increase the reading stability for accurate measurements of DO and ORP. Water quality parameters were recorded on the groundwater sampling log forms. When the water quality parameters were stable for three consecutive readings, samples were collected for chemical analysis. Stabilization is achieved if successive readings are within +/- 0.1 pH units, +/- 1 degree Celsius for temperature, 3 percent for conductivity, and 10 percent for DO. Turbidity and ORP readings were not used as stabilization criteria. In situations where the water quality parameters could not be stabilized in a reasonable timeframe, purging was continued until at least three calculated well volumes were purged prior to sampling.
- 6. For sample collection, the pump flow was reduced to a rate of less than 300 mL/min. Flow rates during the collection of VOC samples were set not to exceed 100 mL/min. The pump was not stopped after stabilization and prior to sample collection unless the water level was approaching the base of the well.
- 7. Samples were collected from the discharge of the pump into the appropriate sample containers. Field quality control samples (e.g., field duplicates) were collected as required.
- For microbial community analysis, 1L of groundwater was passed through Sterivex filters in the field, and the filters were subsequently be used for microbial/gene analysis. Section 5.2.3 and Appendix C provides a detailed description of the procedure.
- 9. The samples were labeled, packaged, and prepared for shipment to the laboratory. All samples were placed in cold storage immediately after collection and until shipping/ transport in coolers on ice.

After sampling was complete at one well, the portable pump was decontaminated before being inserted into each monitoring well. Sampling and measuring equipment reused in multiple wells was decontaminated before use. This includes bladder pumps, water level indicators, and any other instrumentation or material potentially exposed to contaminants. Decontamination of sampling and measurement equipment included the following:

- Initial wash using Alconox or other approved detergent;
- Rinse with potable water; and
- Air dry or drying using a clean towel.

The use of dedicated down-hole tubing and the replacement of sampling bladders between wells precluded the need to decontaminate all but the reusable metallic components of bladder pumps. Peristaltic pumps using dedicated or disposable tubing require no decontamination. Purge water was disposed of following the established practices at each site.

# 5.2.3 qPCR Procedures

Groundwater samples were collected and shipped to the University of Tennessee either as groundwater in 1-L bottles, or the biomass that was collected on Sterivex-GP cartridges *on site* (Millipore, Billerica, MA, Cat.#: SVGPL10RC) (Ritalahti et al., 2010; Ritalahti et al. 2010a,b). Further details on sample collection procedures are provided in **Appendix C**.

DNA was isolated from the Sterivex-GP cartridges using the MoBio PowerLyzer PowerSoil Kit (MoBio, Carlsbad, CA) according to the manufacturer's recommendations. DNA concentrations were quantified using the Qubit dsDNA BR Assay (Life Technologies, Grand Island, NY) according to the manufacturer's manual. DNA solutions were stored at -80°C until analysis.

For qPCR analysis, each sample was diluted to three different dilutions using nuclease-free water (i.e. 1:10, 1:100 and 1:1,000) to determine if any contaminants were present that would interfere with the qPCR analysis. A non-linear response would indicate the presence of inhibitors and the dilute sample(s) that gave the best fit within the template DNA standard curve was included in the analysis (typically the 1:10 dilutions).

Standard curves were prepared using plasmid DNA (pDNA) synthesized and incorporated into *E*. coli by Life Technologies (Grand Island, NY). Alternatively, target gene fragments were amplified with PCR following isolation of gDNA using the MoBio PowerLyzer PowerSoil Kit per kit protocol (MoBio, Carlsbad, CA) and cloned into the pCR<sup>TM</sup>2.1 vector and incorporated into *E*. coli using the Invitrogen TA Cloning<sup>TM</sup> kit per kit protocol (Life Technologies, Grand Island, NY). The *E*. coli transformants were grown in Lysogeny Broth (LB) with ampicilin (100 µg/L) or kanamycin (50 µg/L) at 37°C overnight. pDNA was isolated using the Zymo Research Zyppy<sup>TM</sup> Plasmid Miniprep Kit (Zymo Research Corp., Irvine, CA) and quanitified using a NanoDrop and the Qubit 2.0 Fluorometer. All standard curves had a total of eight calibration points.

# 5.2.4 Passive Flux Meter Procedures

The PFM design for organic contaminants is a self-contained permeable activated-carbon cartridge that is inserted into a well screen or boring such that groundwater flows horizontally through the cartridge under ambient hydraulic gradients. The PFM is constructed in units 1.5 meters long and can be stacked in wells to cover long vertical screen intervals. The activated carbon serves to intercept and retain dissolved organic contaminants present in groundwater flowing through the well screen. The activated carbon is also impregnated with known amounts of soluble 'resident tracers', typically alcohols. These tracers are leached from the sorbent at rates proportional to the water flux.

After a specified period of exposure to groundwater flow, the PFM is removed from the well. Next, the sorbent is carefully extracted to quantify the mass of all organic contaminants intercepted by the PFM and the residual masses of all resident tracers. The contaminant masses are used to calculate time-averaged or cumulative contaminant mass fluxes, while residual resident tracer masses are used to calculate time-averaged or cumulative water flux. Depth variations of fluxes can be measured by vertically segmenting the exposed sorbent at specified depth intervals.

PFMs were installed in monitoring wells following the groundwater sampling event. PFMs remained in the wells for approximately two to four weeks, and then were collected and analyzed by the University of Florida. A description of the PFM protocol is provided in **Appendix D**.

# 5.2.5 Analytical Methods, Sample Preservation, and Containers

The groundwater samples were transported to laboratories for analysis of several parameters. The University of Tennessee received Sterivex cartridges with biomass collected from the monitoring wells. The PFM data were analyzed at the University of Florida. Pace Analytical Laboratory conducted CSIA analysis.

The analytical methods and sample preservation methods used are summarized in **Table 5-1**. Additional details were included in Sampling Plans prepared for each site.

Analyte	Method/ Laboratory	Preservative	Containers		
VOCs	EPA 8260 APTIM	4°C with HCl	40 mL VOA vial (x3)		
Anions	EPA 300.0 APTIM	4°C	120 mL polyethylene screw- cap (x1)		
Reduced Gases	EPA 3810, RSK 175 APTIM	4°C with HCl	40 mL VOA vial (x2)		
Dissolved Hydrogen EPA 3810, RSK 175 APTIM		4°C with HCl	125 mL glass serum bottle with Teflon-lined cap and crimp seal		
тос	EPA 415.1 APTIM	4°C with H <sub>3</sub> PO <sub>4</sub>	120 mL polyethylene screw- cap (x1)		
Volatile Fatty Acids	EPA 300m APTIM	4°C	40 mL VOA vial (x2)		
Microbial Community	RD-qChip / University of Tennessee	4°C	1 L glass bottle filtered through a Sterivex cartridge		
$\begin{array}{c} \textbf{Carbon Isotopes in} \\ \textbf{CSIA (carbon)} \\ \textbf{Pace Analytical} \end{array} \\ \end{array} \\ \begin{array}{c} \textbf{Carbon Isotopes in} \\ \textbf{dominant VOCs } (\delta^{13}\textbf{C}) \\ \textbf{Pace Analytical} \end{array} \\ \end{array}$		4°C with HCl	40 mL VOA vial (x9)		
Metals - dissolved (Fe, Mn)	EPA 6010 Pace Analytical	4°C with HNO <sub>3</sub>	120 mL polyethylene screw- cap (x2)		
DO, ORP, pH, Turbidity	Feld Meter				

# Table 5-1. Analytical Methods, Preservation, and Containers.

## 5.3 SAMPLING RESULTS

## 5.3.1 Former Raritan Arsenal AOC 2

## 5.3.1.1 Site Description, Former Raritan Arsenal AOC 2

Site AOC 2 at the former Raritan Arsenal previously had a TCE plume that extended approximately 3,000 ft downgradient. The primary source of the plume was Area 18C, the Building 256 Ramp Area (Figure 5-2). Building 256 was formerly used for vehicle maintenance. The specific manner of disposal is unclear, but presumably was the surface disposal of solvents at the Ramp Area.

Two sand units are present beneath the site. The Upper Sand contains sand and silty sand and is approximately 20 ft thick (**Figure 5-3a**). A 5-ft thick silty unit separates the Upper Sand from the Lower Sand. The Lower Sand is approximately 10 ft thick (**Figure 5-3b**). The depth to water is approximately 10 ft below ground surface (bgs). Aquifer pumping tests indicate the Upper Sand has a relatively low hydraulic conductivity of 2.3 ft per day, whereas the Lower Sand has a higher hydraulic conductivity of 50 ft per day. Baseline groundwater sampling indicates the aquifer is generally aerobic and acidic, with pH values ranging from 3.3 to 7.4.

Excavation of contaminated soil occurred in 1998 and 2002, and groundwater TCE concentrations subsequently decreased from approximately 3,000  $\mu$ g/L to 100  $\mu$ g/L from 1998 to 2007. Bioremediation was implemented in the source area in 2009 to reduce contaminant concentrations further. The system included recirculation via nine injection wells and nine extraction wells in the Lower Sand. Sodium lactate and nutrients were introduced as amendments along with the SDC-9 bioaugmentation culture. Sodium and potassium hydroxide were used to raise the pH to near neutral levels, to facilitate bioremediation. The recirculation system operated for a 10-month period. Near the end of the recirculation period, an emulsified oil substrate was introduced into the Upper Sand to provide a longer-term, slow release of electron donor to promote biodegradation where recirculation was not applied.

The bioremediation at the site substantially decreased the concentrations of contaminants. Longterm monitoring is currently occurring at the site with natural attenuation selected as the remedy for the remainder of the plume. No further active remediation is planned.

The field investigations at this site assessed rebound in contaminant concentrations since the remediation activities in 2009, and the persistence of the dechlorinating microbial community in a low pH aerobic aquifer environment.

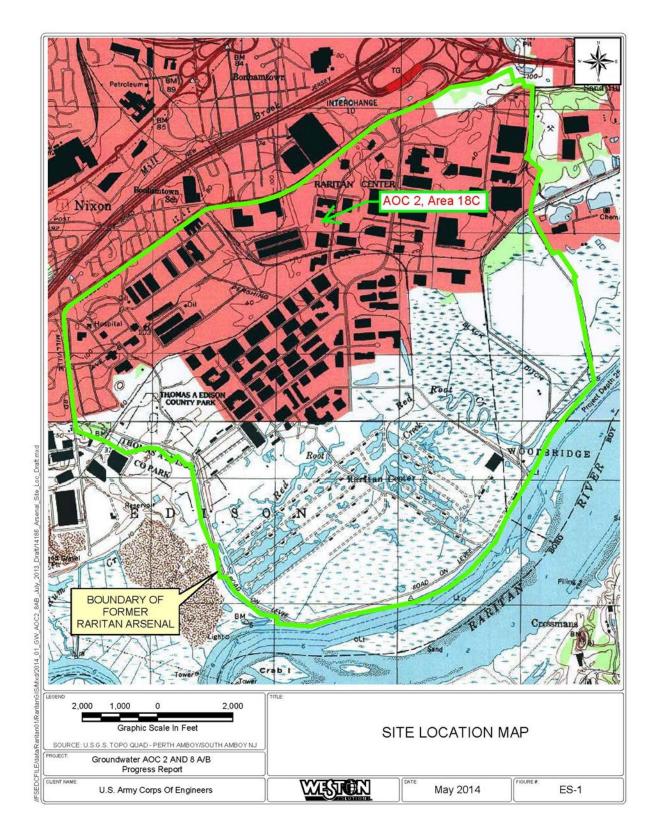


Figure 5-2. Site Location Map, Former Raritan Arsenal AOC 2.

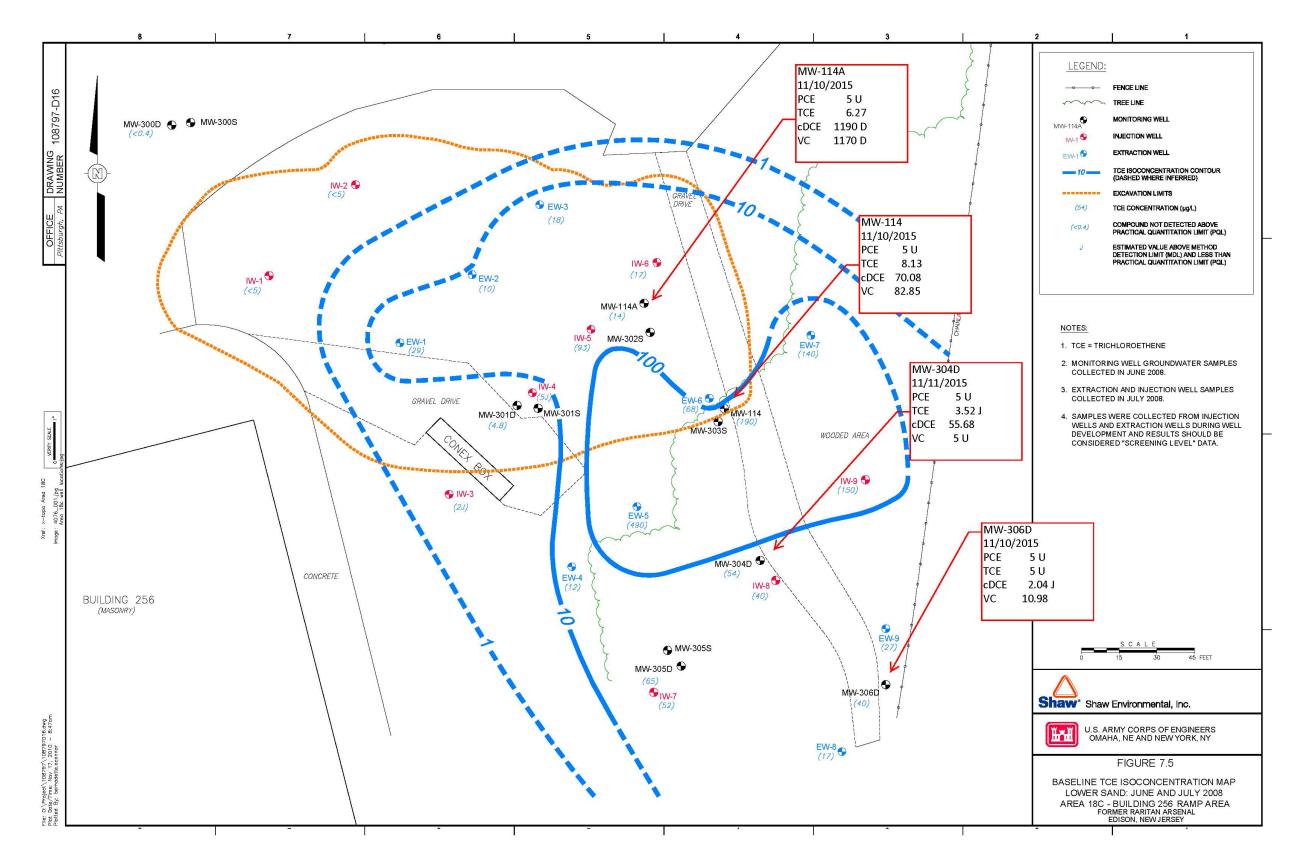


Figure 5-3a. Site Plan, Upper Aquifer, Former Raritan Arsenal AOC 2.

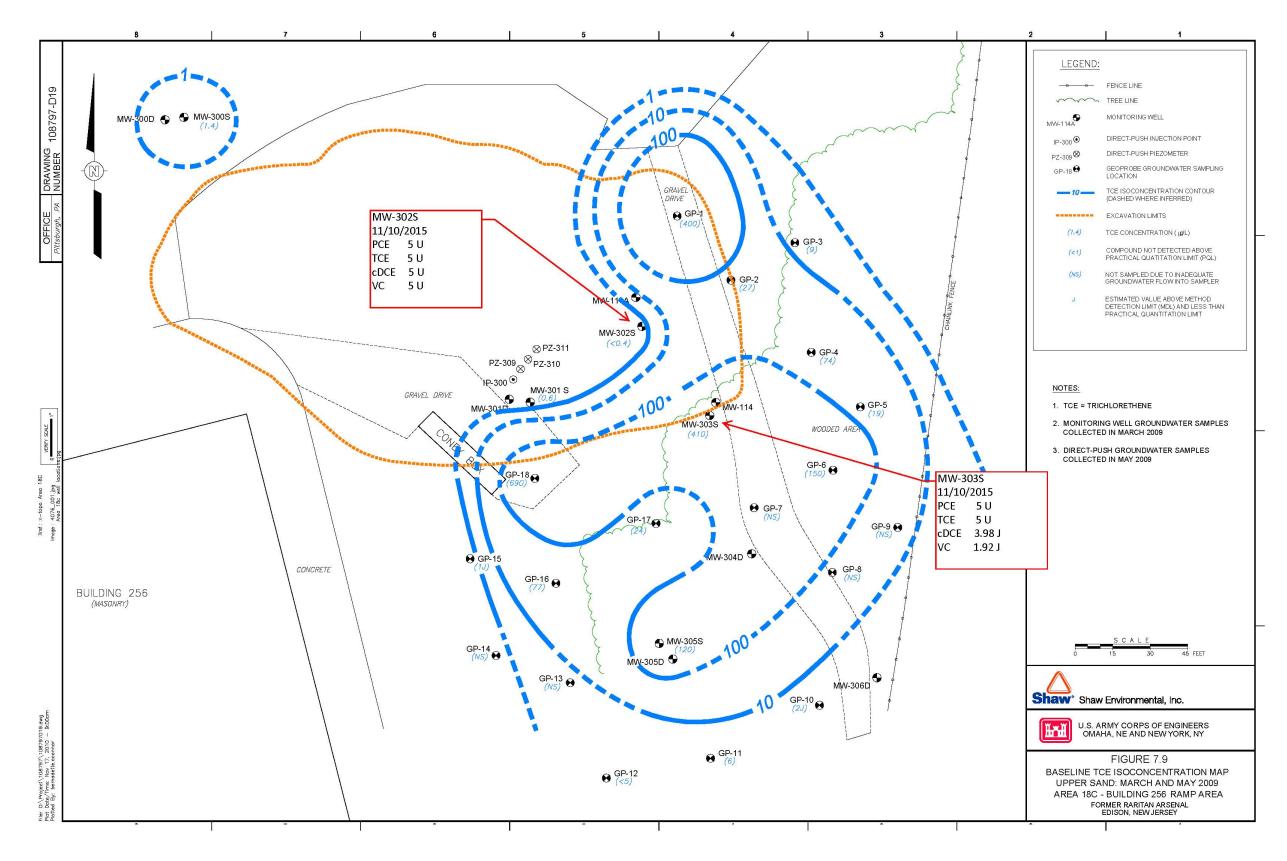


Figure 5-Error! Bookmark not defined.b. Site Plan, Lower Aquifer, Former Raritan Arsenal AOC 2.

## 5.3.1.2 Sampling Rationale, Former Raritan Arsenal AOC 2

The objective of the groundwater sampling was to assess the long-term impacts of the previously performed bioaugmentation treatment on groundwater quality and biogeochemistry. Monitoring locations within and downgradient of the Source Area were evaluated as described below. Monitoring wells sampled are shown in **Figure 5-3a,b**. **Table 5-2** summarizes the monitoring locations and rationale for the sampling program.

Raritan AOC 2, Area 18C Monitoring Well	Sampling Rationale
MW-302S	Within the source area treatment zone and in the Upper Sand. This well represented the long-term effects of the treatment in the Upper Sand.
MW-114A	Within the source area treatment zone and in the Lower Sand. This well represented the long-term effects of the treatment in the Lower Sand.
MW-303S	Located at the downgradient edge of the source area and in the Upper Sand depth horizon. This well monitored contamination and flux exiting the source area in the Upper Sand.
MW-114	Located at the downgradient edge of the source area and in the Lower Sand depth horizon. This well monitored contamination and flux exiting the source area in the Lower Sand.
MW-304D	Downgradient of the treatment zone. This well was evaluated for downgradient effects.
MW-306D	Downgradient of the treatment zone. This well also was evaluated for downgradient effects.

Table 5-2.Groundwater Sampling Locations and Rationale, Raritan Arsenal<br/>AOC 2, Area 18C.

Monitoring wells MW-302S and MW-114A are located within the central part of the original source area treatment zone. MW-303S is screened in the Upper Sand, and MW-114A is screened in the Lower Sand (**Table 5-3**). Monitoring wells MW-303S and MW-114 are located at the downgradient edge of the source area and treatment zone.

The above wells were used to evaluate contaminant concentrations and mass flux exiting the source area. Substantial degradation was observed in these wells during the active treatment period that extended from 2008 to 2010; total VOC concentrations were reduced between 74% and 100% in the various wells. Sampling was conducted approximately 5.5 years after the active treatment, and data from these wells was used to evaluate the long-term effectiveness of the treatment.

Monitoring wells MW-304D and MW-306D are screened in the Lower Sand and located downgradient of the Source Area at distances of approximately 60 and 130 ft, respectively. These wells were used to evaluate conditions downgradient. No degradation was observed in monitoring well MW-304D during the active treatment period, whereas a 90% reduction in total VOCs was observed in monitoring well MW-306D. Data from these wells was used to evaluate changes in downgradient conditions that may have occurred 5.5 years after active treatment.

Monitoring Well	Total depth (bgs)	Top of Screen (bgs)	-			
MW-302S	23	13	23	2		
MW-114A	36	26	36	2		
MW-303S	20	10	20	2		
MW-114	34	24	34	2		
MW-304D	31	21	31	2		
MW-306D	27	17	27	2		

Table 5-3. Monitoring Well Construction Details, Raritan Arsenal AOC 2.

bgs = ft below ground surface

# 5.3.1.3 Sampling Results, Former Raritan Arsenal AOC 2

**Table 5-4** provides a summary of sampling results and **Figures 5-4** through **5-9** provide historical and project data for VOCs for each well.

# 5.3.1.3.1. VOCs

TCE and cis-DCE were both below detection (PQL 5  $\mu$ g/L) the shallow site monitoring wells (MW-302S and MW-303S) more than 5.5 years after treatment with emulsified vegetable oil and bioaugmentation (**Figure 5-4, 5-5**). For MW-303S, concentrations of TCE and cis-DCE were as high as 400 and 1,000  $\mu$ g/L, respectively, prior to/during treatment. A trace of vinyl chloride (1.9  $\mu$ g/L as a J value) was detected in MW-303S but not in well MW-302S. Among the deep wells, > 1,000  $\mu$ g/L of both cis-DCE and VC were detected in MW-114A, which is a source area well that is furthest upgradient among the deep wells (**Figure 5-6**). Moving downgradient to wells MW-114, MW-304D, and MW-306D, those concentrations declined appreciably, with cis-DCE being below detection (< 5  $\mu$ g/L) and VC being 11  $\mu$ g/L at MW-304D, which is the furthest downgradient well (**Table 5-4; Figures 5-7, 5-8, 5-9**). The concentration data suggest continued attenuation of these chlorinated ethenes along the groundwater flow path. CSIA data from the site confirms this hypothesis (see **Section 5.3.1.3.7**). The presence of ethane and ethene in a subset of wells is also indicative of continuing reductive dehalogenation (**Table 5-4**).

# 5.3.1.3.2. Field Parameters

The ORP of all wells at the site were between +2.5 and -100 mV during sampling, indicative of moderately reducing conditions. The groundwater pH was between 5.8 and 6.7 at all wells except MW304D, where the pH was 4.9, which is below that typically considered optimal for reductive dehalogenation ( $\geq$  pH 5.5; Vainberg et al., 2009). The DO in the wells was measured between 0.09 mg/L at MW-306D and 3.13 mg/L at MW-303S. As is often the case when measuring field parameters, the DO and ORP at the site do not necessarily correspond, as the well with the highest measured DO, also had the second lowest ORP value at -79.5 mV.

# 5.3.1.3.3. Anions

Nitrate and nitrite were both below detection in site groundwater. Sulfate ranged from  $\sim 17$  to 48 mg/L in the deep zone wells and from 5 to 13 mg/L in the shallow wells. Chloride was also somewhat higher in the deep compared to the shallow zone (**Table 5-4**).

# 5.3.1.3.4. Dissolved Iron and Manganese

Dissolved iron was detected in all wells at concentrations ranging from ~ 0.7 mg/L to 13 mg/L. Dissolved manganese was also present in each groundwater well, albeit at lower concentrations than for iron (62 - 389  $\mu$ g/L). The presence of dissolved forms of these two metals is consistent with the measured ORPs, indicating continued reducing conditions at the site.

# 5.3.1.3.5. Dissolved Gases

Methane, ranging from ~ 0.4 mg/L to 6.6 mg/L, was detected in wells across the site, indicating reducing conditions. Ethane and ethene were detected in wells MW-114 and MW-114A at concentrations ranging from 4.4 to 13.4  $\mu$ g/L. Well MW-302S had ethane detected at 104  $\mu$ g/L. Both of these gases often represent the final product of reductive dehalogenation and are indicative of the occurrence of this process. Dissolved hydrogen, which is the ultimate electron donor for *Dehalococcoides*, was not detected in any of the groundwater samples above the PQL of 0.0084  $\mu$ g/L.

# 5.3.1.3.6. Total Organic Carbon and Volatile Fatty Acids

Total organic carbon (TOC) in the different wells ranged from a high of 19.8 mg/L in MW-302S to a low of 2.9 mg/L in MW-304D. None of the common fatty acids associated with fermentation of emulsified oils (e.g., formic, lactic, acetic, propionic) were detected above 1 mg/L (PQL). Thus, the measured TOC presumably represents compounds other than these acids.

# Table 5-4. Sampling Results Summary, Former Raritan Arsenal AOC 2.

	LOCATION_CODE		MW-114		MW-11	4A	MW-302S		MW-303S		MW-30	4D	MW-30	6D
Class	SAMPLE_DATE		11/10/15		11/10/15		11/10/15		11/10/15		11/11/15		11/110/15	
	Parameter	Units	Result		Result		Result		Result		Result		Result	
Anion	Chloride	mg/L	223	D	169	D	20.5		7.23		255	D	89.4	D
Anion	Sulfate as SO4	mg/L	26.5		17.4		5.54		12.7		34.8		47.8	
CSIA	del13C cDCE	ppt	-8.62		-8.2		NA		-8.52		-18.7		22.17	
00111	del13C TCE	ppt	-18.6		-20.1		NA		NA		-13.4		-18.8	J
	DO	mg/L	1.33		1.68		2.19		3.13		0.37		0.09	
Field	ORP	mV	-69.9		-100		-70.8		-79.5		2.5		-45	
	pН	su	5.99		6.1		6.65		6.11		4.94		5.84	
	Ethane	μg/L	4.49		8.59		104			U		U		U
Gases	Ethene	μ <b>g</b> /L		J	13.4		5	-	-	U	-	U	-	U
	Methane	μ <b>g</b> /L	539		1280		6600	D	403		718		546	
Metals	Dissolved Iron	μg/L	5190		699		735		2050		3070		13000	
	Dissolved Manganese	μ <b>g</b> /L	204		308		389		93.6		98.1		62.0	
TOC	Total Organic Carbon	mg/L	5.43		9.25		19.8		9.20		2.90		6.29	
	1,1-dichloroethylene	μg/L	5	U	0.7	J	5	U	5	U	5	U	5	U
	1,2-dichloroethane	μg/L	5	U	5	U	4.4	J	5	U	5	U	5	U
	1,3-dichlorobenzene	μg/L	5	U	5	U	0.9	J	5	U	5	U	5	U
	bromobenzene	μg/L	5	U	5	U	1.0	J	5	U	5	U	5	U
	Cis 1,2- Dichloroethylene	μg/L	70.1		1190	D	5	U	4	J	55.7		2	J
	Methyl tertiary butyl ether	μg/L	5	U	5	U	5	U	5	U	2.4	J	5	U
VOC	o-xylene	μ <b>g</b> /L	5	U	0.5	J	1.1	J	5	U	5	U	5	U
	sec-butylbenzene	μg/L	5	U	5	U	7.9		5	U	5	U	5	U
	tert-butylbenzene	μg/L	5	U	5	U	1.7	J	5	U	5	U	5	U
	trans-1,2-dichloroethylene	μ <b>g</b> /L	5	U	2.0	J	5	U	5	U	5	U	5	U
	trichloroethylene	μg/L	8.1		6.3		5	U	5	U	3.5	J	5	U
	vinyl chloride	μ <b>g</b> /L	82.9		1170	D	5	U	1.9	J	5	U	11.0	
	xylenes (m/p)	μg/L	10	U	10	U	1.2	J	10	U	10	U	10	11

Notes:

NA - Not Analyzed

U - Compound not detected above method practical quantitation limit.

D - Sample was diluted prior to analysis

J - Estimated value above MDL and less than PQL

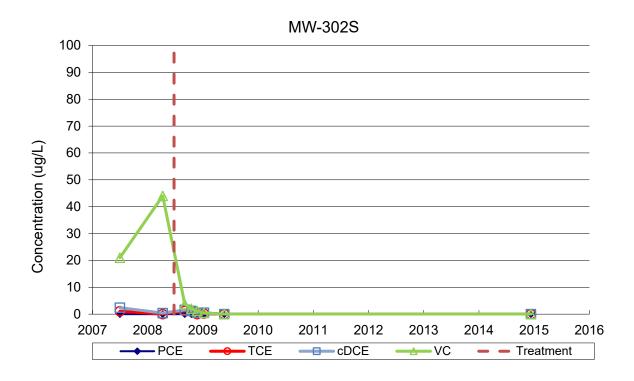


Figure 5-4. VOC Trends for MW-302S, Former Raritan Arsenal AOC 2.

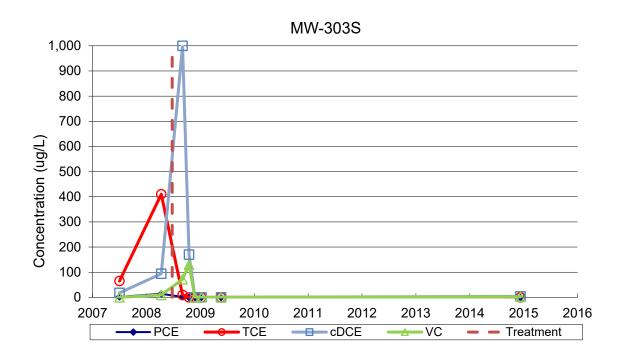


Figure 5-5. VOC Trends for MW-303S, Former Raritan Arsenal AOC 2.

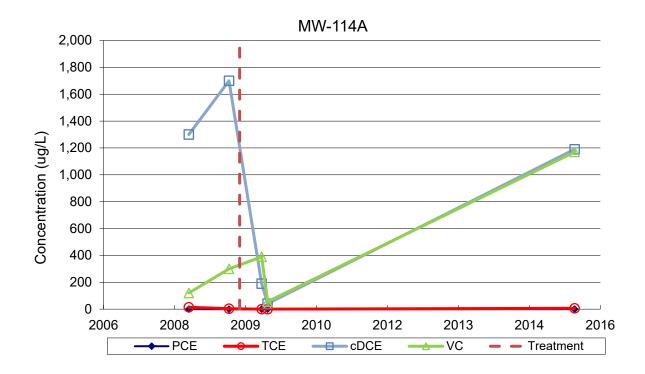


Figure 5-6. VOC Trends for MW-114A, Former Raritan Arsenal AOC 2.



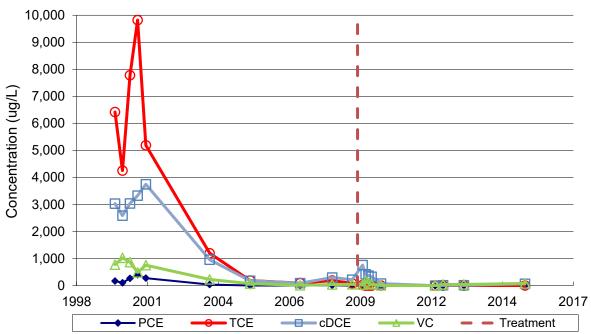


Figure 5-7. VOC Trends for MW-114, Former Raritan Arsenal AOC 2.

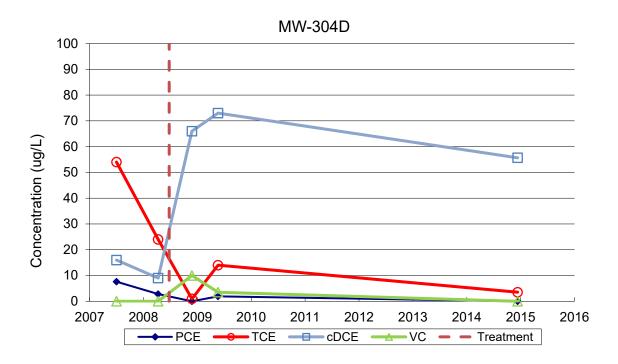


Figure 5-8. VOC Trends for MW-304D, Former Raritan Arsenal AOC 2.

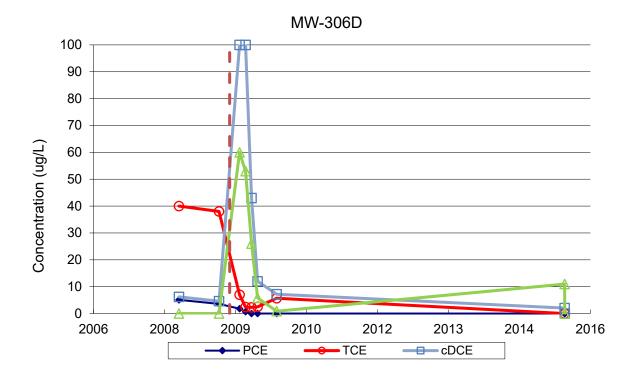


Figure 5-9. VOC Trends for MW-306D, Former Raritan Arsenal AOC 2.

## 5.3.1.3.7. Microbial Community

The microbial community analysis indicated the presence of several different dehalogenationassociated organisms/genes in water samples from the Raritan Site. The greatest diversity of organisms/genes was detected in the deep wells MW-114, MW-114A, and MW-306D (dup) (**Figure 5-10**). Data were not available from samples MW-304D and MW-306D (original not duplicate) due to difficulties with isolating DNA from those samples. *Dehalococcoides* was detected in all of the aforementioned wells, albeit at rather low concentrations ( $\leq 1 \times 10^2$  cells/mL). The vcrA gene and the bvcA gene (both of which encode enzymes that dehalogenate *cis*-DCE) also were detected as was the cobS gene, which encodes a vitamin B12 pathway in *Dehalococcoides*. Interestingly, *Dehaligenomonas* spp, which are known for their ability to dehalogenate chlorinated ethanes, were also detected in each of these wells at concentrations exceeding those of *Dehalococcoides* (10<sup>2</sup> - 10<sup>3</sup> cells/mL). Overall, the data suggest that dehalogenating organisms/genes are present in the deep aquifer zone more than 5.5 years after the last treatment, which included bioaugmentation with SDC-9. This zone also contains the highest concentration of residual VOCs at the site.

*Dehalogenimonas* was also detected in both of shallow wells (MW-302S and MW-303S), and *Dehalococcoides* was detected in one well (MW-302S) but only at ~ 10 cells/mL. No other genes or dehalogenating bacteria were detected. The shallow aquifer showed the lowest residual levels of VOCs. Likely, the much lower levels and/or absence of specific dehalogenating organisms/genes reflects the absence of VOCs required by these organisms for growth. The presence of *Dehalogenimonas*, however, is unusual in that this species is largely associated with chlorinated ethanes, which were never detected at this site to our knowledge.

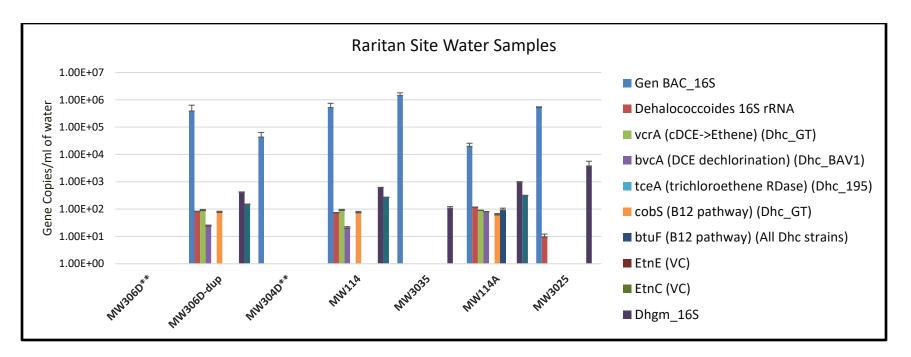


Figure 5-10. Organisms and/or Genes Associated with Reductive Dechlorination Detected in Wells at Former Raritan Arsenal AOC 2.

#### 5.3.1.3.8. CSIA

Values of  $\delta^{13}$ C were obtained for TCE and cis-DCE in monitoring wells where concentrations were sufficient for analysis (**Table 5-5; Figure 5-11**). A typical unfractionated  $\delta^{13}$ C value for manufactured TCE ranges from ~ -34‰ to - 23‰, with a mean of -29‰ (USEPA, 2008). Similarly, for cis-DCE, values range from ~ -30‰ to - 22‰, with a mean of -26‰ (USEPA, 2008). The  $\delta^{13}$ C values for TCE at the Raritan site ranged from -20.1‰ to – 13.4‰, indicating degradation from the parent (i.e., the  $\delta^{13}$ C value is lower/heavier than the range of unfractionated parent values, indicating isotopic enrichment). Similarly,  $\delta^{13}$ C values for cis-DCE at the Raritan site ranged from -18.6‰ to + 22.7‰, indicating significant degradation. In this case, all of the values are significantly heavier than the potential manufactured  $\delta^{13}$ C values for TCE (cis-DCE is a degradation product of TCE), also indicating significant and likely ongoing biodegradation of cis-DCE.

In Figure 5-11, the  $\delta^{13}$ C data for each compound in each deep well, and the corresponding concentrations are provided. The  $\delta^{13}$ C data are subsequently plotted as a function of distance downgradient of the most upgradient well MW-114A in Figure 5-12, and the data are fitted using a linear regression. The slope of each of these lines is subsequently used to estimate the half-lives of the two contaminants, as a simple example of the potential application of CSIA data for such purposes at a site. There are several assumptions made with this simple example as is provided below:

Assuming 1<sup>st</sup> order degradation rates (k), and constant seepage velocity  $(V_x)$  and enrichment factor ( $\varepsilon$ ), then

$$\delta^{13}C = \delta^{13}C_0 - \frac{k * \varepsilon}{V_x} * x$$

DCE: Assuming  $V_x \approx 2$  ft/day and  $\varepsilon \approx -15 \text{ }$ %  $\rightarrow k \approx 2 \times 10^{-2} \text{ d}^{-1}$  and  $t_{1/2} \approx 30 \text{ d}$  for DCE of slope = 0.139

Based on the simple calculations and assumptions shown above, half-lives for the two compounds are estimated at ~ 80 days for TCE and 30 days for cis-DCE. The accuracy of this approach depends on many factors, including the curve fits (which were very good for TCE and less so for cis-DCE), number of overall data points, groundwater velocity estimate, and selection of fractionation factor ( $\varepsilon \approx -15$  ‰ was used) among other variables. However, this exercise shows how half-life data can be determined for compounds via CSIA without using concentration data.

Monitoring Well	δ <sup>13</sup> C TCE (‰)	TCE (µg/L)	δ <sup>13</sup> C DCE (‰)	DCE (µg/L)
MW-114	-20.1	8.1	-8.6	70
MW-114A	-18.6	6.3	-8.2	1190
MW-303D	-13.4	3.5	-18.6	56
MW-306D	NA	< 0.5	+ 22.7	4
MW-302S	NA	< 0.5	NA	< 0.5
MW-303S	NA	< 0.5	-8.5	4

 Table 5-5. CSIA Data, Former Raritan Arsenal AOC 2.

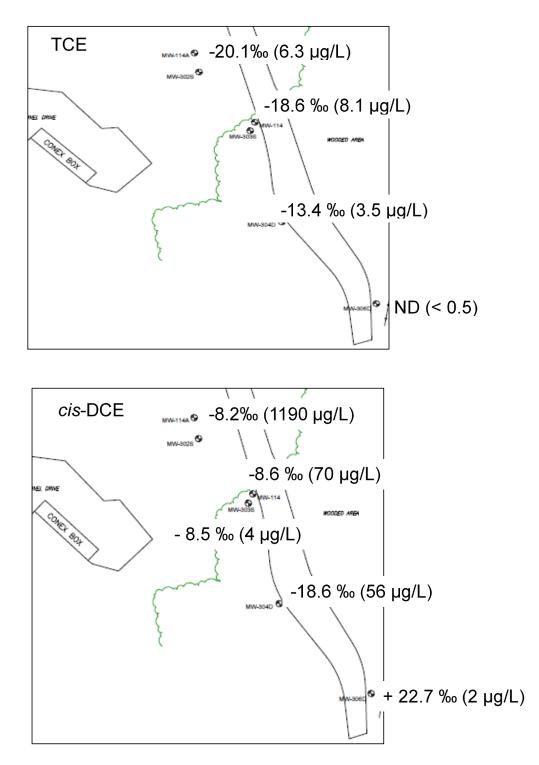


Figure 5-11. Overhead View of the Well Locations, VOC Concentrations (Top Panel; TCE, Bottom Panel; cis-DCE), and δ13C Values for the Relevant VOCs.

ND = Non-detect.

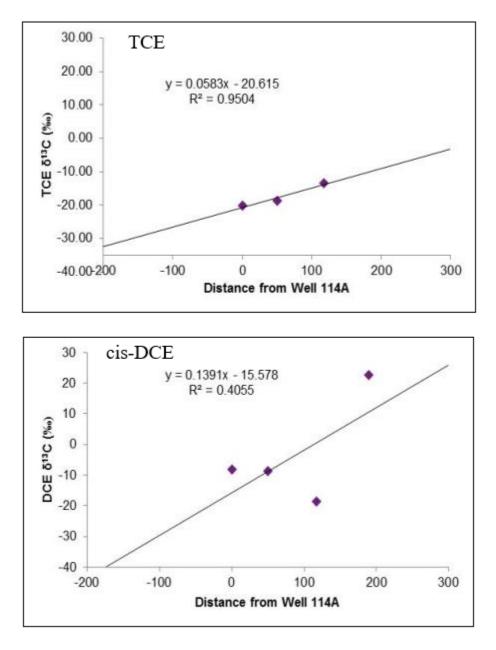


Figure 5-12. Plot of  $\delta^{13}$ C Values for TCE (Top Panel) and *cis*-DCE (Bottom Panel) as a Function of Distance from Upgradient Well 114A.

#### 5.3.1.3.9. Mass Flux

Passive flux meters (PFMs) were deployed in a centerline approach from the source zone down the length of the plume. No chlorinated ethenes or ethene were detected in MW-302S, MW-303S, and MW-306D. VC and cis-DCE, but no ethene, were detected in MW-114A, MW-114, MW-304D. MW-114A, located in the suspected source zone, had the highest fluxes of VC and cis-DCE, which were reported to be about ~67 mg/m<sup>2</sup>/day and 78 mg/m<sup>2</sup>/day. VC and cis-DCE fluxes in MW-114 and MW-304D, located in the plume, ranged from 0.7 to 1.6 mg/m<sup>2</sup>/day (**Table 5-6**).

The PFM flux data can also be used to estimate the flux-averaged contaminant concentration,  $C_f$  (ml<sup>-3</sup>), over the well screen interval/PFM sorbent section using (Basu et al., 2006):

$$C_f = \frac{\int J_{ddz}}{\int q_{ddz}}$$

 $C_f$  is independent of groundwater fluxes, meaning it is not subject to effects of flow convergence towards or divergence around the PFM.  $C_f$  represents a time-averaged contaminant concentration estimate over the deployment period, whereas traditional groundwater sampling techniques yield an instantaneous contaminant concentration that is estimated only at the time the groundwater samples were collected (Basu et al., 2006; Basu et al., 2009).

 $C_f$  values were compared against traditional groundwater concentration data to assess overall remediation performance at sites. In addition to individual contaminant species being repoerted, "Equivalent TCE" was also reported as a measure of overall contaminant mass present at the point sample, to make one to one comparisons of groundwater quality data. "Equivalent TCE" in this study includes degradation products TCE, cis-1,2-dichloroethene (cis-DCE), and vinyl chloride (VC) (Haluska et al., 2019).

The flux-averaged equivalent TCE concentration (TCE+cis-DCE+VC), estimate from the PFM results compared well with existing groundwater monitoring data taken from MW-114A (**Table 5-7**). Downgradient wells MW-114 and MW-304D suggest low concentrations of chlorinated ethenes persist downgradient. The groundwater time concentration profile (**Figure 5-6**) suggests that some rebound occurred in the original source area since the cessation of active treatment, but the increase in VC concentration over time indicates ongoing dechlorination. PFM flux averaged concentration profiles (**Figure 5-13**) corroborate the theory that incomplete dechlorination has occurred as both cis-DCE and VC have accumulated within the source zone. MW-114A and MW-304D groundwater profiles (**Figure 5-6 & 5-8**) and PFM profiles (**Figure 5-13a & 5-13e**) only detected cis-DCE and VC downgradient of the source zone at significantly lower concentrations, consistent with well data. Although no ethene was detected, geochemical analysis suggests aerobic conditions may have persisted, and oxidation of both cis-DCE and VC may be occurring at the site (Fullerton et al., 2013; Gossett, 2010). No chlorinated ethenes or ethene were detected in MW-302S, MW-303S, and MW-306D, suggesting that biological processes have persisted in treating chlorinated solvent after the cessation of active bioremediation in 2009.

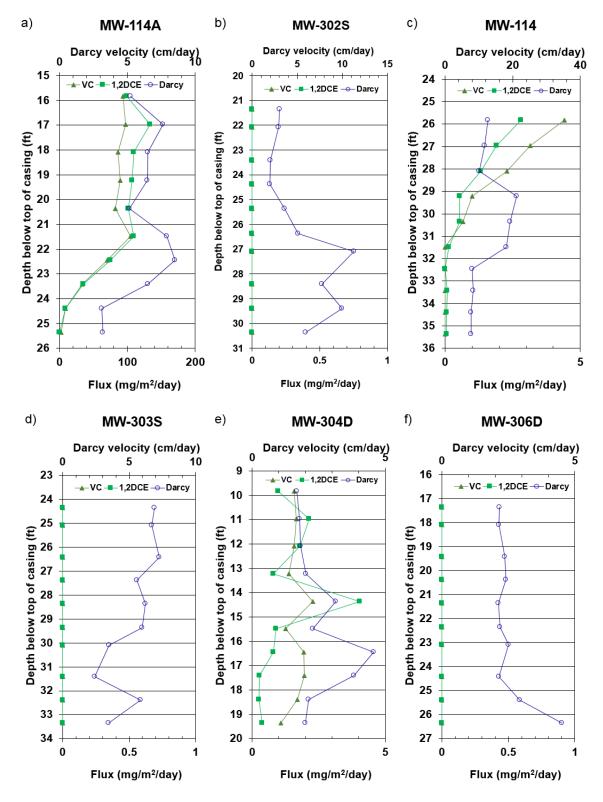


Figure 5-13. Mass Flux Profiles Measured in Select Wells Using the PFMs.

The solid dark green triangles represent VC, the solid light green squares represent DCE, and the open purple circles represent the Darcy Velocity. Panels a-f show flux values sampled on 12/03/2015. Note the changes in scale on both axes to accommodate the data.

Well_ID	Darcy Velocity (cm/day)	VC flux (mg/m²/day)	DCE flux (mg/m²/day)
MW114A	6.0	66.7	77.9
MW114	12.5	1.2	0.7
MW304D	2.5	1.6	1.2
MW303S	5.4	0.0	0.0
MW302S	5.3	0.0	0.0
MW306D	2.5	0.0	0.0

Table 5-6. Average Mass Flux for Each Well.

Well_ID	Equiv. TCE groundwater concentrations (µg/L)	Equiv. TCE PFM flux-averaged concentrations (μg/L)	% Difference
MW114A	4078	3812	6.7
MW114	275	27.6	164
MW304D	79	219.9	94
MW303S	26	<dl< td=""><td>NA</td></dl<>	NA
MW302S	<dl< td=""><td><dl< td=""><td>NA</td></dl<></td></dl<>	<dl< td=""><td>NA</td></dl<>	NA
MW306D	9	<dl< td=""><td>NA</td></dl<>	NA

<sup>a</sup>NA, not analyzed. <dl, less than detection limit.

## 5.3.1.3.10. Site Summary

Groundwater monitoring, CSIA and PFM data suggest that the size of the plume at Raritan has decreased and biodegradation continues 5.5 years after active treatment ended. Source zone groundwater and PFM data show that cis-DCE and VC dominate the source area but dramatic declines in the concentration and flux of both VOCs are apparent downgradient. CSIA data provides further evidence of dechlorination of TCE to cis-DCE and then cis-DCE to VC. Based on the simple calculations using  $\delta^{13}$ C values for TCE and cis-DCE along the groundwater flow path, half-lives for the two compounds were estimated at ~ 80 days for TCE and 30 days for cis-DCE. The microbial community analysis indicated the presence of several different dehalogenation- associated organisms/genes in water samples from the Raritan Site 5.5 years after cessation of active treatment. Dehalococcoides was detected in all of the aforementioned wells, albeit at rather low concentrations (< 1 x  $10^2$  cells/mL). The vcrA gene and the bvcA gene (both of which encode enzymes that dehalogenate cis-DCE) also were detected as was the cobS gene, which encodes a vitamin B12 pathway in *Dehalococcoides*. The data suggest that dehalogenating organisms/genes are present in the deep aquifer zone more than 5.5 years after the last treatment, which included bioaugmentation with SDC-9. This zone also contains the highest concentration of residual VOCs at the site. Overall, site data suggest that removal of organic contaminants is still occurring in the deep zone (shallow zone is clean) and that the potential for anaerobic reductive dechlorination persists, although electron donor is likely limiting at his time. Target treatments with an electron donor, buffer (to ensure optimal pH) and nutrients may be sufficient to promote conditions necessary for complete reductive dechlorination in the deep zone at the site.

# 5.3.2 Dover Air Force Base Area 6

## 5.3.2.1 Site Description, Dover Air Force Base Area 6

Area 6 at Dover Air Force Base was previously characterized by a large plume of chlorinated ethenes and ethanes, approximately 2,500 ft wide and 6,500 ft long. Several source areas were identified, with the most important source area being Site OT14, also known as Building 719. Building 719 was a former jet engine maintenance facility, and trichloroethene (TCE) was used to degrease engine parts in the 1960s. Significant contamination was detected beneath the engine cleaning rooms. Two underground storage tanks (USTs) were formerly present at Building 719 that may have been contributing sources. Maximum TCE concentrations historically exceeded 20,000 micrograms per liter ( $\mu$ g/L). PCE, cis-DCE, and VC were present at lesser concentrations.

The site is underlain by fluvial deposits of the Columbia Formation that extend to a depth of approximately 50 ft bgs. The Columbia Formation contains fine-to-coarse sand with silt and clay lenses and less common lenses of gravel. Monitoring wells are screened in what has been designated the Shallow and Deep Zones of the Columbia Formation. A hydraulic conductivity of 60 ft per day was obtained in an aquifer pumping test, indicating a relatively high permeability. Groundwater is generally oxidizing with a pH of about 5.5 to 6.

A recirculation system for bioremediation was actively operated at Building 719 between 2002 and 2006. The system included twelve injection wells and six extraction wells. Sodium lactate and diammonium phosphate (DAP) were added as biostimulation amendments. The system was converted to a passive mode in 2007 with recirculation temporarily restarted to introduce additional amendments. The system was operated for this purpose on three occasions: September 2008, June 2011 and February 2013.

The bioremediation at Building 719 was generally determined to be successful with a substantial reduction of contamination. The field investigation for this site provided results to evaluate the long-term persistence of source area contamination and associated degradation mechanisms in an aerobic aquifer.

## 5.3.2.2 Sampling Rationale, Dover Air Force Base, Area 6

The objective of the groundwater sampling was to assess the long-term impacts of the previously performed biostimulation-recirculation treatment on groundwater quality and biogeochemistry. Monitoring locations within and downgradient of the source area were evaluated as described below. **Table 5-8** summarizes the monitoring locations and rationale for the sampling program. Monitoring wells sampled are shown on **Figures 5-14** and **5-15**.

# Table 5-8. Groundwater Sampling Locations and Rationale, Dover Air Force Base,<br/>Building 719.

Dover Air Force Base Monitoring Well	Sampling Rationale
MW605S	Within the source area treatment zone and in the Shallow Zone. Treatment was highly effective.
MW605D	Within the source area treatment zone and in the Deep Zone. Treatment was highly effective.
MW608S	Within the source area treatment zone and in the Shallow Zone. Lingering contamination remains.
MW102S	Downgradient of the source area in the Shallow Zone. Treatment was highly effective.
MW102D	Downgradient of the source area in the Deep Zone. Lingering contamination remains.

Monitoring well MW605S is located within the central part of the source area treatment zone and screened in the Shallow Zone. Monitoring well MW605D is also located within the source area treatment zone but screened in the Deep Zone. The treatment has been highly effective at these well locations, with a sustained decrease of VOC concentrations greater than 95%. In contrast, monitoring well MW608S is located in the source area treatment zone, but treatment has not been as effective, and lingering contamination remains.

Similarly, monitoring wells MW102S and MW102D are located downgradient of the source area treatment zone at a distance of approximately 250 ft. Treatment was highly effective at monitoring well MW102S but not at monitoring well MW102D. Monitoring wells MW102S and MW102D area screened in the Shallow and Deep Zones, respectively.

The above wells were selected to evaluate what conditions are associated with highly effective versus less effective treatment. **Table 5-9** presents monitoring well construction details.

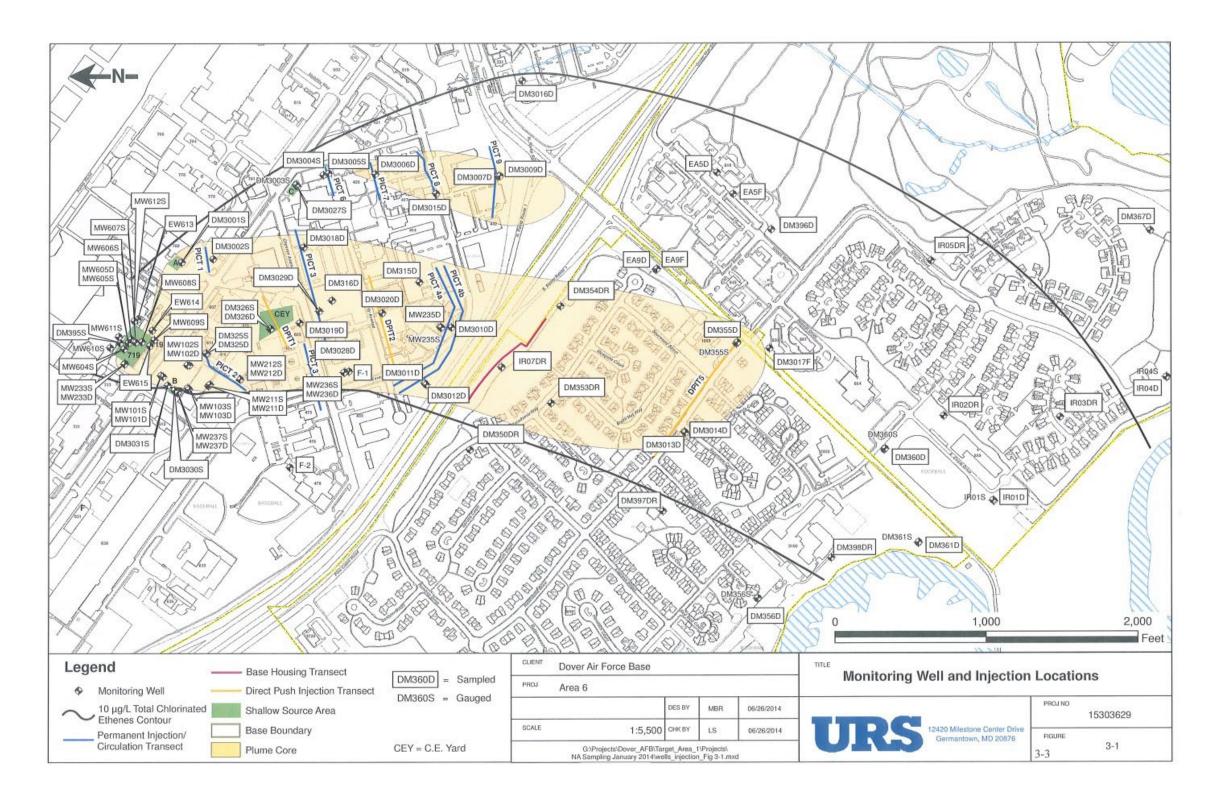


Figure 5-14. Site Location Map Showing the Original VOC Plume, Dover Air Force Base Area 6.

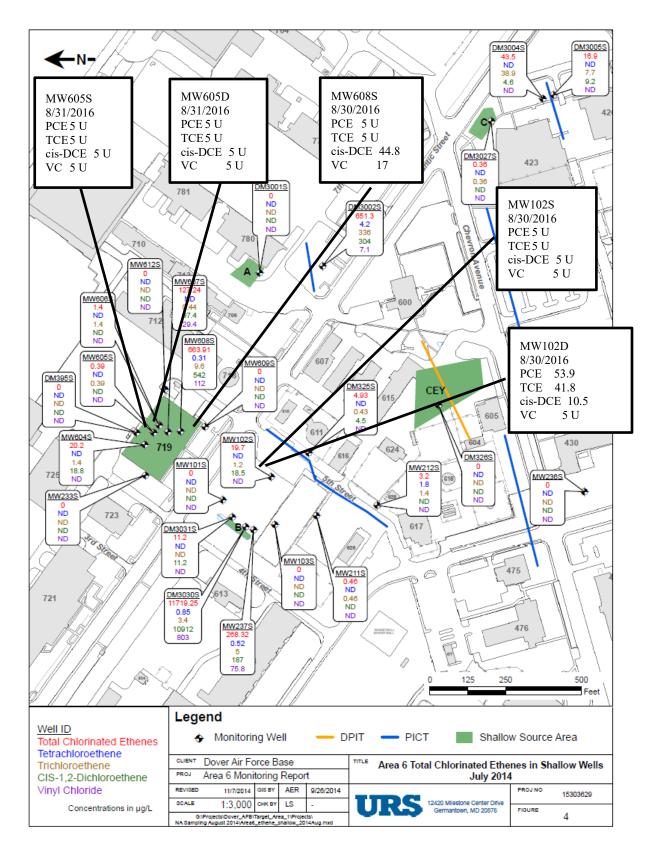


Figure 5-15. Site Plan. Dover Air Force Base Area 6.

Monitoring Well	Total Depth (ft bgs)	Top of Screen (ft bgs)	Bottom of Screen (ft bgs)	Diameter (inches)
MW605S	20	10	20	2
MW605D	30	20	30	2
MW608S	20	10	20	2
MW102S	27	17	27	2
MW102D	40	30	40	4

 Table 5-9. Monitoring Well Construction Details, Dover Air Force Base Area 6.

bgs = below ground surface

### 5.3.2.3 Sampling Results, Dover Air Force Base Area 6

## 5.3.2.3.1. VOCs

**Table 5-10** provides a summary of the sampling results. Among the wells that were sampled, low levels of PCE, TCE, and cis-DCE (10 - 54  $\mu$ g/L) were detected in MW102D, and cis-DCE, VC and 1,1- DCA were detected in MW608S (12 - 45  $\mu$ g/L). These were the same wells that had lingering contamination after the final sampling event in February 2015, with the previous active recirculation treatment with amendment addition ending in 2013. However, it should be noted that MW102D had TCE and PCE exceeding 800  $\mu$ g/L pre-treatment (**Figure 5-16**) and MW608S had concentrations of TCE as high as 7,000  $\mu$ g/L and cis-DCE as high as 100,000  $\mu$ g/L during the first treatment in the early 2000s (**Figure 5-17**). Low concentrations of some chlorinated benzenes were also detected, particularly in MW608S. The other three wells MW102S, MW605D, and MW605S had chlorinated ethenes below detection (< 5  $\mu$ g/L); thus, no rebound was evident, even in wells in the previous source area (MW605D and MW605S). Overall, the biological treatment remedy appears to have been highly effective in this area.

## 5.3.2.3.2. Field Parameters

The ORP of wells at the site had a wide range, from +249 to -153 mV during sampling. The well with the highest ORP, MW102D, was also the only well with residual PCE and TCE, albeit at low concentrations. The groundwater pH was between 6.2 and 7.2 at all wells except MW102D, where the pH was 5.0, which is below that typically considered optimal for reductive dehalogenation ( $\geq$  pH 5.5). Thus, two separate factors appear to contribute to residual VOCs in this well. The DO in the wells was measured between 0.5 mg/L and 1.5 mg/L. In this case, the two wells with the highest ORP values (MW102D and MW102S) also had the highest oxygen concentrations. All of the wells with a negative ORP (-90 to -153 mV) also had measurable oxygen concentrations (0.5-1 mg/L), which is not necessarily consistent with a very negative ORP, as previously noted, but is typical for field data.

## 5.3.2.3.3. Anions

Nitrate was only detected in MW102D (4.12 mg/L and nitrate N), consistent with the high ORP and DO exceeding 1.4 mg/L. All other wells had nitrate < 0.2 mg/L as N. Sulfate ranged from a low of 0.2 mg/L in MW605S to a high of 34.9 mg/L in MW102S. Chloride was also somewhat variable, ranging from 7.28 to 24.7 mg/L in the various wells.

# 5.3.2.3.4. Dissolved Iron and Manganese

Dissolved iron and manganese were detected in some wells, but concentrations were consistently low across the site (<  $600 \mu g/L$  for Fe and <  $115 \mu g/L$  for Mn).

# 5.3.2.3.5. Dissolved Gases

Methane was detected in the different wells, with concentrations ranging from ~ 2  $\mu$ g/L to 22,500  $\mu$ g/L. The highest concentrations were observed in MW605S, MW605D, and MW608S, each having more than 15,000  $\mu$ g/L. The high methane is consistent with the low ORP in these wells (-90 to -115 mV). Each of these wells also had detectable ethane, ranging from 5 to 32  $\mu$ g/L, and one well (MW608S) also had ethene at 8.1  $\mu$ g/L. These gases (ethene and ethene) are often observed as final degradation products of chlorinated ethenes and are indicative of ongoing biodegradation. Hydrogen gas, the ultimate electron donor for reductive dehalogenation, was near or below the MDL (0.0084  $\mu$ g/L) in all wells except MW605D, where 0.021  $\mu$ g/L was detected.

# 5.3.2.3.6. Total Organic Carbon and Volatile Fatty Acids

Total organic carbon (TOC) in the different wells ranged from non-detect (< 2 mg/L) to a high of 9.0 mg/L in MW102S. None of the common fatty acids were detected above 1 mg/L (PQL). Thus, the measured TOC presumably represents compounds other than these acids.

## 5.3.2.3.7. Microbial Community

The microbial community analysis indicated the presence of several different dehalogenationassociated organisms/genes in water samples from the Dover Site. *Dehalococcoides* was detected in wells MW605S, MW605D, and MW608S at concentrations ranging from  $3 \times 10^3$  to  $8 \times 10^4$ cells/mL. These were the wells with high methane, low ORP, and detectable ethane (and ethene in one case). Thus, the data suggest that conditions are conducive to ongoing reductive dehalogenation. The vcrA gene and the bvcA gene, (both of which encode enzymes that dehalogenate cis-DCE) also were detected in two of the three wells (MW605S and MW608S) as were *Dehalobacter* spp., which are known to dehalogenate chlorinated ethanes. Some chlorinated ethanes remain at this site in MW608S.

Conversely, *Dehalococcoides* were not detected in either MW102S or MW102D, nor were any associated genes associated with reductive dehalogenation of chlorinated ethenes. These two wells, only one of which had detectable VOCs (MW102D), also had positive ORPs, low methane, and no detectable ethane or ethene. Thus, geochemical conditions are not conducive to reductive dehalogenation in these wells.

	LOCATION CODE		MW102D	MW102S	MW605D	MW605S	MW608S
Class	SAMPLE DATE		08/30/16	08/30/16	08/31/16	08/31/16	08/30/16
	Parameter	Units	Result	Result	Result	Result	Result
	Chloride	mg/L	24.7	4.42	7.28	23.2	22.0
Anions	Nitrate as N	mg/L	4.12	0.2 U	0.2 U	0.2 U	0.2 U
Anions	Phosphate as P, ortho	mg/L	0.2 U	0.2 U	0.2 U	0.2 U	0.21
	Sulfate as SO4	mg/L	16.6	34.9 D	15.9	0.19 J	18.0
	del13C cDCE	ppt	-21.4	NA	NA	NA	-11.3
CSIA	del13C PCE	ppt	-28.4	NA	NA	NA	NA
COIA	del13C TCE	ppt	-22.4	NA	NA	NA	NA
	del13C VC	ppt	NA	NA	NA	NA	-18
	DO	mg/L	1.46	1.11	0.47	0.68	1.00
Field	ORP	mV	249	57.6	-153	-90.1	-115
	рН	su	5.02	6.16	7.26	6.65	7.05
	Ethane	μg/L	4 U	4 U	5.08	5.47	32.4
Gases	Ethene	μg/L	5 U	5 U	5 U	5 U	8.14
Guebbe	Hydrogen	μg/L	<b>0.006</b> J	0.009	0.021	0.0084 U	0.0084 U
	Methane	μ <b>g</b> /L	<b>1.93</b> J	2 U	22,500 D	<b>19,100</b> D	15,900 D
Metals	Dissolved Iron	μ <b>g</b> /L	70 U	531	70 U	79.4	70 U
	Dissolved Manganese	μ <b>g</b> /L	25.4	7.6	5 U	114	5 U
TOC	Total Organic Carbon	mg/L	2 U	9.04	2.47	2 U	3.06
VFA	Lactic Acid	mg/L	1 U	1 U	1 U	2.98	1 U
	1,1-dichloroethane	μg/L	5 U	5 U	5 U	5 U	12.1
	1,2,4-trimethylbenzene	μ <b>g</b> /L	5 U	5 U	5 U	<b>3.20</b> J	5 U
	1,2-dichlorobenzene	μg/L	5 U	5 U	5 U	5 U	8.34
	1,2-dichloroethane	μ <b>g</b> /L	5 U	5 U	5 U	5 U	<b>2.87</b> J
	1,3-dichlorobenzene	μ <b>g</b> /L	5 U	5 U	5 U	<b>2.37</b> J	<b>1.21</b> J
	1,4-dichlorobenzene	μg/L	5 U	5 U	5 U	11.5	<b>4.62</b> J
VOC	chlorobenzene	μg/L	5 U	5 U	<b>1.62</b> J	53.2	5.70
	chloroethane	μg/L	5 U	5 U	5 U	5 U	16.7
	Cis 1,2- Dichloroethylene	μg/L	10.5	5 U	5 U	5 U	44.8
	naphthalene	μg/L	5 U	5 U	5 U	31.7	5 U
	tetrachloroethylene	μg/L	53.9	5 U	5 U	5 U	5 U
	trichloroethylene	μ <b>g</b> /L	41.8	5 U	5 U	5 U	5 U
	vinyl chloride	μ <b>g</b> /L	5 U	5 U	5 U	5 U	17.0

 Table 5-10.
 Sampling Results Summary, Dover Air Force Base, Area 6.

Notes:

NA - Not Analyzed

U - Compound not detected above method practical quantitation limit.

D - Sample was diluted prior to analysis

J - Estimated value above MDL and less than PQL

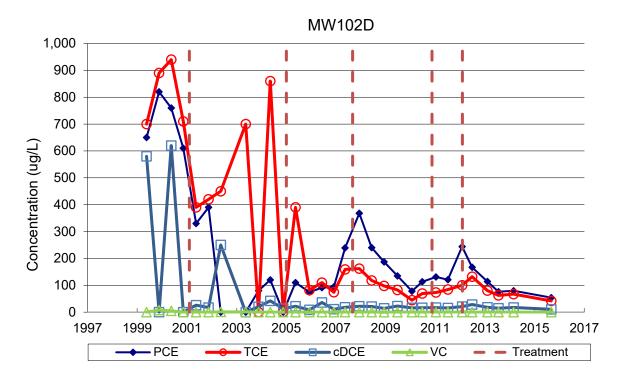


Figure 5-16. VOC Trends for MW102D, Dover Air Force Base Area 6.

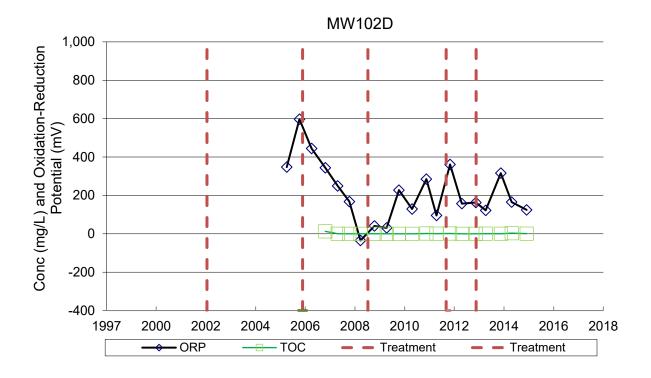


Figure 5-17. ORP and TOC Trends for MW102D, Dover Air Force Base Area 6.

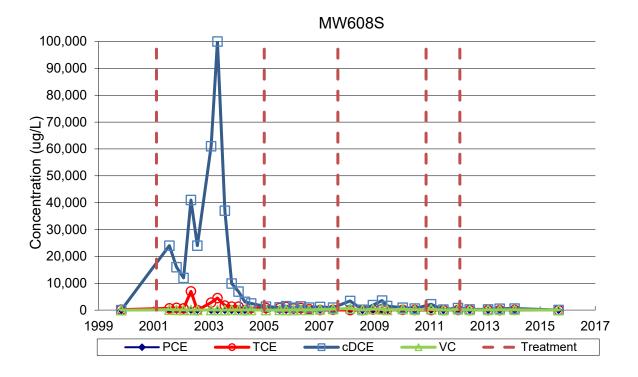


Figure 5-18. VOC Trends for MW608S, Dover Air Force Base Area 6.

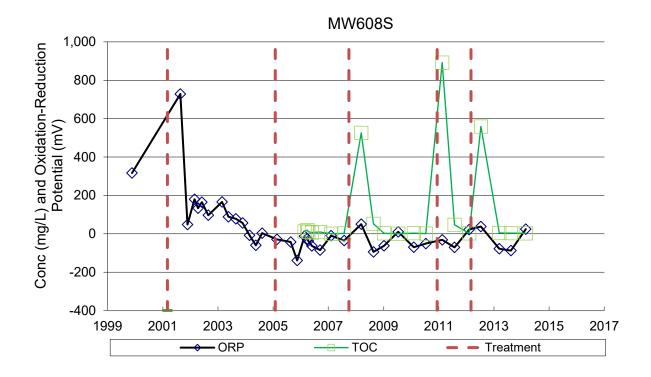


Figure 5-19. ORP and TOC Trends for MW608S, Dover Air Force Base Area 6.

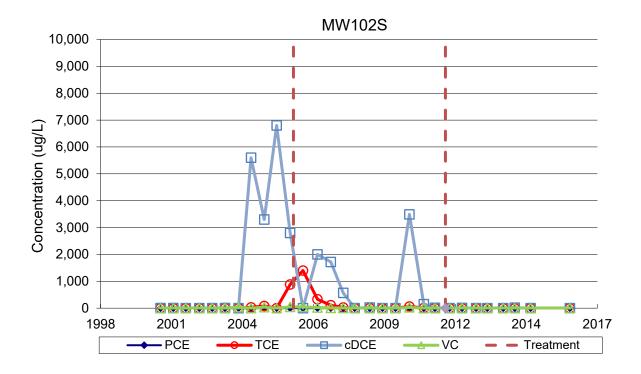


Figure 5-20. VOC Trends for MW102S, Dover Air Force Base Area 6.

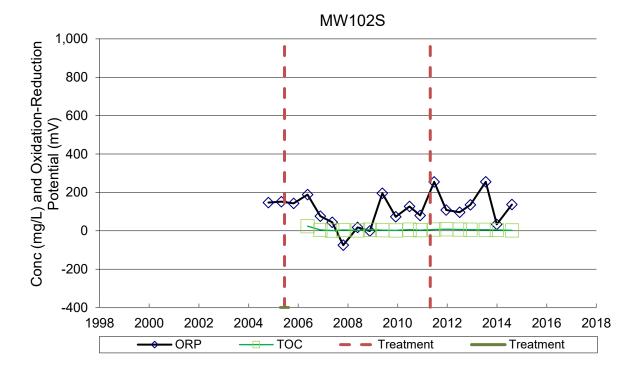


Figure 5-21. ORP and TOC Trends for MW102S, Dover Air Force Base Area 6.

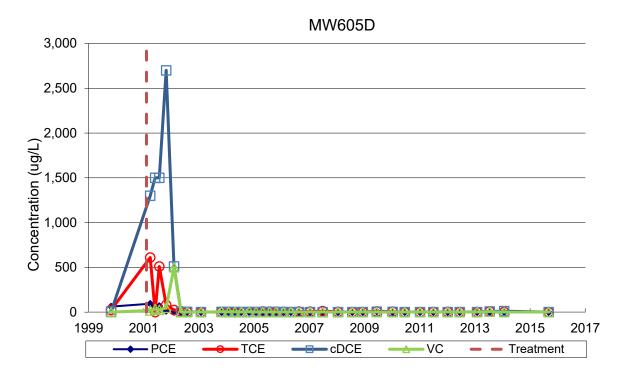


Figure 5-22. VOC Trends for MW605D, Dover Air Force Base Area 6.

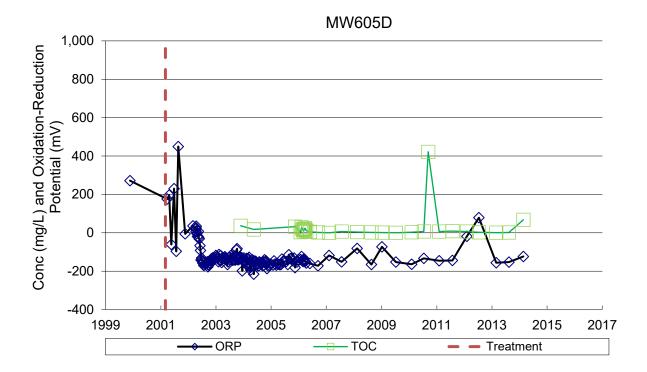


Figure 5-23. ORP and TOC Trends for MW605D, Dover Air Force Base Area 6.

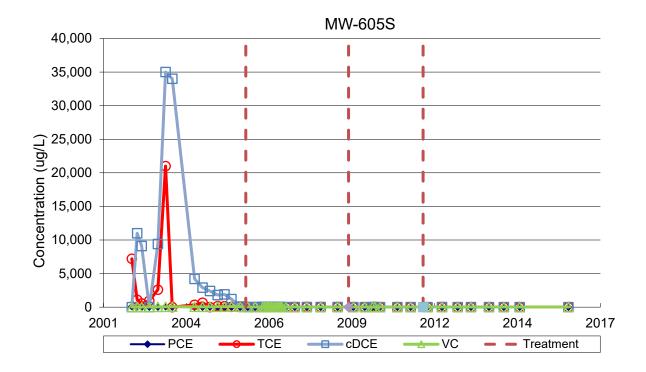


Figure 5-24. VOC Trends for MW605S, Dover Air Force Base Area 6.

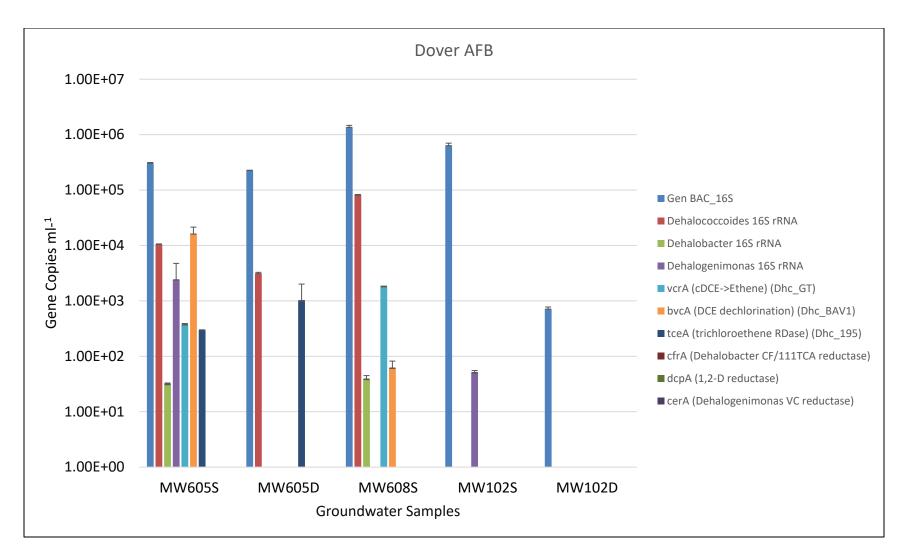
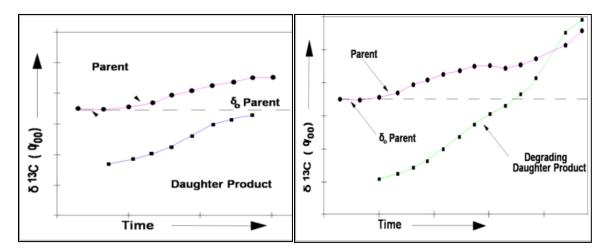


Figure 5-25. Organisms and/or Genes Associated with Reductive Dechlorination Detected in Wells at Dover Air Force Base Area 6.

### 5.3.2.3.8. CSIA

Values of  $\delta^{13}$ C were obtained for PCE, TCE, and *cis*-DCE in MW102D and for cis-DCE in MW608S. Typical unfractionated  $\delta^{13}$ C value for manufactured PCE ranges from~ -37‰ to - 24‰ with a mean of -30%; TCE ranges from  $\sim$  -34% to - 23%, with a mean of -29%; and cis-DCE, values range from ~ -30% to - 22%, with a mean of -26% (USEPA, 2008). The  $\delta^{13}$ C values for PCE, TCE, and cis-DCE in well MW102D were -28.4‰, -22.4‰, and -21.4‰, respectively. The  $\delta^{13}$ C value for cis-DCE in well MW608S was -11.3‰. The  $\delta^{13}$ C value for cis-DCE in MW608S provides clear evidence that the compound is degrading because the delta value is significantly higher (i.e., compound is isotopically heavier or enriched) than the typical range for parent TCE. As shown in Figure 5-26, as a parent VOC (e.g., TCE) biodegrades via reductive dehalogenation, the initial daughter product (cis-DCE in the case of TCE) will initially be much lighter than the parent but will become heavier as the parent biodegrades and also gets heavier. However, if the daughter is not itself biodegrading, it will never become heavier than the original parent TCE (Figure 5-26, left panel). However, if the cis-DCE is biodegrading, it will eventually become heavier than the parent TCE (Figure 5-26, right panel). If one assumes the parent TCE to have a  $\delta^{13}$ C in the range of -34‰ to - 23‰, a value of -11.3‰, as detected in MW608S, is significantly heavier than even the heaviest potential parent value, clearly indicating degradation.

The  $\delta^{13}$ C values of TCE and cis-DCE in well MW102D are heavier than PCE, which may or may not have been the parent of the observed TCE (i.e., TCE may be an initial degradation product of PCE or a parent compound if both were used as solvents). In any event, presuming that TCE is the parent/predecessor of cis-DCE, the  $\delta^{13}$ C value TCE (-22.4 ‰) is just marginally heavier than what would be expected for parent TCE, and cis-DCE is in the same range at -21.4‰. Overall, these data do not support the hypothesis that TCE or cis-DCE is undergoing continued degradation in the vicinity of this well. This conclusion is also consistent with the current geochemistry and microbial data in the groundwater in this region.



# Figure 5-26. Graphs Showing the Expected $\delta^{13}$ C Trajectory in a Case Where the Parent Compound is Biodegrading, and the Daughter Product is Not Biodegrading

(Left Panel). Figure courtesy of Pace Analytical.

#### 5.3.2.3.9. Mass Flux

The contaminant flux profiles, as measured by the PFM deployment, are shown in **Figure 5-27** for the five wells examined. PFMs were deployed in source zone (MW605D, MW605S, MW102D, MW10D) and at the toe of the source zone in the region of the plume (MW608S). No chlorinated ethenes or ethene were detected in wells MW-605D, MW-608S, or MW-102S. No cis-DCE, VC or ethene was detected in MW-102D, but TCE was present, albeit at an average mass flux of 0.02 mg/m<sup>2</sup>/day, and a maximum flux of 0.04 mg/m<sup>2</sup>/day was detected (**Table 5-11**). This maximum mass flux value at well MW-102D is equivalent to a maximum flux averaged TCE concentration of 2.7 µg/L. No PCE or TCE was detected in MW-605S, but the average mass fluxes of 1.19 mg/m<sup>2</sup>/day for cis-DCE, 0.14 mg/m<sup>2</sup>/day for VC and 0.17 mg/m<sup>2</sup>/day ethene (**Table 5-11**). Maximum mass flux concentrations detected in MW-605S included 5 mg/m<sup>2</sup>/day of cis-DCE, 0.72 mg/m<sup>2</sup>/day for VC, and 0.86 mg/m<sup>2</sup>/day for ethene, which translated to maximum flux-averaged aqueous phase concentrations of 280 µg/L cis-DCE, 47 µg/L VC and 56 µg/L ethene (**Table 5-12**).

The flux-averaged equivalent TCE concentration, estimates from the PFM results did not compare well with equivalent TCE groundwater concentration determined using grab samples (**Table 5-13**). Concentrations of contaminants in groundwater withdrawn from wells can be controlled partly by the mass transfer to flowing water during sampling: (1) mass sorbed to aquifer solids; and/or (2) mass trapped in immobile water. Mass transfercan result in extracted groundwater concentrations being underestimated. Flux data averaged across the well screen suggests minimal chloroethene contamination, but target concentrations measured at discrete intervals suggest that dechlorination above MCLs does persist. Taken together, these data suggest that the source of the low concentrations are chlorinated ethenes diffusing from low permeability zone within the aquifer. This suggests that the source zone is likely free of DNAPL and the plume persistence is due to low permeability zone mass release. Finally, as contaminant migrates downgradient dispersion and degradation lower contaminant levels below MCLs.

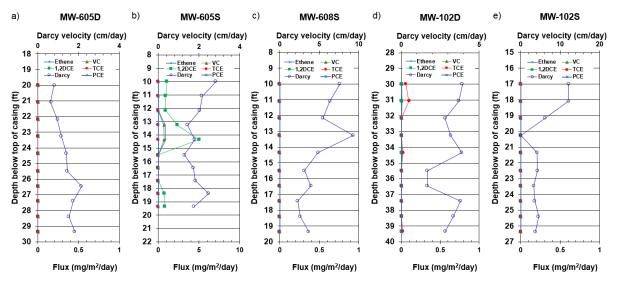


Figure 5-27. Mass Flux Profiles Measured in Select Wells Using the PFMs.

The solid blue diamonds represent ethene, the solid dark green triangles represent VC, the solid light green squares represent DCE, the solid red circles represent TCE, the open purple circles represent the Darcy Velocity, and the blue x's represent PCE. Panels a-e show flux values sampled on 09/22/2016. Note the changes in scale on both axes to accommodate the data.

Well_ID	Average Darcy Velocity (cm/day)	Average Ethene flux (mg/m²/day)	Average VC flux (mg/m²/day)	Average 1,2DCE flux (mg/m²/day)	Average TCE flux (mg/m²/day)
MW-605S	1.93	0.17	0.14	1.19	0.00
MW-605D	1.36	0.00	0.00	0.00	0.00
MW-608S	4.91	0.00	0.00	0.00	0.00
MW-102S	5.41	0.00	0.00	0.00	0.00
MW-102D	3.08	0.00	0.00	0.00	0.02

 Table 5-11.
 Average Mass Discharge for Each Well.

# Table 5-12.Maximum Detected TCE, DCE, and VC Flux Averaged Contaminant<br/>Concentrations for Each Well based on PFMs.

Well_ID	Average Darcy Velocity (cm/day)	Ethene (μg/L)	VC flux (µg/L)	1,2DCE (μg/L)	TCE (µg/L)
MW-605S	1.93	56	47	280	0.00
MW-605D	1.36	0.00	0.00	0.00	0.00
MW-608S	4.91	0.00	0.00	0.00	0.00
MW-102S	5.41	0.00	0.00	0.00	0.00
MW-102D	3.08	0.00	0.00	0.00	2.7

 Table 5-13.
 Equivalent TCE Groundwater vs. PFM Flux-Averaged Concentration.

Well_ID	Equiv. TCE groundwater concentrations (μg/L)	Equiv. TCE PFM flux-averaged concentrations (µg/L)	% Difference
MW-605S	<dl< td=""><td>478</td><td>NA</td></dl<>	478	NA
MW-605D	<dl< td=""><td><dl< td=""><td>NA</td></dl<></td></dl<>	<dl< td=""><td>NA</td></dl<>	NA
MW-608S	96	<dl< td=""><td>NA</td></dl<>	NA
MW-102S	<dl< td=""><td><dl< td=""><td>NA</td></dl<></td></dl<>	<dl< td=""><td>NA</td></dl<>	NA
MW-102D	101	2.7	190

<sup>a</sup>NA, not analyzed. <dl, less than detection limit.

### 5.3.2.3.10. Site Summary

The data collected from the site indicate highly effective bioremediation of chlorinated ethenes across the site. Low concentrations of chlorinated solvents persist in a few of the wells, possibly as a result of diffusion from low permeability materials, but concentrations are orders of magnitude lower than pre-treatment. In wells MW605S, MW605D, and MW608S, *Dehalococcoides* was detected at concentrations ranging from  $3 \times 10^3$  to  $8 \times 10^4$  cells/mL, methane concentrations were relatively high (ranging from 19.1 to 22.5 mg/L) and the presence of ethene and low ORP suggest that conditions are conducive to ongoing reductive dehalogenation. CSIA data supported this conclusion in MW608S (with VOC concentrations too low to measure in the other two wells).

Conversely, in wells MW102S and MW102D, *Dehalococcoides* were not detected, ORP values were positive, methane concentrations were low, and no ethane or ethene were detected, suggesting the conditions were not conducive to reductive dechlorination at these well. However, MW102S had no residual VOCs. MW102D had low concentrations of PCE, TCE, and cis-DCE that, perhaps due to the elevated ORP and low pH, appear to be undergoing further degradation based on CSIA data. Overall, the data from this site suggest that treatment has been highly effective and that reductive dechlorination at the site continues in many areas 5 years after cessation of active treatment.

# 5.3.3 Former Charleston Air Force Base, Zone 4

## 5.3.3.1 Site Description Former Charleston Air Force Base, Zone 4

Zone 4 at Charleston Air Force Base (CAFB) is an industrial area near the CAFB flight line. Trichloroethene (TCE) was formerly used for aircraft parts cleaning (**Figure 5-28**).

Two distinct plumes of TCE contamination, designated Plume 1 and Plume 2, are present at Zone 4. Plume 1 is associated with an oil/water separator at Building 543 (Figure 5-29). TCE was historically present in Plume 1 at concentrations exceeding 100,000  $\mu$ g/L, and the plume extended 600 ft. The origin of Plume 2 is a solvent spill that occurred in 1978 at Building 532 (Figure 5-30). Historical TCE concentrations at Plume 2 exceeded 10,000  $\mu$ g/L. Plume 2 is also approximately 600 ft in length.

Monitoring wells are installed at three depth horizons within the Ladson Formation aquifer consisting of interbedded fine sand, silt, and clay. Groundwater at the site is generally anaerobic and moderately acidic.

Multiple phases of bioremediation have been implemented at Zone 4. The first phase of treatment began in 2003 with the introduction of molasses and bicarbonate buffer using injection wells at Plume 1. Follow-up treatments were initiated in 2004 using molasses, whey, and bicarbonate. This phase of treatment was expanded to include both Plume 1 and Plume 2 and used direct-push technology (DPT) in addition to injection wells. A third phase was implemented in 2008 with the injection of emulsified vegetable oil (EVO) in Plume 1, and a fourth phase was implemented in 2010 with additional EVO injections at both Plume 1 and Plume 2. The EVO injections included both grids and transects of injection wells.

Performance monitoring is currently being conducted. The results to date are somewhat mixed with significant contamination reduction in some areas and minimal reduction in others. Additional injections and/or source remediation are likely in the future. The objective of the field investigation at this site will be to evaluate differences in successful remediation areas versus less successful areas with similar site conditions.

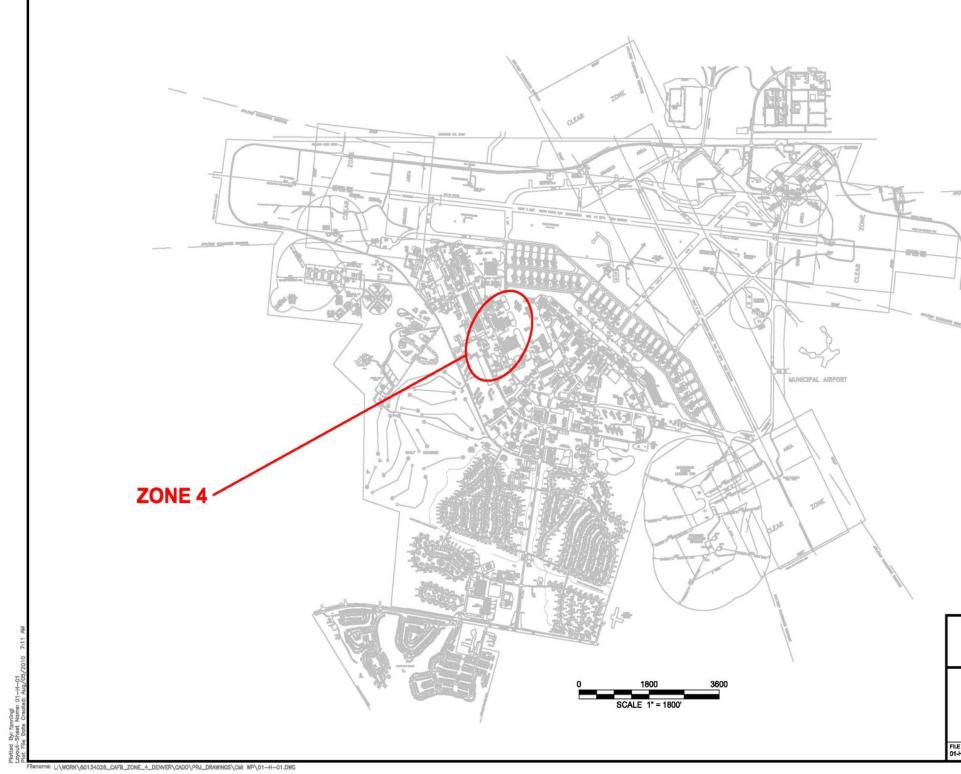


Figure 5-28. Site Location Map, Former Charleston Air Force Base Zone 4.



CORRECTIVE MEASURES IMPLEMENTATION WORK PLAN

### SITE LOCATION MAP

		FORCE BASE SOUTH CARC		
NAME:	DRN	PROJECT NO.	DATE	FIGURE NO.
H-01.dwg	LF	60134026	AUG 2010	1

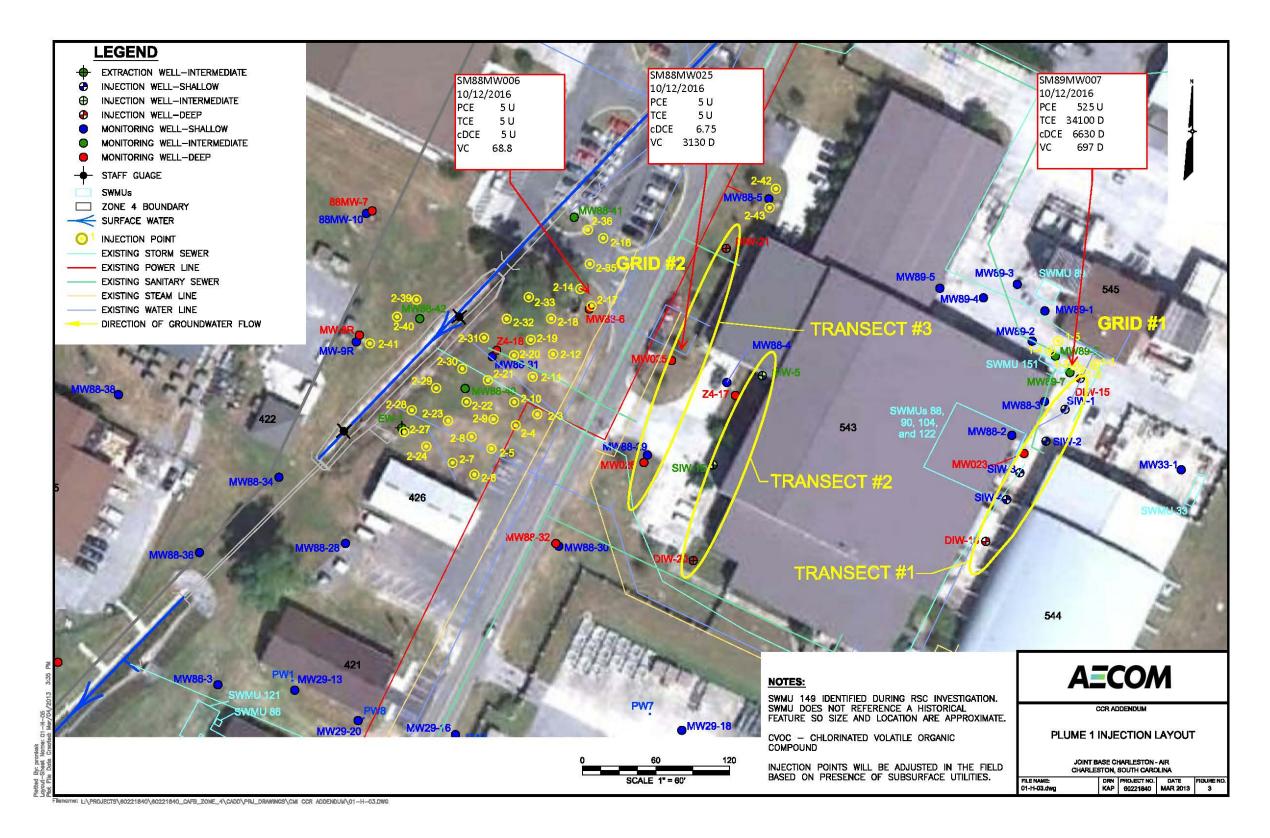


Figure 5-29. Plume 1 Injection Layout, Former Charleston Air Force Base Zone 4.

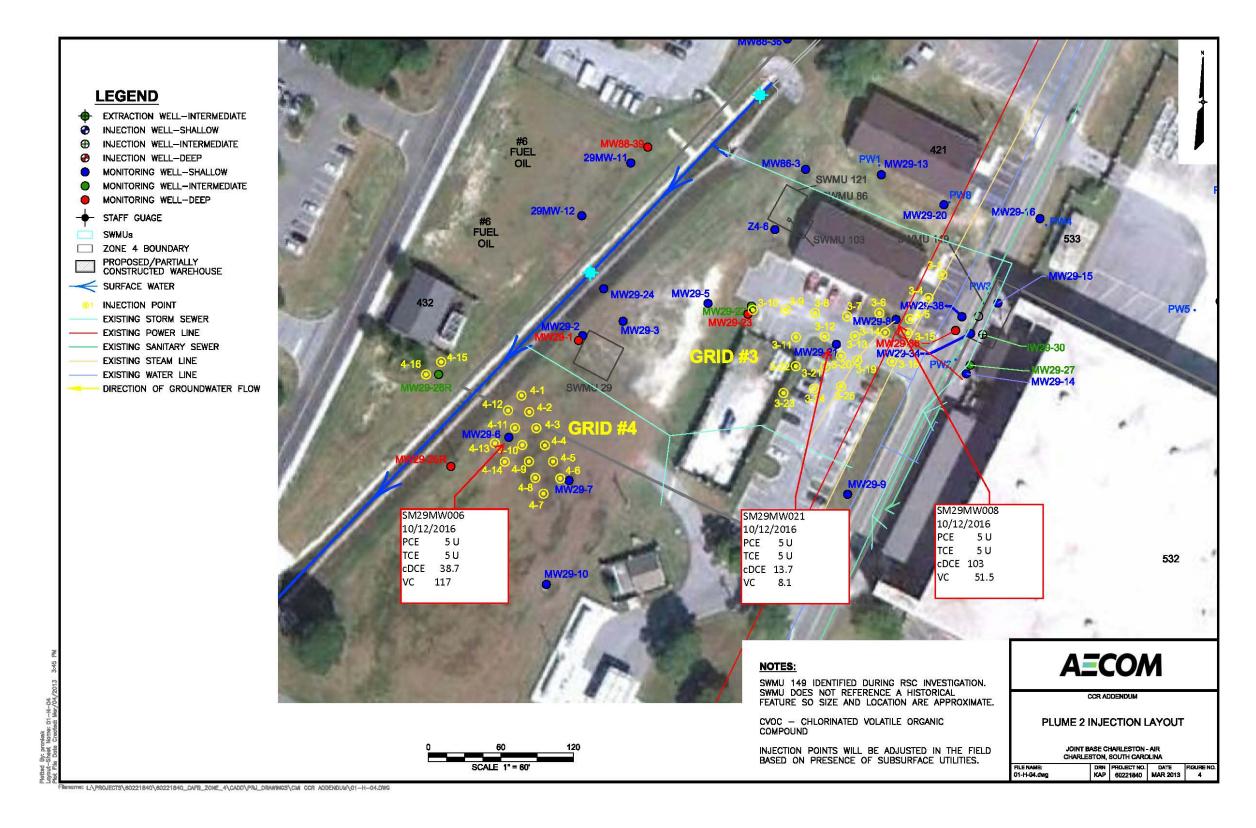


Figure 5-30. Plume 2 Injection Layout, Former Charleston Air Force Base Zone 4.

## 5.3.3.2 Sampling Rationale Former Charleston Air Force Base Zone 4

The objective of the groundwater sampling was to assess the long-term impacts of the previously performed biostimulation treatments on groundwater quality and biogeochemistry. Amendments were added during various treatment phases between 2004 and 2010, and contaminant reduction in the individual wells was variable. The sampling included wells with varying degrees of contaminant reduction to investigate factors that may contribute to treatment effectiveness. Also, monitoring wells were selected from the various depth horizons and treatment areas at the site. **Table 5-14** summarizes the monitoring locations and rationale for the sampling program and contains general comments on contaminant reduction. **Table 5-15** provides a summary of well construction details. The monitoring wells that were sampled are shown in **Figures 5-29** and **5-30**.

Monitoring Well	Contaminant Reduction	Depth Zone	Treatment Area
MW025	Negligible	Deep	Plume 1, Transect 3
MW29-8	Substantial	Shallow	Plume 2, Grid 3
MW29-21	Moderate	Shallow	Plume 2, Grid 3
IW29-30	Negligible	Intermediate	Plume 2, Grid 3
MW29-6	Moderate	Shallow	Plume 2, Grid 4
MW88-06	Substantial	Deep	Plume 1, Grid 2
MW89-07	Negligible	Shallow	Plume 1, Grid 1

Table 5-14.Groundwater Sampling Locations and Rationale, Former Charleston Air<br/>Force Base Zone 4.

Table 5-15.	Monitoring Well Construction Details, Former Charleston Air Force Base
	Zone 4.

Monitoring Well	Top of Screen (ft bgs)	Bottom of Screen (ft bgs)	Diameter (inches)		
MW025	25	35	2		
MW29-8	10	20	2		
MW29-21	22	27	2		
IW29-30	21	26	2		
MW29-6	10	20	2		
MW88-06	30	35	2		
MW89-07	22	27	1		
Z4-17*	26.5	31.5	2		

bgs = below ground surface

\*PFMs only deployed in this well. VOCs, CSIA, and geochemical analysis were not performed on wells due to scheduling conflicts.

## 5.3.3.3 Sampling Results, Former Charleston Air Force Base Zone 4

## 5.3.3.3.1. VOCs

**Table 5-16** provides a summary of sampling results, and **Figures 5-31** to **5-46** provide historical and current VOC data and corresponding ORP and TOC data for each well. The primary parent contaminant at this site was TCE. Concentrations were below detection ( $< 5 \mu g/L$ ) in 5 of the 7 wells sampled. The remaining wells, IW29-30 and MW29-07 had 62.6  $\mu g/L$  and 34,100  $\mu g/L$  of TCE, respectively (**Figure 5-31** and **5-32**). The latter well is a shallow source zone well where treatment has been relatively ineffective, most likely due to the continued presence of DNAPL based upon the dissolved TCE concentrations. Among the wells, *cis*-DCE was also detected at the highest concentrations in IW29-30 (1,160  $\mu g/L$ ) and MW29-07 (6,630  $\mu g/L$ ). Among the other 5 wells, *cis*-DCE ranged from below detection ( $< 5 \mu g/L$ : MW88-6) to 103  $\mu g/L$  in MW29-8. VC was detected at the highest concentration in MW-025 (3,130  $\mu g/L$ ), a well without TCE and with only 6.75  $\mu g/L$  or *cis*-DCE, indicating substantial reductive dechlorination, but perhaps with a much slower rate of VC reduction. All the other wells also VC, ranging from a low of 8.1  $\mu g/L$  in MW29-21 to 697  $\mu g/L$  in MW89-7. The prevalence of daughter products indicates ongoing reductive dechlorination.

### 5.3.3.3.2. Field Parameters

The ORP of wells across the site ranged from +40.7 to -137 mV, and DO was consistently low (0.1 to 0.42 mg/L). Thus, the conditions across the site were anoxic and mildly reducing. The historical ORP at this site, as shown in the accompanying figures for each well (see Figures 5-31 to 5-46) have also been in this general range since 2007, with some expected decreases after the addition of organic substrate. The groundwater pH was between 5.9 and 6.8 SU, which is in a range suitable for reductive dehalogenation to occur.

### 5.3.3.3.3. Anions

Nitrate, nitrite, and phosphate were all below detection (< 0.2 mg/L) across the site, and sulfate ranged from 0.14 to 2.42 mg/L. Previous data collected in 2003 showed sulfate ranging from a high of 34 mg/L to a low of ~1.5 mg/L across the site, indicating that sulfate is not naturally elevated in site groundwater.

### 5.3.3.3.4. Dissolved Iron and Manganese

Dissolved iron was high across the site, ranging from 3,510  $\mu$ g/L in MW89-7 to 59,100  $\mu$ g/L in IW29-30. Similarly, dissolved manganese was detected in every well ranging from 148  $\mu$ g/L in MW89-6 to 794  $\mu$ g/L in IW29-30. These data are indicative of iron and manganese reduction occurring in site groundwater.

## 5.3.3.3.5. Dissolved Gases

Methane was detected in groundwater across the site ranging from ~ 2,480  $\mu$ g/L to 9,510  $\mu$ g/L. The highest concentrations were observed in MW-025 and MW-89-7, each having more than 9,000  $\mu$ g/L. Each of these wells also had detectable ethane, ranging from 4.1  $\mu$ g/L in MW29-21 to 2,180  $\mu$ g/L in MW-025. Moreover, all wells except MW29-21 had ethene ranging from 17  $\mu$ g/L in MW29-6 to 5,840  $\mu$ g/L in MW-025. These gases (ethene and ethene) are typically detected as final degradation products of chlorinated ethenes and are indicative of ongoing reductive dehalogenation.

Hydrogen gas, the ultimate electron donor for reductive dehalogenation, was near or below the MDL (0.0084  $\mu$ g/L) in all wells except MW-025 (0.0785  $\mu$ g/L) and MW88-6 (0.028  $\mu$ g/L).

# 5.3.3.3.6. Total Organic Carbon and Volatile Fatty Acids

Total organic carbon (TOC) in the different wells ranged from 1.48 mg/L in MW29-8 to 204 mg/L in IW29-30. None of the common fatty acids were detected above 1 mg/L (PQL) except acetate, which was measured at 10.2 mg/L in MW29-21 and 46.2 mg/L in IW29-30. Thus, with the exception of these two wells, the measured TOC largely represents compounds other than fatty acids. Historical TOC data are provided in accompanying figures (Figures 5-31 to 5-46).

## 5.3.3.3.7. Microbial Community

The microbial community analysis indicated the presence of numerous different dehalogenationassociated organisms/genes in water samples from across the Charleston Site. *Dehalococcoides* was detected in water from wells MW29-21, MW29-8, MW-025, MW88-6, MW29-6, and IW29-30 at an abundance of between 9 x 10<sup>1</sup> cells/mL (MW29-21) to 3.3 x 10<sup>6</sup> cells/mL in IW-29-30 (**Figure 5-47**). The qPCR analysis indicated that *vcrA* gene (enzyme degrades cis-DCE to ethene) was present in MW29-8, MW-025, MW88-6, MW29-6, and MW29-30 at 1.4 x 10<sup>2</sup> cells/mL (MW88-6) to 4.2 x 10<sup>5</sup> cells/mL in IW-29-30.The *bvcA* gene, whose coded enzyme performs the same function as *vcrA* was only detected in three wells, but concentrations in IW29-30 were very high (3.1 x 10<sup>5</sup> cells/mL) as observed for *vcrA*. The *tceA* gene (enzyme degrades TCE to VC) was detected in MW29-8, MW-025, MW29-6, and IW29-30 at 4.5 x 10<sup>1</sup> cells/mL (MW29-8) to 3.0 x  $10^4$  cells/mL (IW29-30). *Dehalobacter* spp., known for degrading chlorinated ethanes, were detected in samples from MW88-6, MW29-6, and MW29-30 in amounts ranging from 6.0 x 10<sup>1</sup> cells/mL to 1.0 x 10<sup>6</sup> cells/mL, and *Dehalogenimonas* (also chlorinated ethane degraders) were observed in MW29-8, MW-025, MW88-6, MW29-6, and MW29-30 in amounts from 6.4 x 10<sup>1</sup> cells/mL to 1.7 x 10<sup>4</sup> cells/mL.

Overall, qPCR results showed a wide array of dehalogenating organisms and genes in the samples. The numbers and diversity were generally greatest in IW29-30 followed by MW-025, MW29-6, and MW29-8. MW29-21 and MW89-7 had the lowest overall numbers and cell/gene diversity.

	LOCATION CODE		MW-025	MW29-21	MW29-30	MW29-6	MW29-8	MW88-6	MW89-7
Class	SAMPLE DATE		10/12/16	10/12/16	10/12/16	10/12/16	10/12/16	10/12/16	10/12/16
	Parameter	Units	Result	Result	Result	Result	Result	Result	Result
Anion	Chloride	mg/L	92.3 D	11.9	332 D	14.3	13.9	190 D	15.1
	Nitrate as N	mg/L	0.2 U	0.2 U	0.2 U	0.2 U	0.2 U	0.2 U	0.2 U
	Nitrite as N	mg/L	0.2 U	0.2 U	0.2 U	0.2 U	0.2 U	0.2 U	0.2 U
	Phosphate as P, ortho	mg/L	0.2 U	0.2 U	0.2 U	0.2 U	0.2 U	0.2 U	0.2 U
	Sulfate as SO4	mg/L	2.42	0.14 J	0.56	1.47	0.52	0.19 J	0.57
	del13C cDCE	ppt	NA	-13.2	-22	-3.78	-7.84	NA	-25.7
CSIA	del13C TCE	ppt	NA	NA	-21	NA	NA	NA	-25.1
	del13C VC	ppt	NA	-9.29	-25.6	-13.3	-11.9	1.20	NA
Field	DO	mg/L	0.34	0.33	0.1	0.38	0.42	0.23	0.24
	ORP	mV	36.3	-15.5	-137	37.7	-8.6	-42.3	40.7
	pН	su	6.29	5.99	6.81	5.95	5.99	6.52	5.92
Gases	Acetylene	μ <b>g</b> /L	10 U	10 U	10 U	10 U	10 U	10 U	10 U
	Ethane	μ <b>g/L</b>	<b>2,180</b> D	4.09	23.1	4.34	23	332	131
	Ethene	μ <b>g/L</b>	5,840 D	5 U	55.2	17	6.52	82.9	210
	Hydrogen	μ <b>g/L</b>	0.0785 D	0.0084 U	0.0084 U	0.0084 U	0.0084 U	0.028	0.0084 U
	Methane	μ <b>g/L</b>	9,510 D	4,080 D	2,480 D	6,990 D	5,200 D	6,380 D	9,490 D
	Propane	μg/L	6 U	6 U	6 U	6 U	6 U	6 U	6 U
Metals	Dissolved Iron	μg/L	12500	7800	59100	12400	8540	19800	3510
<b>T</b> 00	Dissolved Manganese	μg/L	299	134	794	148	130	278	163
TOC	Total Organic Carbon	mg/L	14.4	26.5	204 D	25.7	1.48 J	160 D	9.02
	Acetic Acid	mg/L	1 U	10.2	46.2	1 U	1 U	1 U	1 U
VFA	Butyric Acid	mg/L	1 U	1 U	1 U	1 U	1 U	1 U	1 U
	Formic Acid	mg/L	1 U 1 U	1 U 1 U	1 U 1 U	1 U 1 U	1 U 1 U	1 U 1 U	1 U 1 U
	Lactic Acid Propionic Acid	mg/L mg/L	10	1 U	1 U	1 U	1 U	10	10
	Pyruvic Acid	mg/L	10	1 U	10	1 U	10	10	10
	Valeric Acid	mg/L	10	10	1 U	10	10	10	10
	1,1-dichloroethylene	μg/L	5 U	5 U	2.30 J	5 U	5 U	5 U	525 U
	1,2,4-trimethylbenzene	μg/L	<b>4.91</b> J	5 U	8.33	5 U	5 U	5.50	525 U
				5 U	5 U	5 U	5 U		
	1,2-dichlorobenzene	μg/L	2.84 J					1.00 J	525 U
	1,3,5-trimethylbenzene	μ <b>g/L</b>	<b>4.26</b> J	5 U	<b>4.44</b> J	5 U	5 U	<b>3.32</b> J	525 U
	1,3-dichlorobenzene	μ <b>g</b> /L	75.9	5 U	5 U	5 U	5 U	45.9	525 U
	1,4-dichlorobenzene	μ <b>g/L</b>	51.8	5 U	5 U	5 U	5 U	20.2	525 U
	benzene	μ <b>g/L</b>	<b>3.3</b> J	5 U	5 U	5 U	5 U	6.15	525 U
	carbon disulfide	μ <b>g/L</b>	5.45	5 U	5 U	5 U	5 U	<b>3.86</b> J	525 U
	chlorobenzene	μg/L	783 D	5 U	5 U	<b>0.97</b> J	5 U	885 D	525 U
VOC	Cis 1,2- Dichloroethylene	μg/L	6.75	13.7	1150 D	38.7	103	5 U	6630 D
	ethylbenzene	μg/∟ μg/L	6.91	5 U	1.00 J	50.7 5 U	5 U	7.21	525 U
	methylene chloride	μg/∟ μg/L	5 U	5 U	1.00 5 5 U	5 U	5 U	5 U	79.8 JD
	o-xylene	μg/∟ μg/L	7.92	5 U	<b>4.28</b> J	5 U	5 U	5.80	525 U
	tert-butylbenzene	μg/L	5 U	5 U 10 U	0.71 J	5 U	5 U	5 U 10 U	525 U
	Tetrahydrofuran (THF)	μg/L	10 U		93.1	10 U	87.0		1050 U
	toluene	μg/L	127	5 U	7.41	5 U	0.45 J	4.64 J	525 U
	trans-1,2-dichloroethylene	μ <b>g/L</b>	5 U	<b>1.64</b> J	<b>0.50</b> J	5 U	7.59	5 U	525 U
	trichloroethylene	μ <b>g/L</b>	5 U	5 U	62.6	5 U	5 U	5 U	34100 D
	vinyl chloride	μ <b>g/L</b>	3130 D	8.10	179 JD	117	51.5	68.8	697 D
	xylenes (m/p)	μg/L	25.8	10 U	11.6	10 U	10 U	27.9	1050 U

 Table 5-16.
 Sampling Results Summary, Former Charleston Air Force Base Zone 4.

Notes:

NA - Not Analyzed

U - Compound not detected above method practical quantitation limit.

D - Sample was diluted prior to analysis

J - Estimated value above MDL and less than PQL

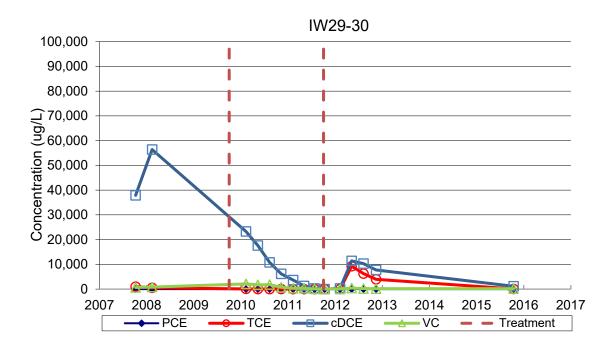


Figure 5-31. VOC Trends for IW29-30, Former Charleston Air Force Base Zone 4.

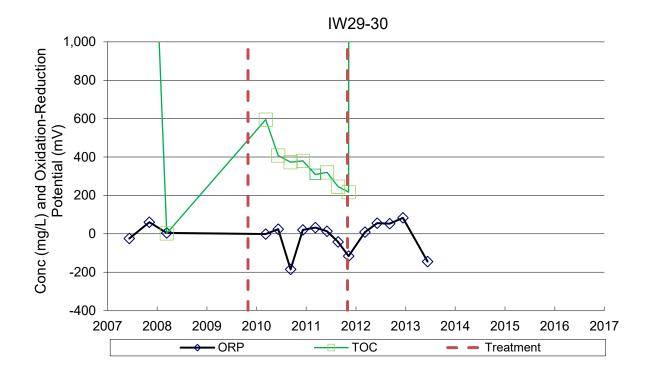


Figure 5-32. ORP and TOC for IW29-30, Former Charleston Air Force Base Zone 4.

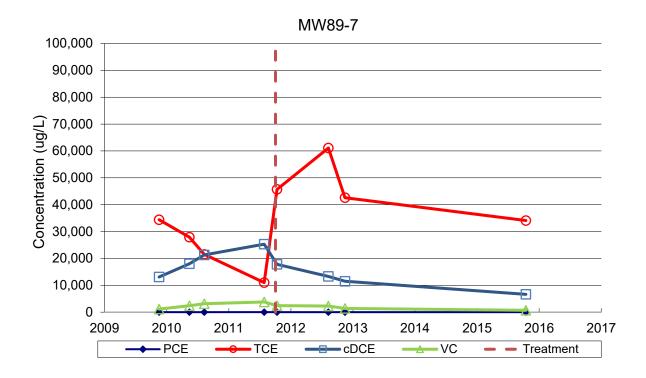


Figure 5-33. VOC Trends for MW89-7, Former Charleston Air Force Base Zone 4.

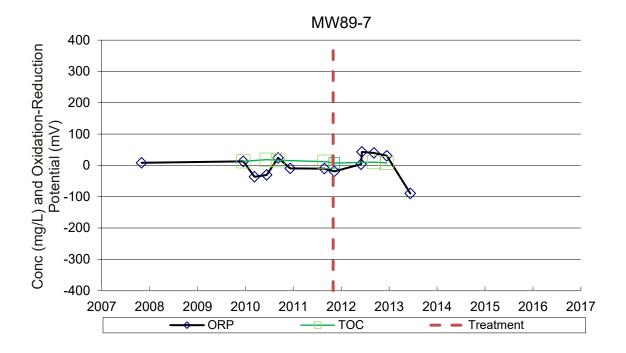


Figure 5-34. ORP and TOC for MW89-7, Former Charleston Air Force Base Zone 4.

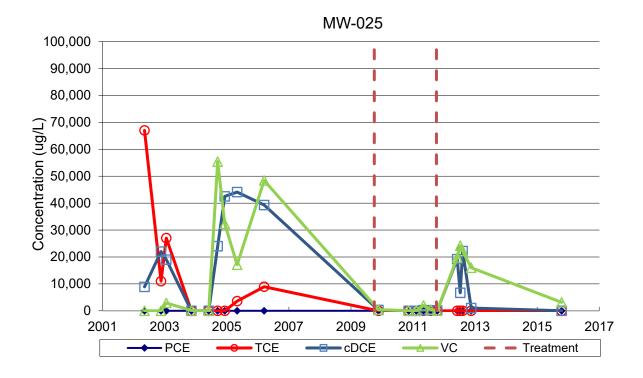


Figure 5-35. VOC Trends for MW-025, Former Charleston Air Force Base Zone 4.

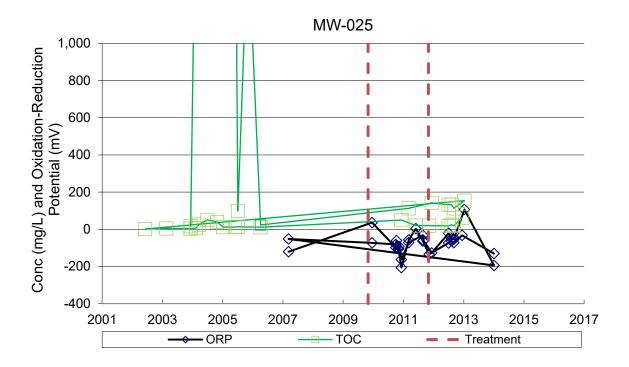


Figure 5-36. ORP and TOC for MW-025, Former Charleston Air Force Base Zone 4.

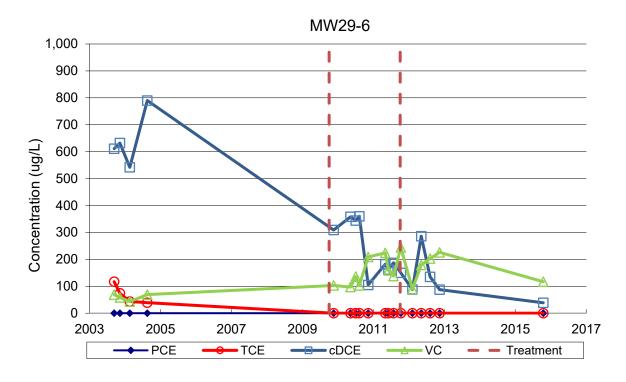


Figure 5-37. VOC Trends for MW29-6, Former Charleston Air Force Base Zone 4.

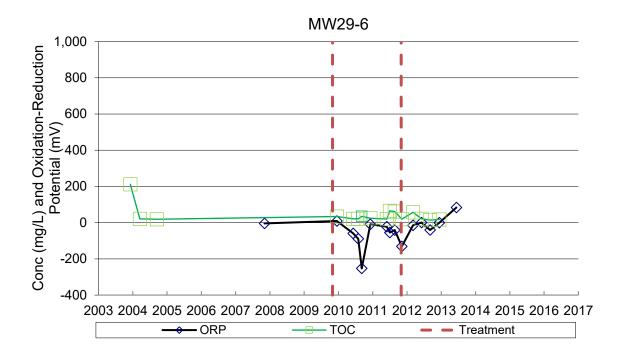


Figure 5-38. ORP and TOC for MW29-6, Former Charleston Air Force Base Zone 4.

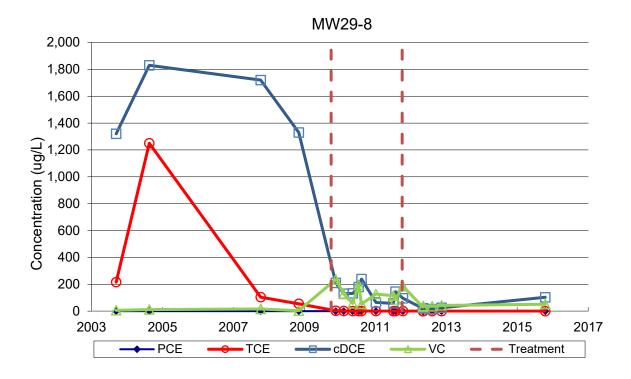
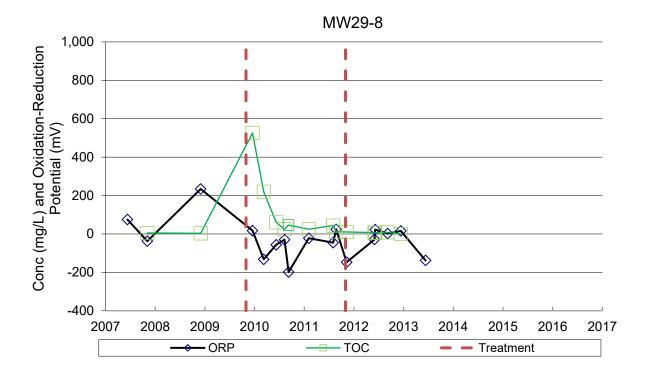
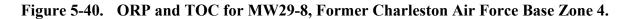


Figure 5-39. VOC Trends for MW29-8, Former Charleston Air Force Base Zone 4.





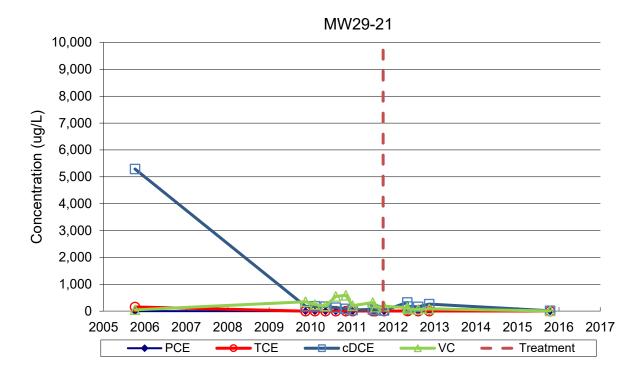


Figure 5-41. VOC Trends for MW29-21, Former Charleston Air Force Base Zone 4.

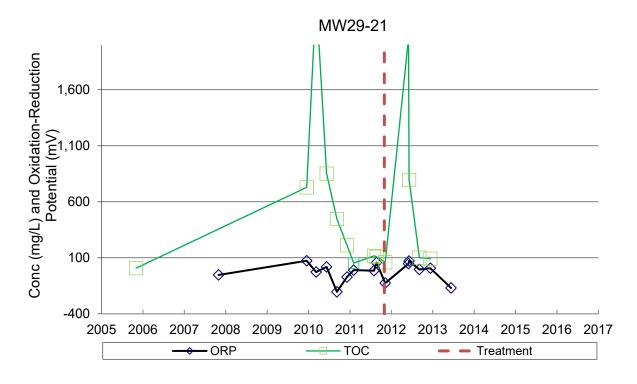


Figure 5-42. ORP and TOC for MW29-21, Former Charleston Air Force Base Zone 4.

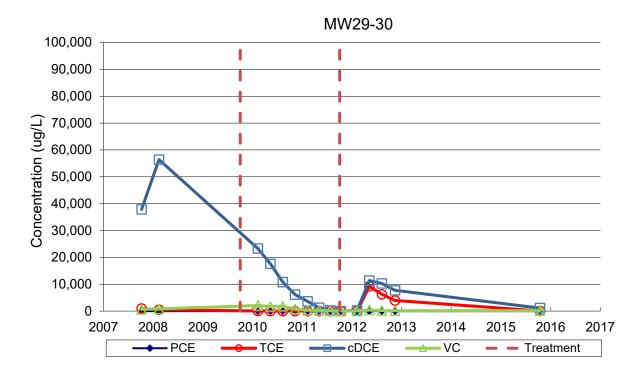
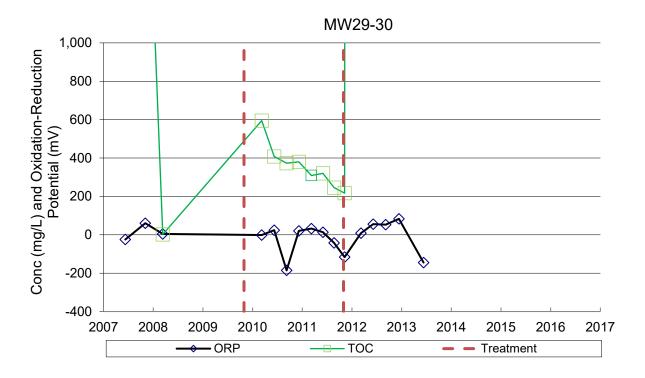
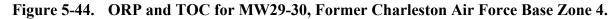


Figure 5-43. VOC Trends for MW29-30, Former Charleston Air Force Base Zone 4.





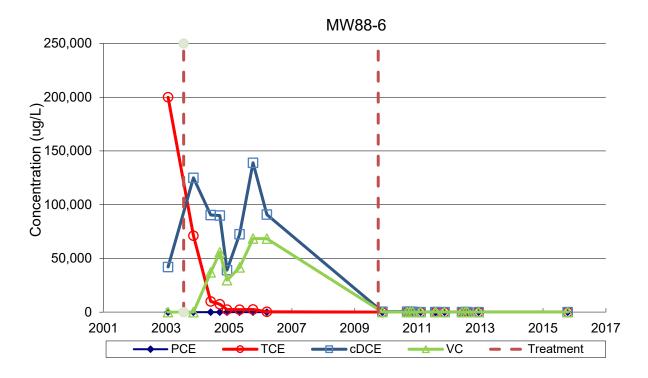


Figure 5-45. VOC Trends for MW88-6, Former Charleston Air Force Base Zone 4.

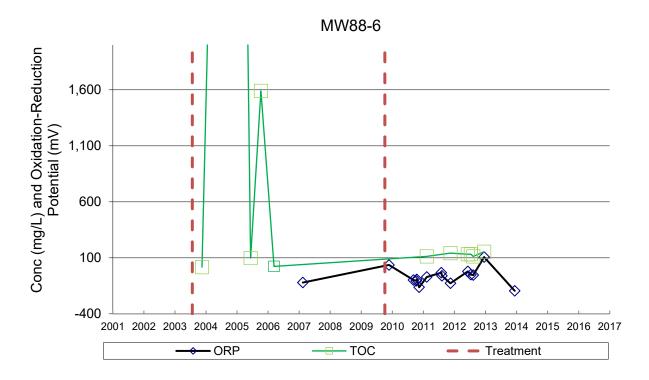


Figure 5-46. ORP and TOC for MW88-6, Former Charleston Air Force Base Zone 4.

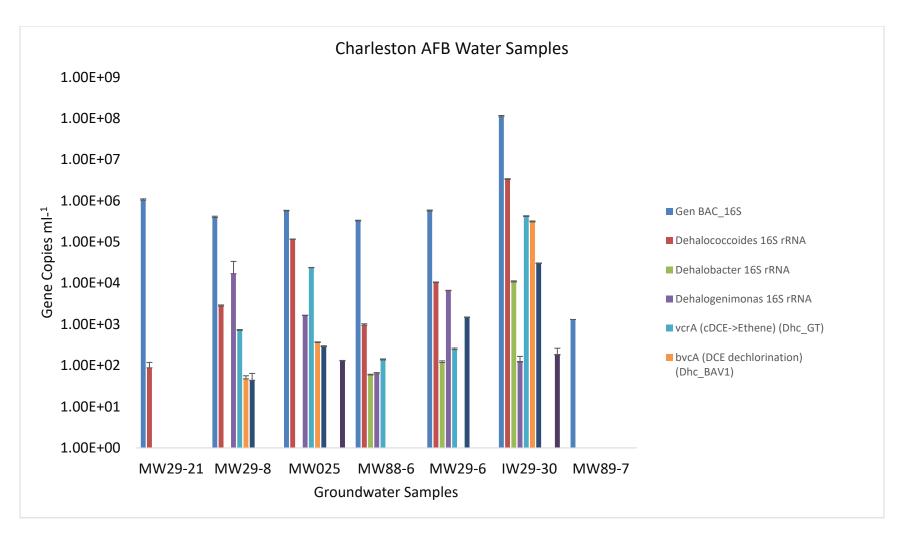


Figure 5-47. Organisms and Genes Associated with Reductive Dechlorination Detected in Wells at Charleston Air Force Base.

#### 5.3.3.3.8. CSIA

Values of  $\delta^{13}$ C were obtained for TCE in MW29-30 and MW89-7, for cis-DCE in MW29-21, MW29-30, MW29-6, MW29-8, and MW89-7, and for VC in MW29-21, MW29-30, MW29-6, MW29-8, and MW88-6 (Table 5-16). The  $\delta^{13}$ C values are presented with VOC concentration data and well location (i.e., Plume 1 or Plume 2), with the most upgradient wells listed first in Table 5-17. There are not enough  $\delta^{13}$ C data to prepare plots as presented for the Raritan data. However, the data do provide some insight concerning ongoing biodegradation at the Charleston site. For Plume 1, the most upgradient well MW89-7 (a source zone well),had TCE at 34,100 µg/L and a  $\delta^{13}$ C value of -25.1‰, which is in the typical range for parent TCE (-34‰ to - 23‰). cis-DCE, which was detected at 6,630  $\mu$ g/L, had a reported  $\delta^{13}$ C value of -25.7‰, very close to TCE. No value was reported by the analytical laboratory for VC, presumably because the concentration was too low to analyze VC based on the concentration of TCE and cis-DCE in the same sample. At first view, the CSIA data suggest that biodegradation of TCE and cis-DCE is not occurring in this zone. This is supported by the general absence of dechlorinating organisms/genes in this sample (Figure 5-47). However, the concentration of TCE indicates that this zone may contain residual DNAPL. In this case,  $\delta^{13}$ C from TCE may not be a good measure of ongoing transformation because partitioning of parent TCE from the DNAPL is likely to reduce and isotopic enrichment in parent TCE caused by degradative processes. Because  $\delta^{13}$ C in cis-DCE did not exceed the typical parent value for TCE, whether or not cis-DCE degradation is ongoing also cannot be determined in this zone (see Section 5.3.2.3.8).

Moving downgradient in Plume 1, no  $\delta^{13}$ C data were available for TCE or cis-DCE in MW025, due to low concentrations. The laboratory also did not provide data for VC despite a reported concentration of 3,130 µg/L. Inquiries concerning the analytical issue with this sample were not answered. However, it is interesting to note that only VC was detected in this well, which is downgradient of the source area with high remaining TCE and cis-DCE. Thus, ongoing VOC biodegradation is indicated by the data and supported by the previous qPCR data. The final well in this general plume transect, MW88-6, had VC detected at 68.6 µg/L, and with a  $\delta^{13}$ C value of +1.20 ‰, indicative of significant ongoing biodegradation. This result is again consistent with the qPCR data, which showed the presence of *Dehalococcoides* and the *vcrA* gene.

In Plume 2, TCE was only detected in the most upgradient well (IW29-30) at 63 µg/L and with a  $\delta^{13}$ C value of -21‰. In this same well cis-DCE was detected at 1,150 µg/L with a  $\delta^{13}$ C of -22.0‰ and VC was observed at 179 µg/L with a  $\delta^{13}$ C value of -25.6‰. The daughter products are indicative of TCE dehalogenation consistent with the high concentrations of dechlorinating organisms and reductive dehalogenase genes in this well. The  $\delta^{13}$ C of the three VOCs (parent and two daughter products), however, does not conclusively support ongoing biodegradation in this region. Moving downgradient, however, a general trend is apparent for *cis*-DCE degradation based on  $\delta^{13}$ C, as isotopic enrichment is apparent along the flow path from -22.0‰ for IW29-30 to -7.8‰ for MW29-8, -13.2‰ for MW29-21, and -3.78 ‰ for MW29-6. The groundwater in this plume passes through two injection zones along this path (Grid #3 for the first 3 wells and Grid #4 for the last well), but cis-DCE degradation is clearly indicated by the isotopic enrichment (i.e., from -22.0‰ to -3.78‰. A large initial increase in the  $\delta^{13}$ C value of VC is also observed from MW29-21 at -25.6‰ to MW29-21 at -11.9‰, with a corresponding concentration decrease of about 4-fold. The  $\delta^{13}$ C value then decreases further to -9.29‰ for MW29-21. The a  $\delta^{13}$ C in the final well in Grid #4 then increases slightly to -13.3‰ with a corresponding increase in concentration.

The data indicate ongoing VC biodegradation along the flow path through Grid#3, but perhaps a lack of further degradation of VC at the final well. Further sampling would be required to get a better long-term picture in this region.

#### 5.3.3.3.9. Mass Flux

The contaminant flux profiles, as measured by the PFM deployments, are shown in Figure 5-48 for the eight wells sampled at the Former Charleston Air Force Base, Zone 4. Because of well sizes and other variables, PFMs were, in some cases, deployed in wells not sampled for other parameters, but in the same vicinity as those companion wells. Calculated mass fluxes for each VOC in each well are provided in Table 5-18. As shown in Figures 5-29 and 5-30 and previously described, the site is represented by four "Grids" Two PFMs were deployed in Plume 1, Grid #1, located to the Northwest of Building 543, in wells MW89-2 (Figure 5-48a) and MW89-7 (Figure 5-48b). Only cis-DCE was detected in MW89-2 in the shallow portion of the aquifer. TCE, cis-DCE, VC and ethene were detected in MW89-7, consistent with the analytical from this well, and as noted previously in the CSIA section, it is likely that DNAPL remained in this well. Three PFMs were deployed to the Southwest of Building 542 in Grid #2 and Transect #3 in wells MW88-6 (Figure 5-48c), MW025 (Figure 5-48d) and Z4-17 (Figure 5-48e). Ethene, VC, cis-DCE, and TCE were all detected in MW88-6, but cis-DCE was by far the highest with a maximum flux of  $20 \text{ mg/m}^2/\text{day}$ . These data are inconsistent with the analytical data from this well in which only VC was detected above the PQL. One potential reason for the discrepancy is diffusion of cis-DCE into the aquifer from a low permeability zone during the relatively long incubation period of the PFM. In that well samples generally pull only form the most conductive zones, cis-DCE may not have been detected during aqueous phase groundwater sampling. Cis-DCE and VC were detected in MW025 (Transect #3), but their fluxes were relatively small in comparison to the ethene flux where the average and maximum mass flux detected was  $101 \text{ mg/m}^2/\text{day}$  and  $211 \text{ mg/m}^2/\text{day}$ . The high flux of ethene (as well as the high detected concentration in this well; 5,840 µg/L) suggests high biological activity near/upgradient of this well. VC and ethene were both detected in MWZ4-17 with average mass fluxes of  $\sim 1300 \text{ mg/m}^2/\text{day}$  and  $\sim 470 \text{ mg/m}^2/\text{day}$ , respectively. This well was not otherwise sampled. Two PFMs were deployed in Plume 2, Grid #3, located to the Northwest of Building 532. In MW29-8 (Figure 5-48f), no TCE was detected, but cis-DCE, VC, and ethene were detected. Ethene was detected in small quantities compared to VC and cis-DCE, which were detected (consistent with the analytical data from this well) with maximum mass fluxes of  $\sim 28 \text{ mg/m}^2/\text{day}$  and  $\sim 15 \text{ mg/m}^2/\text{day}$ . Overall, contaminant mass fluxes at this well were much lower than for those in the Plume #1. Only cis-DCE was detected in MW29-21 (Figure 5-48c) with a relatively low average mass flux of  $\sim 2 \text{ mg/m}^2/\text{day}$ . A final PFM was deployed in Plume #2, Grid 4 in MW29-6. Ethene, VC, and cis-DCE were detected consistent with the analytical data from this location. Ethene flux was relatively small compared to cis-DCE and VC, which had average mass fluxes of  $\sim 32 \text{ mg/m}^2/\text{day}$  and  $21 \text{ mg/m}^2/\text{day}$ , respectively.

The flux-averaged Equiv. TCE concentration (TCE+cis-DCE+VC), estimates from the PFM results, and groundwater monitoring wells were compared (**Table 5-19**). Groundwater concentration data suggest no chlorinated ethenes were present in MW 89-2 during the time of sampling, but PFM data suggested that cis-DCE was present, but was below MCL. MW 89-7 groundwater and PFM data showed the TCE, cis-DCE, and VC were present above MCLs and free phase DNAPL is likely present since equivalent TCE concentrations were reported to be ~40 mg/L, The proportion of contaminant present in groundwater is difficult to discern since TCE was

detected at the highest concentrations in groundwater samples, whereas cis-DCE was detected at the highest concentration in PFM samples. Regardless, both measurements indicated that complete reductive dechlorination is occurring since ethene was detected in both measurement types. Downgradient of the source zone, VC and ethene were detected in MW Z4-17 with flux-averaged concentrations of ~23,300  $\mu$ g/L and ~8,200  $\mu$ g/L, respectively, suggesting that complete dechlorination is occurring between wells MW 89-7 and Z4-17. Groundwater samples were not taken for this well, the PFM data from Z4-17 suggest that free phase DNAPL remains in the source zone, which is similar to what was indicated by MW 89-2 data. Immediately downgradient of well Z4-17, MW-025 groundwater and PFM data concurred that VC is above its MCLs but that no TCE is present. For cis-DCE, groundwater data indicated that it was below the MCL whereas PFM data indicated the opposite. Ethene was also detected in well MW-025, indicating that complete dechlorination is occurring, but PFM data showed that ethene concentrations were 101,000  $\mu$ g/L, whereas groundwater data showed a concentration of 5,840  $\mu$ g/L. There high values by both methods clearly indicate ongoing degradation of chlorinated ethenes.

For Plume 2, groundwater and PFM data for MW 29-8 showed that cis-DCE and VC were present above MCLs, and ethene was detected, suggesting that complete dechlorination is ongoing. VC and cis-DCE were detected above MCLs at MW 29-21and ethane (not ethene) was present in groundwater samples. For this same well, the PFM detected cis-DCE (below the MCL) but no VC or ethane were measurable. Groundwater and PFM data for Well MW 29-6, located at the toe of the plume, both detected ethene, VC and cis-DCE. However, groundwater data showed that cis-DCE was below the MCL, whereas PFM data suggested it was above the MCL.

In some instances, groundwater and PFM data compared very well and in others they did not. Past studies have shown that flux-averaged values are better estimates than groundwater data because flux-average concentration is independent of divergence and is both a temporally and spatially average concentration as opposed to "instantaneous" measurements of groundwater data (Basu et al., 2006; Brooks et al., 2008). It is informative to have both measures to evaluate at a given site.

#### 5.3.3.3.10. Site Summary

Results from Charleston indicate that biogeochemical conditions remained generally favorable for reductive dechlorination of chlorinated ethenes. The presence of daughter products as well as ethane and/or ethene in most wells based on both sampling and PFM data indicates ongoing biodegradation, although significant mass was still likely present in the Plume 1 source area (e.g., near MW 89-7). The microbial community analysis showed the presence of numerous different dehalogenation-associated organisms/genes in water samples from across the site with *Dehalococcoides* concentrations ranging from  $9 \ge 10^1$  cells/mL to  $3.3 \ge 10^6$  cells/mL. The overall CSIA data did not support degradation in the Plume 1 source area well (MW 89-7), but this may reflect the presence of DNAPL, and the continued resupply of unenriched TCE as biodegradation occurs. All other measures indicated ongoing degradation in this area. Moreover, CSIA clearly indicated ongoing biodegradation downgradient of this source. The same scenario was found in Plume 2: ambiguous data in the most upgradient well (MW29-30), with clear isotopic enrichment in daughter products (TCE is largely gone), downgradient of this well, indicative of ongoing biodegradation. Groundwater and PFM data showed that free-phase DNAPL was still present at the site in and or around well MW 89-7. Overall, the results suggest that biological degradation processes have persisted  $\sim$  4 years after cessation of active bioremediation.

Well	δ <sup>13</sup> C TCE (‰)	TCE (µg/L)	δ <sup>13</sup> C DCE (‰)	DCE (µg/L)	δ <sup>13</sup> C VC (‰)	VC (µg/L)	Grid/depth
MW89-7	-25.1	34,100	-25.7	6,630	NA	697	Plume 1, Grid 1/ Shallow
MW025	NA	< 5	NA	6.75	NA	3,130	Plume 1, Transect 3/ Deep
MW88-6	NA	< 5	NA	< 5	+ 1.20	68.6	Plume 1, Grid 2/ Deep
IW29-30	-21.0	63	-22.0	1,150	-25.6	179	Plume 2, Grid 3/ Intermediate
MW29-8	NA	< 5	-7.84	103	-11.9	51.5	Plume 2, Grid 3/ Shallow
MW29-21	NA	< 5	-13.2	13.7	-9.3	8.1	Plume 2, Grid 3/ Shallow
MW29-6	NA	< 5	-3.78	38.7	-13.3	117	Plume 2, Grid 4/ Shallow

Table 5-17.Groundwater Sampling Locations and Rationale, Former Charleston Air<br/>Force Base Zone 4.

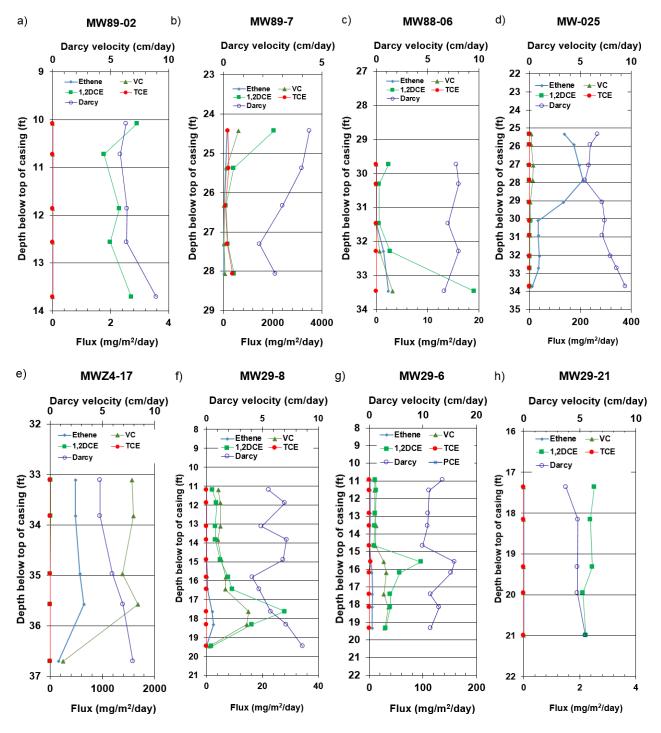


Figure 5-48. Mass Flux Profiles Measured in Select Wells using PFMs.

The solid blue diamonds represent ethene, the solid dark green triangles represent VC, the solid light green squares represent DCE, the solid red circles represent TCE, and the open purple circles represent the Darcy Velocity. Panels a-h show flux values sampled on 11/04/2016. Note the changes in scale on both axes to accommodate the data.

Well_ID	Average Darcy Velocity (cm/day)	Average Ethene flux (mg/m²/day)	Average VC flux (mg/m²/day)	Average 1,2DCE flux (mg/m²/day)	Average TCE flux (mg/m²/day)
MW 89-02	6.78	0.00	0.00	2.34	0.00
MW 89-7	3.16	61.20	170.67	639.36	204.96
MW 88-06	7.51	0.74	0.75	5.05	0.00
MW 025	7.18	101.40	5.51	1.31	0.00
Z4-17	6.09	471	1296	5.01	0.00
MW 29-6	12.37	3.09	20.51	31.59	0.26
MW 29-8	6.18	0.53	6.77	8.03	0.00
MW 29-21	4.73	0.00	0.00	2.34	0.00

 Table 5-18.
 Average Mass Discharge for Each Well.

Table 5-19.VOC Concentration and GW and PFM Concentration Comparisons for<br/>Wells in Zone 4, Charleston Air Force Base.

Chemical	Туре	MW 89-02	MW 89-7	Z4-17	MW 025	MW 88-06	IW 29- 30	MW 29-8	MW 29-21	MW 29-6
Equiv. TCE	GW	na	44,640	na	6,968	145	376	249	205	299
	PFM	47	40,037	49,117	13,364	122	na	435	68	665
	% diff.	na	11	na	63	17	na	54	100	76
TCE	GW	na	34,100	na	<dl< td=""><td><dl< td=""><td>63</td><td><dl< td=""><td><dl< td=""><td><dl< td=""></dl<></td></dl<></td></dl<></td></dl<></td></dl<>	<dl< td=""><td>63</td><td><dl< td=""><td><dl< td=""><td><dl< td=""></dl<></td></dl<></td></dl<></td></dl<>	63	<dl< td=""><td><dl< td=""><td><dl< td=""></dl<></td></dl<></td></dl<>	<dl< td=""><td><dl< td=""></dl<></td></dl<>	<dl< td=""></dl<>
	PFM	<dl< td=""><td>7078</td><td><dl< td=""><td><dl< td=""><td><dl< td=""><td>na</td><td><dl< td=""><td><dl< td=""><td>2</td></dl<></td></dl<></td></dl<></td></dl<></td></dl<></td></dl<>	7078	<dl< td=""><td><dl< td=""><td><dl< td=""><td>na</td><td><dl< td=""><td><dl< td=""><td>2</td></dl<></td></dl<></td></dl<></td></dl<></td></dl<>	<dl< td=""><td><dl< td=""><td>na</td><td><dl< td=""><td><dl< td=""><td>2</td></dl<></td></dl<></td></dl<></td></dl<>	<dl< td=""><td>na</td><td><dl< td=""><td><dl< td=""><td>2</td></dl<></td></dl<></td></dl<>	na	<dl< td=""><td><dl< td=""><td>2</td></dl<></td></dl<>	<dl< td=""><td>2</td></dl<>	2
	% diff.	na	131	na	na	na	na	na	na	na
cis-DCE	GW	na	6630	na	6.8	<dl< td=""><td>1150</td><td>103</td><td>137</td><td>39</td></dl<>	1150	103	137	39
	PFM	34.8	17498	88.5	1,307	73.2	na	139.5	50.4	237
	% diff.	na	90	na	198	na	na	na	na	n
VC	GW	na	697	na	3310	68.8	179	51.5	8.1	117
	PFM	0.0	4402	23,305	5,515	10.4	na	117	0.0	163
	% diff.	na	145	na	50	148	na	78	200	33
Ethene	GW	na	210	na	5840	82.9	55.2	6.5	<dl< td=""><td>17</td></dl<>	17
	PFM	0.0	1665	8,245	101,396	7.51	na	8.7	0.0	24.7
	% diff.	na	155	na	178	167	na	29	na	37
Ethane	GW	na	131	na	2,180	332	23	23	4.09	4.3
	PFM	na	na	na	na	na	na	na	na	na
	% diff.	na	na	na	na	na	na	na	na	na

\*PFMs only deployed in this well. VOCs, CSIA, and geochemical analysis were not performed on wells due to scheduling conflicts.

Note: na = not available

#### 5.3.4 Seal Beach Naval Weapons Station Site 70

#### 5.3.4.1 Site Description, Seal Beach Naval Weapons Station Site 70

Seal Beach Site 70 is also known as the Research, Testing, and Evaluation Area. TCE was used during various research and development activities that occurred at the site between 1962 and 1985. A site location map is provided in **Figure 5-49**.

Groundwater contamination is present in the source area where dense non-aqueous phase liquid (DNAPL) is suspected to occur based on the groundwater concentrations; however, DNAPL has not been observed in the various subsurface investigations. A dissolved-phase plume currently extends at least 4,000 ft downgradient to a depth of approximately 160 ft below ground surface (bgs). TCE was the primary contaminant in the plume before bioremediation (**Figure 5-50**). *cis*-DCE was detected at low concentrations throughout the plume, indicating a low level of natural biological attenuation before treatment.

The uppermost geologic material at the site contains recent alluvial sediments to a depth of approximately 60 ft bgs. This zone is designated the Upper Fines Unit and contains interbedded silts, clays, and silty clays. Heterogeneity is high in this unit, and measured hydraulic conductivities vary by an order of magnitude or more over several ft. The hydraulic conductivity value is low in the Upper Fines Unit. Water level monitoring indicates that water levels decline approximately 2.2 ft over the 45-foot saturated thickness of the Upper Fines Unit, yielding a vertical hydraulic gradient of approximately 0.05 ft per foot. The horizontal hydraulic gradient in most of the source area is generally to the south.

The Upper Fines Unit is underlain by what has been designated the First Sand, which extends from approximately 105 to 135 ft bgs. This unit is predominantly composed of well-sorted, fine-grained sand and silty sand. Gravelly coarse sand is sometimes observed at 80 to 95 ft bgs. The First Sand has a relatively high hydraulic conductivity of 68 ft per day.

Groundwater is generally anaerobic with a neutral pH. Groundwater is brackish with elevated chloride concentrations. High sulfate concentrations (>1,000 mg/L) are present throughout much of the site.

An ESTCP demonstration was implemented in the source area from 2008 to 2010 (ESTCP Project ER-200513). The primary objective of the demonstration was to compare the ability to distribute *Dehalococcoides* via passive versus active approaches. Two test cells were established in the source area, designated the Passive Cell and the Active Cell, as shown in **Figure 5-51**. Sodium lactate and the SDC-9 bioaugmentation culture were introduced into both cells. The amendments were allowed to migrate with the natural groundwater gradient in the Passive Cell. In the Active Cell, the amendments were distributed by recirculation established between two injection wells and two extraction wells. The treatment time for the ESTCP demonstration extended over a 10-month period, and significant dechlorination occurred in both test cells as a result. The results of the demonstration indicated that bacterial distribution was similar in both the Active and Passive Cells in terms of travel time and area influenced.

A full-scale bioremediation program was implemented shortly after the ESTCP demonstration, with a combination of a grid of injection wells in the source area and a series of biobarriers perpendicular to the dissolved-phase plume further downgradient. Figure 5-51 provides the layout of the site with aerial imagery to show site features, and Figure 5-52 shows the site plan without the aerial imagery. The Source Area Treatment Grid is an irregular shape approximately 140 ft wide and 220 ft long, containing 57 injections wells. The Source Area Biobarrier is located approximately 200 ft downgradient and contains 14 injection wells aligned roughly perpendicular to groundwater flow.

The biobarriers include permanent injection wells at a spacing of 25 ft. EVO and the KB-1 bioaugmentation culture were the amendments used in the full-scale design. The full-scale system was implemented in areas of the plume with TCE exceeding 250  $\mu$ g/L. The initial injections in the biobarriers occurred in 2009 and in the source area grid in 2010. The second round of injections occurred in 2013 in the source area and all but one of the biobarriers.

The results of the bioremediation activities have been mixed, with substantial dechlorination occurring in some areas and partial dechlorination in other areas. A significant accumulation of DCE and VC was observed at many monitoring locations.



Figure 5-49. Site Location Map, Seal Beach Naval Weapons Station Site 70.

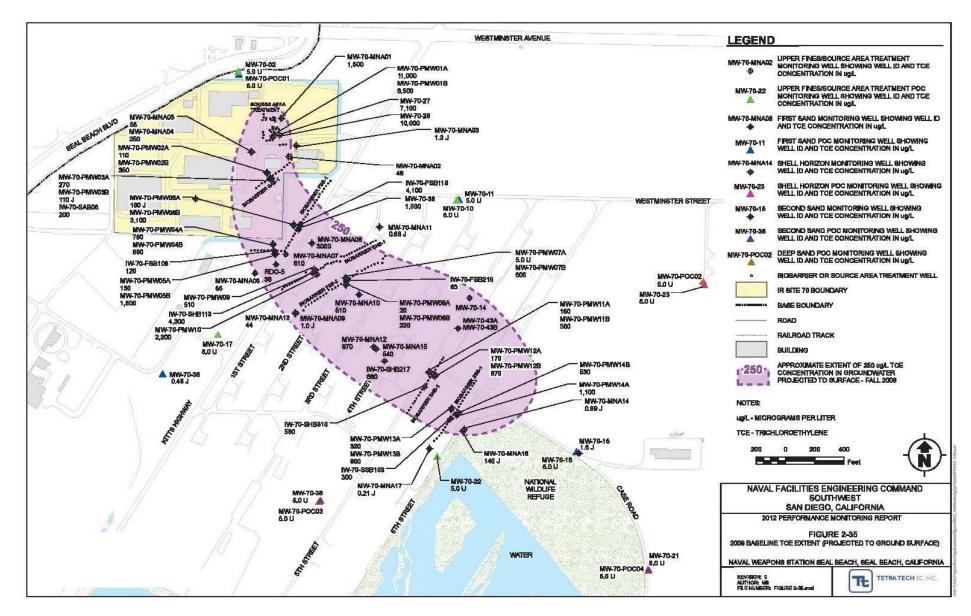


Figure 5-50. Extent of Plume Before Bioremediation, Seal Beach Naval Weapons Station Site 70.



Figure 5-51. Site Plan with Aerial Imagery, Seal Beach Naval Weapons Station Site 70.

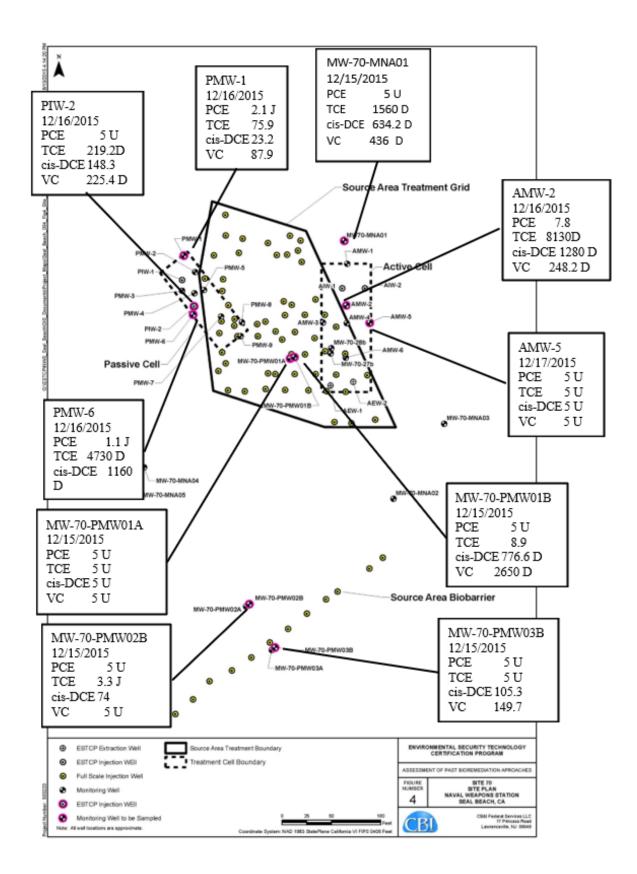


Figure 5-52. Site Plan without Aerial Imagery, Seal Beach Naval Weapons Station Site 70.

#### 5.3.4.2 Sampling Rationale, Seal Beach Naval Weapons Station Site 70

The objective of the groundwater sampling was to assess the long-term impacts of the previously performed bioaugmentation treatment on groundwater quality and biogeochemistry. Monitoring locations upgradient, within, and downgradient of the Source Area Treatment Grid and Source Area Biobarrier were evaluated as described below. Monitoring wells that were sampled are shown in **Figures 5-52**. **Table 5-20** summarizes the monitoring locations and rationale for the sampling program.

Seal Beach Site 70 Monitoring Well	Sampling Rationale
MW70-MNA01	Within the source area but upgradient from both the ESTCP Active Cell and full-scale treatment zones. This well served as an untreated source area location.
AMW-2	Within the ESTCP Active Cell treatment zone but not within the full-scale treatment zone. This well represented the long-term effects of the ESTCP Active Cell treatment.
AMW-5	Similar to AMW-2 described above, this well is within the ESTCP Active Cell treatment zone, and represented the long-term effects of the ESTCP Active Cell treatment.
PMW-1	On the upgradient edge of the ESTCP Passive Cell treatment zone. This well represented the upgradient conditions associated with the ESTCP Passive Cell treatment.
PIW-2	Within the ESTCP Passive Cell treatment zone but not within the full-scale treatment zone. This well represented the long-term effects of the ESTCP Passive Cell treatment.
PMW-6	This well is also within the ESTCP Passive Cell treatment zone and represented the long-term effects of the ESTCP Passive Cell treatment
MW-70-PMW01A	Within the Source Area Treatment Grid but not within either of the ESTCP treatment zones. This well represented the long-term effects of full-scale treatment. Substantial contaminant reduction has been observed in this well.
MW-70-PMW01B	Located adjacent to MW-70-PMW01A within the Source Area Treatment Grid but screened at a deeper depth horizon. Substantial contaminant reduction has not been observed in this well.
MW-70-PMW02B	Downgradient of the ESTCP and full-scale treatment zones. This well was evaluated for downgradient effects associated with the Source Area treatment.
MW-70-PMW03B	Downgradient of the Source Area Biobarrier. This well was evaluated for downgradient effects associated with the Source Area Biobarrier treatment.

## Table 5-20.Groundwater Sampling Locations and Rationale, Seal Beach Naval Weapons<br/>Station Site 70.

Monitoring well MW70-MNA01 is located within the source area but upgradient of both the ESTCP Active Cell and the full-scale Source Area Treatment Grid. The working assumption was that this well had not received any treatment due to its upgradient location. This well was evaluated as a source area well that did not receive treatment.

Monitoring wells AMW-2 and AMW-5 are located within the ESTCP Active Cell treatment zone but not within the Source Area Treatment Grid. The working assumption for these wells was that they only received treatment from the ESTCP Active Cell. They were evaluated for the long-term effects of the ESTCP Active Cell treatment only. During the ESTCP demonstration, significant degradation of TCE was observed in each of these wells. A significant amount of historical VOC, geochemical and microbial data are available for these wells.

Monitoring well PMW-1 is located on the upgradient edge of the ESTCP Passive Cell treatment zone and will represent upgradient conditions. Monitoring wells PIW-2 and PMW-6 are located within the ESTCP Passive Cell treatment zone but not within the Source Area Treatment Grid. The working assumption for these wells was that they only received treatment from the ESTCP Passive Cell. They were evaluated for the long-term effects of the ESTCP Passive Cell treatment that was completed in 2009.

Monitoring wells MW-70-PMW01A and MW-70-PMW01B are located at two different depths within the full-scale Source Area Treatment Grid. The initial treatment in this area included the injection of EVO and KB-1 in 2010. The second injection of EVO was conducted in 2013. Post-treatment monitoring results indicate a substantial reduction in contaminants in well MW-70PMW01A, but not in well MW-70-PMW01B. These wells were evaluated to determine what factor(s) may influence the difference in the effectiveness of the treatment between the two wells.

Monitoring well MW-70-PMW02B is located approximately 180 ft downgradient of the Source Area Treatment Grid and approximately 45 ft upgradient of the Source Area Biobarrier. This well is screened in the First Sand depth horizon and downgradient of the source area, which is located within the Upper Fines depth zone. This well was evaluated for effects downgradient of the Source Area.

Monitoring well MW-70-PMW03B is located approximately 10 ft downgradient of the Source Area Biobarrier and was used to evaluate the effects of the barrier treatment approach. This well is also screened in the First Sand depth horizon.

The monitoring well locations are shown in **Figures 5-52**. **Table 5-21** provides a summary of well construction details.

Monitoring Well	Total Depth (bgs)	Top of Screen (bgs)	Bottom of Screen (bgs)	Diameter (inches)
MW70-MNA01	50	40	50	4
AMW-2	35	15	35	4
AMW-5 Zone 1	34	33	34	1.7
PMW-1	35	15	35	4
PIW-2	35	15	35	4
PMW-6	35	15	35	4
MW70—PMW01A	35	25	35	4
MW70—PMW01B	55	45	55	4
MW70—PMW02B	100	90	100	4
MW70—PMW03B	100	90	100	4

# Table 5-21.Monitoring Well Construction Details, Seal Beach Naval Weapons Station<br/>Site 70.

#### 5.3.4.3 Sampling Results, Seal Beach Naval Weapons Station Site 70

#### 5.3.4.3.1. VOCs

Table 5-22 provides a summary of sampling results, and Figures 5-53 to 5-70 provide historical and current VOC data and corresponding ORP and sulfate data for the different wells sampled. The primary parent contaminant at this site was TCE, although traces of PCE were detected in a few wells. For the upgradient, untreated well, MW70-MNA01, the concentration of TCE at the time of sampling was 1560 µg/L, cis-DCE was 634 µg/L, and VC was detected at 436 µg/L. Historical data at this well showed that TCE declined substantially from the initial sampling in 2011 (2,800 µg/L) while cis-DCE and VC increased over this same timeframe (Figure 5-53). Thus, the data are consistent with some natural attenuation of TCE occurring at this site. The two wells that previously were treated via active addition of lactate and SDC-9 (AMW-2 and AMW-5) several years before the current sampling had very different VOC profiles. AMW-2, which had shown substantial treatment during the ESTCP study in 2008, with TCE declining from ~ 10,000  $\mu$ g/L to 210  $\mu$ g/L (Figure 5-54), showed a substantial rebound in 12/2015 when the current sample collection occurred. TCE was detected at 8,130 µg/L. Interestingly, cis-DCE showed only a modest increase over this same timeframe while VC declined substantially, from 4,200 µg/L in 2008 at the final sampling point of the demonstration to 248 µg/L in 2015. For AMW-5, TCE, cis-DCE, and VC showed an initial steep increase after treatment, likely due to the active recirculating of groundwater combined with ongoing biodegradation, with TCE reaching 6,400  $\mu$ g/L from a starting concentration of 710  $\mu$ g/L before treatment (Figure 5-56). During the final sampling event for the original ESTCP demonstration in 10/2009, this well had TCE, cis-DCE, and VC at 940 µg/L, 1400 µg/L, and 5400 µg/L, respectively. During sampling for this project in 12/2015, all three VOCs were  $< 5 \mu g/L$ . Interestingly, 1,2-dichloropropane, bromochloromethane, and bromomethane were all detected in this well at  $> 1,000 \mu g/L$ . These constituents were also observed in MW70-PMW01B at far lower concentrations, but not elsewhere.

Wells PMW-1, PIW-2, and PMW-6 were located in the ESTCP passive treatment cell, with PMW-1 being at the upgradient edge of the cell and the other two wells within the treated region. TCE in PMW-1 was ~ 1,100  $\mu$ g/L at the initiation of the ESTCP demonstration in 2008, increased somewhat during the treatment period, and then remained near 1,500 µg/L for the duration of the study (final sampling event 10/2009) (Figure 5-58). VC and cis-DCE were generally  $< 100 \mu g/L$ during this same interval. In 12/2015, TCE in this well was 76 µg/L, while cis-DCE and VC were 23 µg/L and 88 µg/L, respectively. Thus. TCE showed a substantial decline after the end of the ESTCP study. It is not clear whether this represents a decline in the upgradient TCE concentrations or a longer-term impact of the ESTCP treatment. The downgradient well PMW-6 had TCE detected at 11,000 µg/L prior to substrate and culture injection and declined to a low of 470 µg/L during the field trial (Figure 5-60). Cis-DCE and VC reached 640 µg/L and 370 µg/L, respectively. During the 12/15 sampling event, the TCE in this well had partially rebounded to 4,730  $\mu$ g/L, while cis-DCE was 1,160 µg/L, and VC was 584 µg/L, respectively. The final well sampled in this plot, PIW-2, was an injection well in which TCE declined from 20,000 µg/L to < 250 µg/L during the demonstration, with minimal formation of cis-DCE or VC (< 50  $\mu$ g/L each) (Figure 5-62). During the 12/2015 sampling event, TCE remained low compared to historical concentrations, being detected at 219 µg/L, and cis-DCE and VC also remained relatively low at 225 µg/L and 148 µg/L, respectively. Thus, overall, well PMW-6 showed a different long-term outcome (i.e., TCE rebound) than either PMW-1 or PIW-2.

Wells MW-70-PMW01A and MW-70-PMW01B were within the full-scale Source Area Treatment Grid that received EVO and KB-1 in 2010 and a second injection of EVO was conducted in 2013. The wells were adjacent, with the former being screened at a shallower interval. Well MW70-PMW01A responded very well to treatment, with TCE declining from > 11,000 µg/L in 2011 to < 5 µg/L during sampling 12/2015 (**Figure 5-63**). Cis-DCE and VC were also below detection (< 5 µg/L). Similarly, TCE in MW-70-PMW01B declined appreciably after treatment, falling from 8,500 µg/L in 2011 to 8.9 µg/L during the 12/2015 sampling event (**Figure 5-65**). In contrast to MW-70-PMW01A, however, significant concentration of cis-DCE (777 µg/L) and VC (2,560 µg/L) were detected in this well during the 12/2015 sampling. Thus, the wells showed a somewhat different response to treatment with respect to daughter products of reductive dehalogenation.

The final two wells, MW-70-PMW02B and MW-70-PMW03B, are located approximately 45 ft upgradient and 10 ft downgradient of the Source Area Biobarrier, respectively. Both wells are appreciably downgradient of the ESTCP treatment plots. MW-70-PMW02B had TCE detected at 410  $\mu$ g/L and cis-DCE at 770  $\mu$ g/L in 2011, presumably the latter resulting from upgradient treatment (**Figure 5-67**). Each of these compounds showed a steady decline. By 2015, TCE was observed at < 5  $\mu$ g/L, and *cis*-DCE was at 74  $\mu$ g/L, an order of magnitude lower than in 2011. TCE was not detected in Well MW-70-PMW03B between 2011 and 2015 (**Figure 5-68**). Cis-DCE was present and showed an increase from 380  $\mu$ g/L to 1700  $\mu$ g/L between 2011 and 2013. After the second round of EVO injection in 2013, however, *cis*-DCE dropped to 12  $\mu$ g/L, and then rebounded somewhat, being detected at 105  $\mu$ g/L in 12/2015. VC was consistently detected < 100  $\mu$ g/L in this well until the second EVO injection, after which VC increased to 240  $\mu$ g/L. During the 12/2013 sampling event, the concentration of VC had declined somewhat to 140  $\mu$ g/L.

#### 5.3.4.3.2. Field Parameters

The ORP of wells across the site covered a wide range from +61 mV in AMW-2 to -351 mV in MW70-PMW01B (**Table 5-18**). The wells in the region receiving the most recent treatment with EVO (i.e., the four MW70 wells) had the lowest OPR values, ranging from -232 mV to -351 mV. DO was < 1 mg/L in 9/10 well sampled, exceeding this value (1.98 mg/L) only in PMW-6. The groundwater pH was between 6.31 and 7.61 SU, which is in a range suitable for reductive dehalogenation.

#### 5.3.4.3.3. Anions

One of the characteristics of this site is naturally high sulfate, which can be inhibitory to reductive dehalogenation, likely due to the toxic action of hydrogen sulfide, which is the final product of sulfate reduction (Mao et al., 2017). Competition for H<sub>2</sub> as an electron donor between dehalogenating organisms and sulfate reducers also may contribute to issues observed at some sites (e.g., Aulenta et al., 2007). The sulfate concentrations during the 12/2015 sampling event ranged from 301 mg/L in MW70-PMW02B to 12,200 mg/L in AMW-2. The wells receiving the most recent treatment with EVO (i.e., the four MW70 wells) had the lowest sulfate values, indicative of substantial sulfate reduction. These wells also had the lowest ORP values, as previously noted. mV to -351 mV. Nitrate, nitrite, and phosphate were all below detection (< 0.2 mg/L) across the site.

## 5.3.4.3.4. Dissolved Iron and Manganese

As with other parameters, dissolved iron exhibited a range of concentrations across the site, from a low of  $<70 \ \mu g/L$  in AMW-2 to a high of 35,800  $\mu g/L$  in MW70-PMW01B. Dissolved manganese was detected in every well ranging from 129  $\mu g/L$  to 2,400  $\mu g/L$ . These data are indicative of iron and manganese reduction occurring in much of the groundwater across the site, which is consistent with the general ORP and DO values.

#### 5.3.4.3.5. Dissolved Gases

Methane was detected in groundwater across the site ranging from ~ 48  $\mu$ g/L to 10,900  $\mu$ g/L. The AMW and PMW wells from the early ESTCP study all had > 1,900  $\mu$ g/L, with two of the wells exceeding 7,000  $\mu$ g/L. Ethane was present in 5/10 wells sampled but at a maximum concentration of only 12  $\mu$ g/L. Ethene was more prevalent, being detected in 9/10 wells and ranging as high as 134  $\mu$ g/L. These gases (ethene and ethene) are often typically detected as final degradation products of chlorinated ethenes and are indicative of ongoing reductive dehalogenation. Hydrogen gas, the ultimate electron donor for reductive dehalogenation was below the MDL (0.0084  $\mu$ g/L) in all wells.

## 5.3.4.3.6. Total Organic Carbon and Volatile Fatty Acids

Total organic carbon (TOC) in the different wells ranged from 6.24 mg/L in MW70-PMW02B to 37 mg/L in AMW-2. Except for two wells (MW70-PMW02B and MW70-PMW03B), all wells had at least 15 mg/L. Lactic acid was observed in all of the wells except MW70-MNA01, the well upgradient of the ESTCP active plot. The highest concentrations were present in the two wells from the historical active ESTCP Plot, AMW-2, and AMW-5 (3.76 and 11.1 mg/L, respectively, and in two wells from the historical passive ESTCP Plot, PIW-2, and PMW-1 (9.57 and 35.2 mg/L, respectively). This is interesting in that these plots had not received the carbon source for several years before this sampling event. The presence of ethene in each of these wells is also indicative of ongoing dehalogenation activity.

#### 5.3.4.3.7. Microbial Community

The microbial community analysis indicated the presence of dehalogenation-associated organisms/genes in several water samples from the Seal Beach Site. Interestingly, no such organisms/genes were detected in upgradient well MW70-MNA01, which presumably represents non-augmented conditions at the site (Figure 5-69). Far downgradient wells MW-70-PMW02B and MW-70-PMW03B also showed largely non-detectable levels of dehalogenating organisms/genes. In contrast, each of the wells that were sampled from the former ESTCP passive treatment plot (PIW-1, PIW-2, and PMW-6) contained *Dehalococcoides* at  $\sim 9 \times 10^2$  to  $4 \times 10^3$ cells/mL. These wells also had detectable vcrA (enzyme degrades cis-DCE to ethene), tceA (enzyme degrades TCE to VC), and *btuF* (enzyme involved in the B12 synthesis pathway in Dehalococcoides). The genes bvcA (second enzyme that degrades cis-DCE to ethene) and cobS (associated with the vitamin B12 pathway), also were each detected in at least one of these wells. Thus, key organisms/genes involved in reductive dehalogenation were readily detectable in this previously bioaugmented passive treatment plot, but not in upgradient or far downgradient wells. The microbial data were similar to the passive plot for the wells representing full-scale treatment (MW-70-PMW01A and MW-70-PMW01B), which received EVO and KB-1 in 2010 and a second injection of EVO in 2013. As observed for the three passive treatment plot wells, each of these wells had detectable levels (generally between  $10^2$  and  $10^3/mL$ ) of *Dehalococcoides*, vcrA, bvcA, tceA, cobS and btuF, indicating the persistence of dehalogenating organisms in this region.

Interestingly, compared to the passive treatment plot, and the wells receiving full-scale treatment with EVO, *Dehalococcoides* and most reductive dehalogenation associated genes were absent in the well sampled from the historical active ESTCP plot (AMW-2), indicating that bioaugmented organisms did not persist as long under these conditions. Well AMW-5 was not sampled for molecular analysis.

## 5.3.4.3.8. CSIA.

Values of  $\delta^{13}$ C were obtained for TCE, cis-DCE, and VC in most of the monitoring wells where concentrations were sufficient for analysis (**Table 5-23**). It should be noted that  $\delta^{13}$ C values for TCE, as well as cis-DCE and VC, were obtained for a few samples in which the concentrations were below detection at a PQL of 5 µg/L. It is assumed in these cases that, while present below the PQL, there was enough VOC present for isotopic analysis. However, in a complex matrix, the possibility exists that compounds were not adequately separated before isotopic analysis (i.e., a second compound coeluted with the compound of interest and was analyzed). We assume this is not the case.

As noted previously, unfractionated  $\delta^{13}$ C values for manufactured TCE typically range from ~ -34‰ to - 23‰, with a mean of -29‰ (USEPA, 2008). The  $\delta^{13}$ C values for TCE at the Seal Beach site ranged from -25.0‰ to -0.04‰. The wells with the least fractionated TCE (i.e., closest to parent values) were AMW-2 (-23.9 ‰), AMW-5 (-22.6 ‰), MW70-MNA01 (-25.0 ‰), and PMW-6 (-22.2 ‰). Among these wells, AMW-2 and MW70-MNA01, were also noted to have very low concentrations of dehalogenating genes and organisms (AMW-5 was not tested). Consistent with these data, AMW-2, MW70-MNA01, and PMW-6 had the highest residual TCE among the wells tested. Two of the wells in the former ESTCP passive treatment plot, PIW-2, and PMW-1 had  $\delta^{13}$ C values of -13.8‰ and -17.4‰, respectively, indicative of ongoing biodegradation in this zone. Similarly, wells in the full-scale treatment area, MW-70-PMW01A, and MW-70-PMW01B, had  $\delta^{13}$ C values of -17.5‰ and -0.04‰, respectively, also indicative of ongoing TCE biodegradation, particularly in the latter well. Among the remaining wells, which are downgradient, TCE was below detection in MW70-PMW03B, and no  $\delta^{13}$ C value was obtained, and MW70-PMW02B showed evidence of continuing dehalogenation of TCE with a  $\delta^{13}$ C value of -18.6 ‰.

 $\delta^{13}$ C values for cis-DCE at the Seal Beach site ranged from -26.0 ‰ to + 4.2 ‰, again indicating significant degradation of *cis*-DCE in some areas of the site. In most cases, wells in which TCE biodegradation was indicated by CSIA also showed evidence of ongoing cis-DCE biodegradation (i.e.,  $\delta^{13}$ C values heavier than typical parent TCE) and vice versa. A similar trend was apparent for VC, with the exception perhaps of well MW70-PMW01B, which had  $\delta^{13}$ C values of -0.04‰ for TCE, + 4.2 ‰ for cis-DCE and - 25.9 ‰ for VC. The  $\delta^{13}$ C data, as well as the high VC concentration in this well (2,650 µg/L), may indicate a VC stall.

#### 5.3.4.3.9. PFMs.

Due to a scheduling conflict, PFMs were not deployed at the Seal Beach Site.

#### 5.3.4.3.10. Site Summary

Results from the Seal Beach site varied by treatment approach and, in some instances, by well in a given treatment area. VOC concentrations in upgradient well MW70-MNA01 declined since the conclusion of the ESTCP test project and daughter products were present, but CSIA data provided no indication of ongoing degradation, and relevant dehalogenating cultures and genes were absent.

In the former ESTCP active treatment plot, one well had high concentrations of TCE and daughter products, and the second well had non-detect concentrations. CSIA data showed no clear indication of ongoing degradation of TCE or daughter products. In the former ESTCP passive treatment area, 2/3 wells appeared to have ongoing degradation based on CSIA data. Microbial community analysis indicated the presence of dehalogenation-associated organisms/genes in several but not all locations at the Seal Beach Site. Wells sampled in the passive treatment plots contained *Dehalococcoides* at ~ 9 x  $10^2$  to 4 x  $10^3$  cells/mL, whereas *Dehalococcoides* was not detected in the active treatment cell. Groundwater data in both the active and passive cell suggest that TCE rebound had occurred in some areas, but below the historically high concentration. In the full-scale treatment area, all indications suggest ongoing VOC biodegradation. Taken in sum, the data suggest that biodegradation has persisted in the former ESTCP passive treatment plot ~3 years after injection of cultures and electron donor, whereas dechlorination has largely ceased in the former ESTCP cell that underwent active treatment. The full-scale area that received EVO and bioaugmentation culture also appears to have ongoing VOC biodegradation based on all relevant measures.

	LOCATION_CODE		AMW-2	AMW-5	MW70-MNA01	MW70-PMW01	A MW70-PMW01B	MW70-PMW02B	MW70-PMW03B	PIW-2	PMW-1	PMW-6
Class	SAMPLE_DATE		12/16/15	12/17/15	12/15/15	12/15/15	12/ 15/ 15	12/15/15	12/ 15/ 15	12/ 16/ 15	12/16/15	12/16/15
	Parameter	Units	Result	Result	Result	Result	Result	Result	Result	Result	Result	Result
Anions	Chloride	mg/L	1,910 D	2840 D	1590 D	2030 D	1410 D	185 D	263 D	1210 D	580 D	2380 D
Anona	Sulfate as SO4	mg/L	12,200 D	6680 D	2730 D	1490 D	504 D	301 D	419 D	4040 D	2,370 D	3120 D
	del13C c DCE	ppt	-24.87	-19.16	-23.2	0.01	4.16	-26	2.21	-14.95	-10.51	-21.54
CSIA	del13C PCE	ppt	NA	NA	NA	NA	NA	NA	NA	NA	-17.64 J	NA
00.71	del13C TCE	ppt	-23.89	-22.61	-25.02	-17.47	-0.04	-18.61	NA	-13.78	-17.35	-22.22
	del13CVC	ppt	-22.99	-22.97	-23.72	-0.87	-25.86	NA	-15.88	-21.49	-20.26	-30.76
	DO	mg/L	0.96	0.31	0.29	0.68	0.18	0.18	0.98	0.22	0.39	1.98
Field	ORP	m∨	61	-53	-141	-351.2	-232.2	-261.4	-252	-84	14.7	-47.1
	pH	su	7	6.85	6.67	6.7	6.31	7.48	7.29	7.26	7.61	6.96
	Ethane	μ <b>g/L</b>	4 U	3.02 J	4 U	7.79	12.0	4 U	4 U	4 U	4.58	1.89 J
Gases	Ethene	μg/L	6.51	67	16.7	134	41.8	5 U	21.1	23.2	19.3	34.1
	Methane Dissolved Iron	μg/L μg/L	3,340 D 70 U	7110 D 254	1270 457	5990 D 210	10900 D 35800	47.9 256	6750 D 3380	1970 1710	7,460 D 116	4550 D 9070
Metals	Dissolved Manganese	μg/L μg/L	491	1260	1300	129	2400	175	556	860	116	1830
TOC	Total Organic Carbon	mg/L	37	21.1	16.1	39.4	16.5	6.24	7.26	28.7	28.6	23.1
VFA	Lactic Acid	mg/L	3.76	11.1	10.1	11.1	0.43 J	1.1	0.77 J	9.57	35.2	0.57 J
MA.	1,1,1-tric hloroethane	μg/ L	5 U	4	5 U	5 U	50	5 U	5.0	5.07	5 U	5.07
	1,1-dichloroethane	μg/ L	1.1 J	18.8	5 U	15	50	5 0	5 U	5 0	5 0	5 U
	1,1-dichloroethylene	μg/ L	4.4 J	5 U	3.9 J	5 U	6.0	5 U	5 U	3.7 J	1.8 J	35.7
	1,2-dichloropropane	μg/ L	5 U	1010 D	5 U	24.4	5 U	5 U	5 U	5 U	5 U	5 U
	1,3-dichloropropane	μg/ L	5 U	<b>2.1</b> J	5 U	5 U	5 U	5 U	5 U	5 U	5 U	5 U
	2,2-dichloropropane	μg/L	5 U	5.4	5 U	5 U	5 U	5 U	5 U	5 U	5 U	5 U
	2-Butanone (MEK)	μg/ L	10 U	5 U	10 U	16	10 U	10 U	10 U	10 U	10 U	10 U
	bromochloromethane	μg/ L	5 U	1170 D	5 U	100	5 U	5 U	5 U	5 U	5 U	5 U
VOC	bromomethane	μg/ L	5 U	3280 D	5 U	203	5 U	5 U	5 U	5 U	5 U	5 U
	carbon disulfide	μg/ L	5 U	<b>10</b> U	5 U	10 U	18.5	5 U	5 U	5 U	5 U	5 U
	chloroform	μg/ L	178	5 U	5 U	5 U	<b>1.9</b> J	5 U	5 U	5 U	<b>1.6</b> J	7.6
	Cis 1,2- Dichloroethylene	μg/L	1280 D	5 U	634 D	5 U	777 D	74.0	105	148	23.2	1160 D
	methylene chloride	μg/ L	5 U	5.9	5 U	5 U	5 U	5 U	5 U	5 U	5 U	5 U
	tetrac hloroethylene	μg/ L	7.8	5 U	5 U	5 U	5 U	5 U	5 U	5 U	<b>2.1</b> J	<b>1.1</b> J
	trans-1,2-dic hloroethylene	μg/L	21.2	5 U	73.3	5 U	26.1	5 U	1.8 J	<b>4.3</b> J	2.8 J	18.8
	trichloroethylene	μg/ L	8130 D	5 U	1560 D	5 U	8.9	<b>3.3</b> J	5 U	219 D	75.9	4730 D
	vinyl chloride	μg/ L	248 D	5 U	436 D	5 U	2650 D	5 U	150	225 D	87.9	584 D

 Table 5-22.
 Sampling Results Summary AMW-2, Seal Beach Naval Weapons Station Site 70.

Notes: NA - Not Analyzed

U - Compound not detected above method practical quantitation limit.

D - Sample was diluted prior to analysis

J - Estimated value above MDL and less than PQL

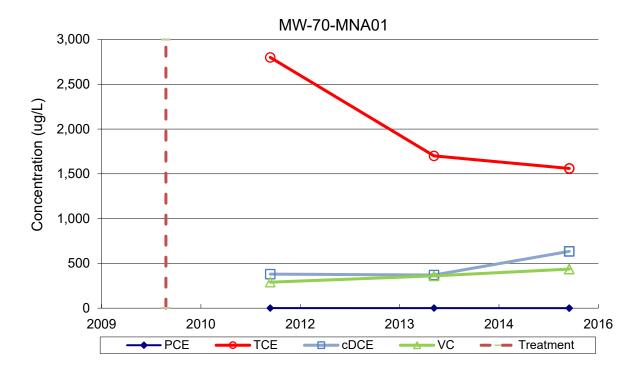


Figure 5-53. VOC Trends for MW70-MNA01, Seal Beach Naval Weapons Station Site 70.

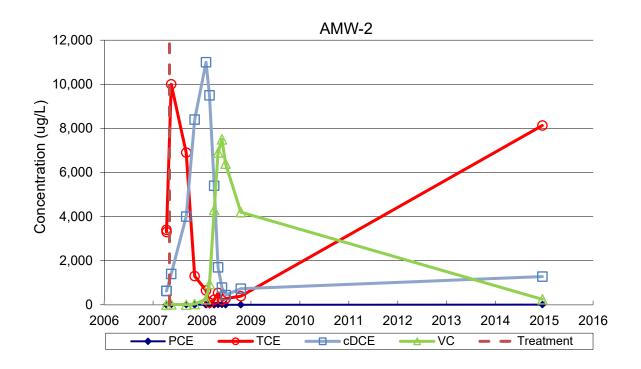


Figure 5-54. VOC Trends for AMW-2, Seal Beach Naval Weapons Station Site 70.

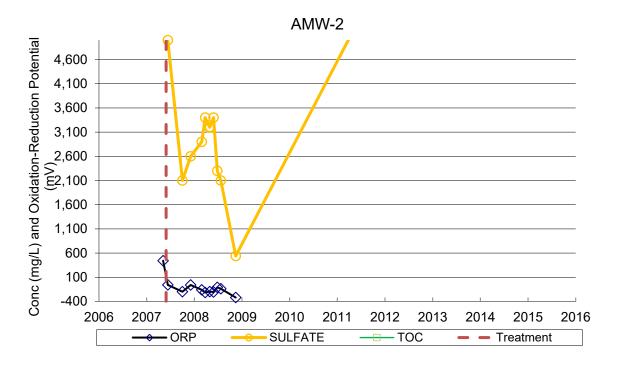


Figure 5-55. ORP and Sulfate Trends for AMW-2, Seal Beach Naval Weapons Station Site 70.

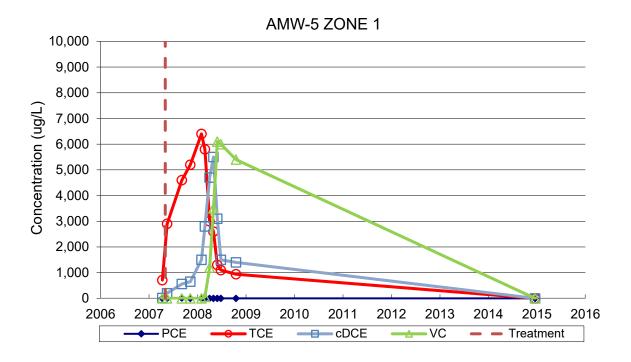


Figure 5-56. VOC Trends for AMW-5, Seal Beach Naval Weapons Station Site 70.

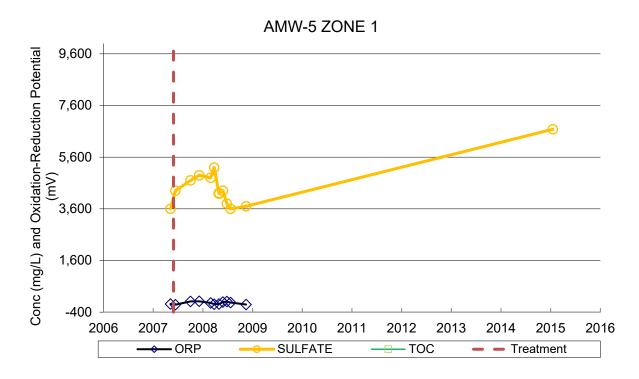


Figure 5-57. ORP and Sulfate Trends for AMW-5, Seal Beach Naval Weapons Station Site 70.

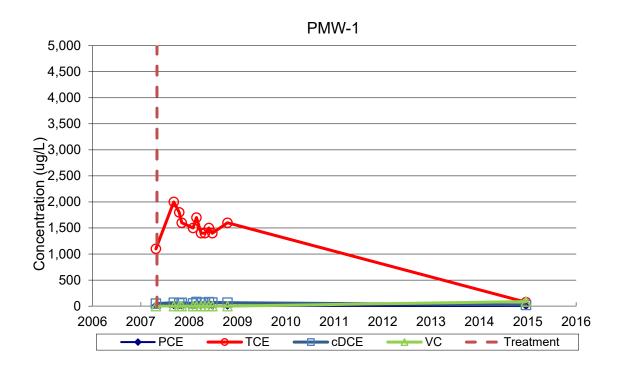


Figure 5-58. VOC Trends for PMW-1, Seal Beach Naval Weapons Station Site 70.

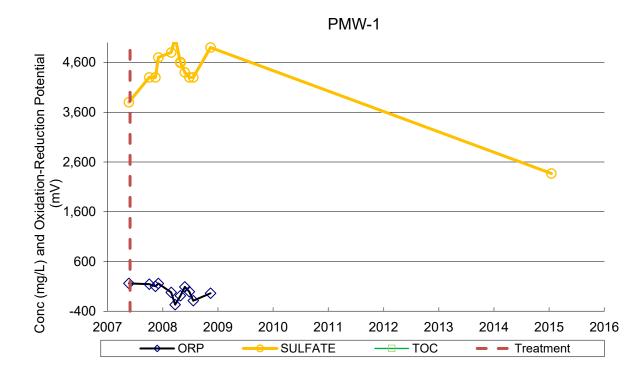


Figure 5-59. ORP and Sulfate Trends for PMW-1, Seal Beach Naval Weapons Station Site 70.

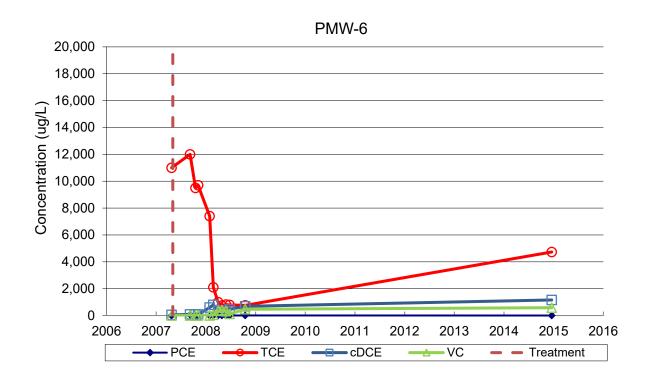


Figure 5-60. VOC Trends for PMW-6, Seal Beach Naval Weapons Station Site 70.

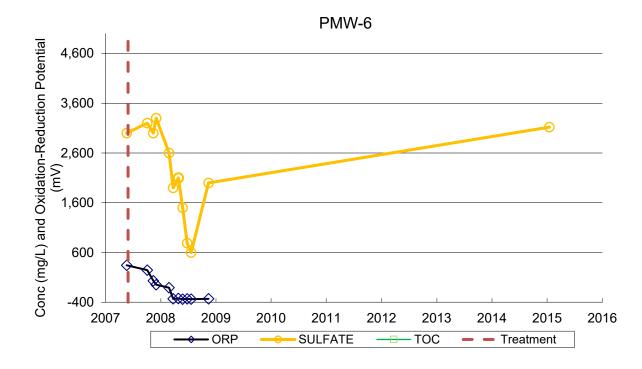


Figure 5-61.ORP and Sulfate Trends for PMW-6, Seal Beach Naval Weapons Station Site 70.

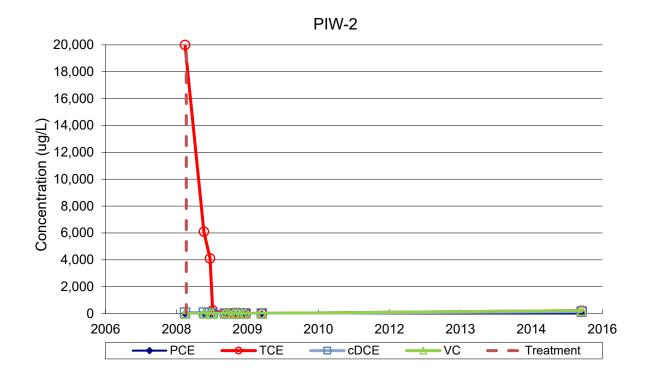


Figure 5-62. VOC Trends for PIW-2, Seal Beach Naval Weapons Station Site 70.

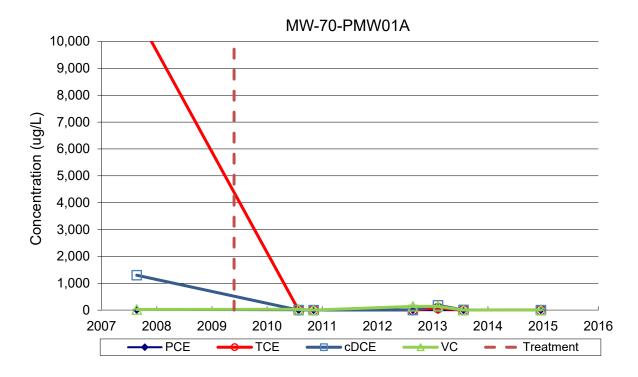


Figure 5-63. VOC Trends for MW70-PMW01A, Seal Beach Naval Weapons Station Site 70.

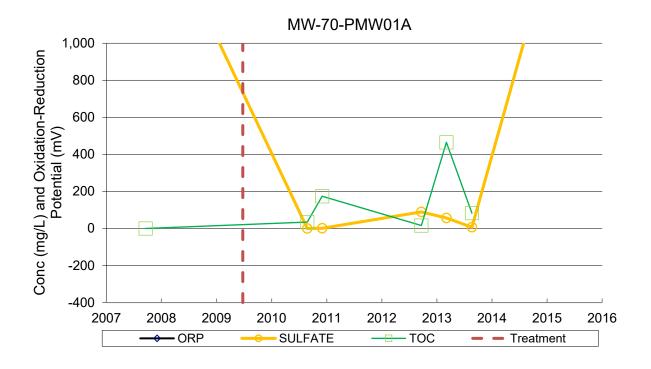


Figure 5-64. TOC and Sulfate Trends for MW70-PMW01A, Seal Beach Naval Weapons Station Site 70.

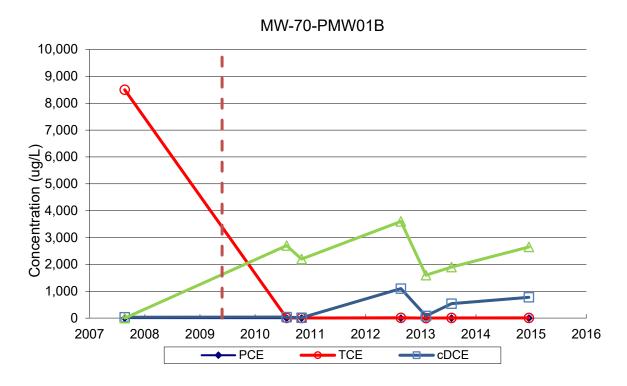


Figure 5-65. VOC Trends for MW70-PMW01B, Seal Beach Naval Weapons Station Site 70.

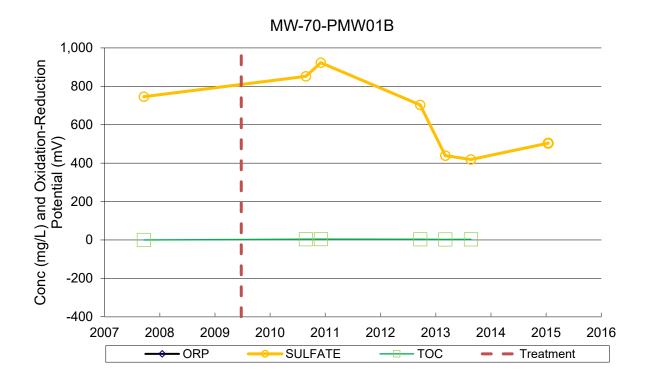


Figure 5-66. TOC and Sulfate Trends for MW70-PMW01B, Seal Beach Naval Weapons.

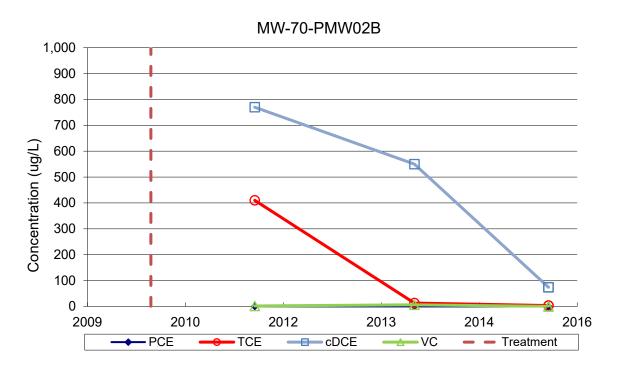


Figure 5-67. VOC Trends for MW70-PMW02B, Seal Beach Naval Weapons Station Site 70.

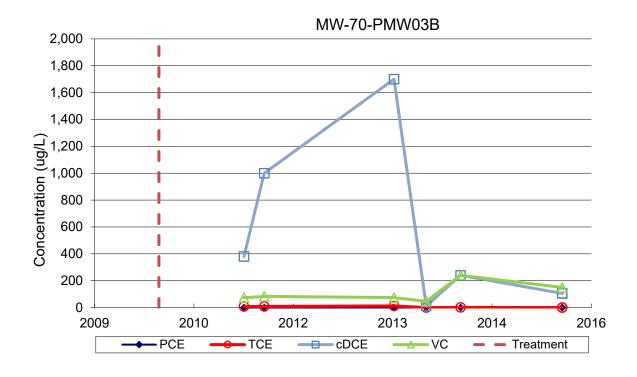


Figure 5-68. VOC Trends for MW70-PMW03B, Seal Beach Naval Weapons Station Site 70.

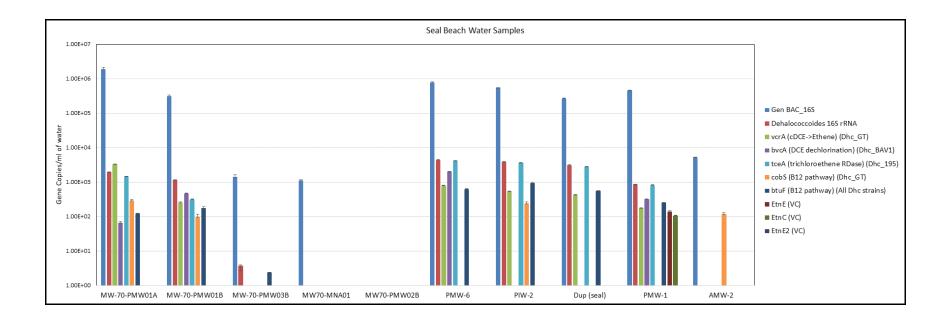


Figure 5-69. Organisms and Genes Associated with Reductive Dechlorination Detected in Wells at the Seal Beach Site.

Well	δ <sup>13</sup> C TCE (‰)	TCE (µg/L)	δ <sup>13</sup> C DCE (‰)	DCE (µg/L)	δ <sup>13</sup> C VC (‰)	VC (µg/L)	Location
MW70-MNA01	-25.0	1,560	-23.2	634	-23.7	436	Upgradient
AMW-2	-23.9	8,130	-24.9	1,280	-23.0	248	Active
AMW-5	-25.0	< 5	-23.2	< 5	-23.0	< 5	Active
PMW-1	-17.4	75.9	-10.5	23.2	-20.3	87.9	Passive
PIW-2	-13.9	219	-15.0	148	-21.	225	Passive
PMW-6	-22.2	4,730	-21.5	1,160	-30.8	584	Passive
MW70 -PMW01A	-17.5	< 5	-0.01	< 5	-0.9	< 5	Full-scale
MW70 -PMW01B	-0.04	8.9	+ 4.2	777	-25.9	2,650	Full-scale
MW70-PMW02B	-18.6	3.3	- 26.0	74	NA*	< 5	Downgradient
MW70-PMW03B	NA	< 5	+2.2	105	-15.9	150	Downgradient

 Table 5-23.
 CSIA Summary Seal Beach Naval Weapons Station Site 70.

\*NA - Data not available because concentration was too low for CSIA analysis.

#### 5.3.5 Treasure Island Naval Station Site 24

#### 5.3.5.1 Site Description, Treasure Island Naval Station Site 24

Site 24 at Naval Station Treasure Island is contaminated with PCE, and a plume of contamination extends over 1,100 ft to the San Francisco Bay (**Figure 5-70**). The main source area, Building 99, contained dry cleaning facilities that operated between 1942 and 1977. Total VOC concentrations above 1 mg/L occurred throughout much of the plume before remediation.

Treasure Island is a manmade island created from dredged fill material. The fill material contains discontinuous layers of sand, silt, clay, and shell hash to depths of up to 40 ft. The measured hydraulic conductivity of the material ranges from 5 to 16 ft per day, indicating a moderate permeability. A freshwater aquifer is present in the upper part of the fill material, but the groundwater becomes saline at depths of greater than 30 ft. The aquifer is generally anaerobic, has a neutral pH, and has relatively high sulfate concentrations.

An initial pilot study was conducted at the Building 99 source area from 2003 to 2004. The pilot study system included recirculation with three injection wells and three extraction wells arranged to provide three test loops. Sodium lactate, hydrogen gas, and SDC-9 were applied to various degrees in each of the loops. The pilot study results data showed a substantial decrease in contamination due to the treatment.

Following the successful pilot study, the full-scale treatment was implemented in three phases. Phase 1, conducted from November 2004 to May 2007, consisted of injecting and recirculating a lactic acid solution, hydrogen gas, and SDC-9 across the plume downgradient of the pilot study area. Phase 2, conducted from June 2008 to October 2010, consisted of a combination of recirculating a sodium lactate solution in some areas and injections of emulsified vegetable oil substrate in other areas.

The biotreatment was successful throughout much of the plume, but several pockets of contamination remained. Phase 3 was conducted from 2011 to 2012 to address the remaining hot spots of contamination and included additional source area recirculation with lactic acid and SDC-9, along with direct-push applications of Lactoil<sup>TM</sup> in other parts of the plume. Three areas there were targeted for additional treatment included the South Source Area Treatment Area, the EW12 Treatment Area, and the EW30 Treatment Area (**Figure 5-71**).

The bioremediation was considered effective, and no further active remediation is planned. This site was sampled to assess current conditions, including VOC and DHC concentrations, as well as a passive flux in different areas of the site. This site is an example of successful full-scale bioremediation.



Figure 5-70. Location Map, Treasure Island Naval Station Site 24.

## 5.3.5.2 Sampling Rationale, Treasure Island Naval Station Site 24

The objective of the groundwater sampling was to assess the long-term impacts of the previously performed bioaugmentation treatment on groundwater quality and biogeochemistry. Monitoring locations within the various treatment areas were evaluated, as described below. **Table 5-24** summarizes the monitoring locations and rationale for the sampling program. Monitoring wells sampled are shown in **Figure 5-71**.

Treasure Island Monitoring Well	Sampling Rationale
24-EW04	Located in the South Source Area. Phase 1 resulted in substantial reduction followed by a rebound to pre-treatment concentrations. Phase 2 treatment caused substantial reduction but without a subsequent rebound.
24-EW12	Located in the EW-12 Treatment Area. Substantial reduction was initially achieved following Phase 1. Some rebound occurred that was substantially reduced in subsequent treatments.
24-TW-11	Located in the South Source Treatment Area. Treated in Phases 1, 2, and 3. Initial rebound followed by substantial degradation.
24-TW-14	Located in the EW30 Treatment Area. Treated in Phase 2 only. Treatment was highly effective.
24-TW-41	Located in the EW30 Treatment Area. Treated in Phase 2 only. Treatment was highly effective.
24-TW-53	Located in the EW12 Treatment Area and received treatments during Phases 1, 2, and 3. Substantial accumulation of DCE and VC after Phase 1 but then substantial reduction in Phase 3.

Table 5-24.Groundwater Sampling Locations and Rationale, Naval Station TreasureIsland Site 24.

Monitoring wells 24-TW-11 and 24-TW-48 are located in the South Source Treatment Area. Well 24-MW-11 had repeated treatments in Phases 1, 2, and 3. Substantial rebound of contaminant concentrations occurred between treatments. Substantial degradation was observed in this well in the final treatment (Phase 3). Monitoring well 24-EW4 and 24-IW05 also had substantial reduction in Phase 1, followed by substantial rebound. Concentrations were substantially reduced in Phase 2.

Monitoring wells 24-TW-14 and 24-TW-53 are located at two other treatment areas at the site – the EW30 and EW12 Treatment Areas, respectively. Treatment was highly effective at well 24-TW-14 in the EW30 Treatment Area but not at well 24-TW-53 in the EW12 Treatment Area, where substantial accumulation of 1,2-dichloroethene and vinyl chloride was observed after treatment. Similar to the wells described above, these wells were selected to evaluate what conditions are associated with highly effective versus less effective treatment.

 Table 5-25 provides monitoring well construction details.

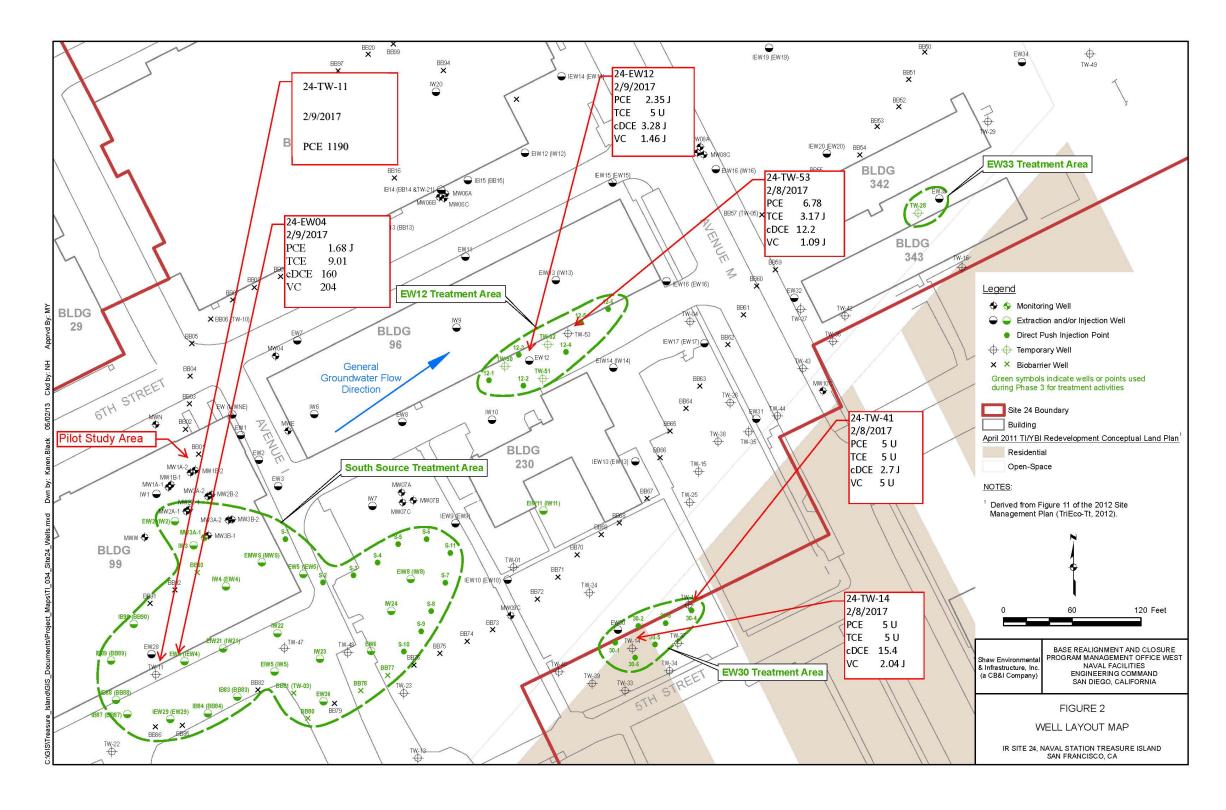


Figure 5-71. Well Layout Map, Naval Station Treasure Island Site 24.

Monitoring Well	Top of Screen (ft bgs)	Bottom of Screen (ft bgs)	Diameter (inches)
24-EW04	10	30	4
24-EW-12	10	30	4
24-TW-11	10(e)	30(e)	1
24-TW-14	5	29.7	1
24-TW-41	5	29.7	1
24-TW-53	5	30	1

 Table 5-25.
 Monitoring Well Construction Details, Naval Station Treasure Island Site 24.

bgs = below ground surface

(d) = dual screen

(e) = estimated

## 5.3.5.3 Sampling Results, Treasure Island Naval Station Site 24

#### 5.3.5.3.1. VOCs

**Table 5-26** provides a summary of sampling results, and **Figures 5-72** to **5-82** provide historical and current VOC data and corresponding ORP and sulfate data for the different wells sampled. The primary parent contaminant at this site was PCE. The highest concentration of residual PCE was detected in 24-TW-11, in the South Source Treatment Area, which had 1,190  $\mu$ g/L as well as 779  $\mu$ g/L of TCE, 7,840  $\mu$ g/L of cis-DCE and 937  $\mu$ g/L of VC, respectively. This well has shown rebound of PCE after various treatment phases with levels as high as 16,000  $\mu$ g/L detected in 2011, but minimal rebound was observed after treatment in 2012 until this sampling event approximately 5 years later (**Figure 5-72**). The other well in this same treatment area, 24-EW04 had much lower levels of VOCs than 24-TW-11, having PCE at 1.7  $\mu$ g/L, TCE at 9.0  $\mu$ g/L, and VC at 204  $\mu$ g/L (**Figure 5-74**). This well had concentrations as high as 36,000  $\mu$ g/L in 2005. The other 4 wells that were sampled at Site 24 had residual PCE < 7  $\mu$ g/L, cis-DCE < 16  $\mu$ g/L, and VC < 3  $\mu$ g/L. Thus, overall treatment effectiveness in the EW-12 and EW-30 areas showed little to no rebound after treatment. These wells were primarily contaminated with *cis*-DCE and VC during previous years.

#### 5.3.5.3.2. Field Parameters

Except for well 24-TW-53, which had a measured ORP of +66.5 mV, a negative ORP was measured in all other wells tested ranging from -257.6 mV in 24-EW04 to -23.2 mV in 24-TW-41 (**Table 5-26**). Thus, even though the region had not presumably received any additional electron donor in ~ 5 years before this sampling event, groundwater ORP remained predominantly negative. The one well showing significant rebound, 24-TW-11 had an ORP of -145. DO was < 0.6 mg/L in all of the wells sampled. The groundwater pH was between 6.23 and 6.98 SU, which is in a range suitable for reductive dehalogenation.

#### 5.3.5.3.3. Anions

Sulfate concentrations in groundwater at the site ranged from ~ 200 to 400 mg/L in the early 2000s (**Figures 5-73, 5-75, 5-77, 5-79, 5-81**). The sulfate concentrations during the current sampling event ranged from a low of 22.9 mg/L in 24-TW-11 to a high of 202 mg/L in 24-TW-14. Most of the wells had much lower sulfate concentrations than measured before or during the active treatment phases, indicating continued sulfate reduction across most of the site. This is consistent with the negative ORP values as well. Nitrate was detected at 0.88 mg/L in 24-TW-53, the well that also had a positive ORP, but was < 0.2 mg/L in all other wells. Interestingly, orthophosphate was detected in 4 of the 7 wells at > 1 mg/L. This is unusual since phosphate forms insoluble precipitates in most environments.

#### 5.3.5.3.4. Dissolved Iron and Manganese

Dissolved iron ranged from  $<70 \ \mu g/L$  in 24-TW-53 (ORP +66.5) to 268,000  $\mu g/L$  in 24-EW04 (ORP -257.6). Dissolved manganese was detected in every well ranging from 343  $\mu g/L$  in 24-TW-53 to 7,800  $\mu g/L$  in 24-EW04. These data are indicative of iron and manganese reduction occurring in much of the groundwater across the site, consistent with the general ORP values.

#### 5.3.5.3.5. Dissolved Gases

Methane was detected in all wells ranging from 10  $\mu$ g/L in 24-TW-41 to 3,920  $\mu$ g/L in 24-TW-11. Ethane was present in 5/6 wells sampled, at concentrations ranging from 4.6 to 68  $\mu$ g/L. Ethene was only detected > 5  $\mu$ g/L in two wells, 24-EW04 (31  $\mu$ g/L) and 24-TW-11 (107  $\mu$ g/L), both of which are in the South Source Treatment Area, and have the highest residual VOCs overall. Ethene and ethane are indicative of ongoing reductive dehalogenation. Hydrogen gas, the ultimate electron donor for reductive dehalogenation was detected in 5/6 wells, with well 24-EW04 exceeding 1.5  $\mu$ g/L.

#### 5.3.5.3.6. Total Organic Carbon and Volatile Fatty Acids

TOC was < 11 mg/L in 4/6 wells. In the other two, 24-EW04 and 24-TW-11, TOC measured 1,520 mg/L and 212 mg/L, respectively. Each of these wells also had appreciably quantities of fatty acids present, presumably as degradation products of the Lactoil<sup>TM</sup> injected in 2012 during Phase 3 treatment. Fatty acids were not detected in any of the other wells.

#### 5.3.5.3.7. Microbial Community

The microbial community analysis indicated the presence of quantifiable dehalogenationassociated organisms/genes only in well 24-TW-11 (**Figure 5-83**). This was one of two wells (along with 24-EW04) that showed significant residual concentrations of VOCs as well as the presence of fatty acids to support microbial growth. *Dehalococcoides* was detected in this well at ~ 1 x 10<sup>3</sup> cells/mL, and the *vcrA* gene was measured at approximately the same concentration. The only other organism/gene detected was the *tceA* gene at slightly > at ~ 1 x 10<sup>2</sup> cells/mL. Among the other wells, it is not surprising that dehalogenating organisms/genes were not detected in 24-EW12, 24-TW-53, 24-TW-14, or 24-TW-41 as concentrations of VOCs in these wells were all < 20 µg/L, and TOC was very low with no fatty acids were detected. The moderately surprising result was that no quantifiable dehalogenation-associated cells or genes were detected in 24-EW04, as this well is close to 24-TW-11 and does have some low to moderate concentrations of both cis-DCE and VC remaining (~ 160 µg/L and 204 µg/L, respectively) as well as high concentrations of fatty acids. However, overall VOC concentrations in this well were appreciably lower than in 24-TW-11, which also had TCE and PCE concentrations each exceeding 750 µg/L. PCE and TCE in 24-EW04 were each < 10 µg/L.

#### 5.3.5.3.8. CSIA

Values of  $\delta^{13}C$  were obtained for TCE, cis-DCE, and VC in monitoring wells where concentrations were sufficient for analysis (Table 5-27). The primary contaminant at this site was initially PCE. Unfractionated  $\delta^{13}$ C values for manufactured PCE typically range from ~ -37‰ to - 24‰, with a mean of -29‰, very similar to TCE (USEPA, 2008).  $\delta^{13}$ C values were not obtained for PCE at any of the well locations. The  $\delta^{13}$ C values for residual TCE in the three wells where it was measurable ranged from -24.4‰ to -29.5‰. These values do not necessarily suggest that continued degradation of TCE is occurring as they are not above the range for parent PCE. It should be noted, however, that concentrations at two of the wells (24-TW-53 and 24-EW04) were  $< 10 \mu g/L$  at the time of sampling, and that dehalogenating organisms/genes were not detected in the wells.  $\delta^{13}$ C values for cis-DCE at the Treasure Island site ranged from -24.1% to + 8.8% in the 5/6 wells where concentrations were high enough to measure. The isotope analysis shows clear evidence for continued cis-DCE biodegradation in 24-MW-41, 24-EW04, and 24-TW-14. It should be noted that the concentration of cis-DCE in two of these wells  $(24-MW-41 \text{ and } 24-TW-14) \text{ was} < 20 \mu g/L$ . The  $\delta^{13}$ C value for cis-DCE in 24-TW-11 (-24.1‰), while in the general range of parent PCE, was ~ 5‰ heavier than the comparable value for TCE (-29.1 %.), which does indicate potential ongoing biodegradation. Finally, for VC, CSIA analysis indicated ongoing degradation in 24-EW04 ( $\delta^{13}$ C of -12.9‰), and 24-EW12 ( $\delta^{13}$ C of -1.1‰). The other wells for which  $\delta^{13}$ C of VC was measured had values in near or lower than the parent compound (PCE in this case); thus, evidence of its continuing degradation is not provided by the analysis.

Overall, the CSIA data from this site suggest a rather complex scenario. Wells in the EW30 area, where concentrations of VOCs are very low, each show isotopic evidence of ongoing *cis*-DCE degradation. One of the two wells in the South Source Zone Treatment Area (24-EW04) showed clear evidence of both cis-DCE and VC degradation, but no measurable numbers of relevant dehalogenating organisms/genes. The other well in this region (24-TW-11) showed some evidence of ongoing cis-DCE biodegradation, and very light VC (-38.9‰) indicating that VC was being formed from cis-DCE (i.e., very light daughter product is expected initially as a parent VOC degrades), but probably not biodegrading further. This well had detectable dehalogenating organisms/genes and the highest residual VOC concentrations. For the remining wells in the EW12 area, VC is clearly still degrading in 24-EW12 but not necessarily in 24-EW-53 (although residuals are exceedingly low in both wells).

	LOCATION_CODE		24-EW0	4	24-EW	12	24-TW-	11	24-TW-	14	24-TW-	41	24-TW-5	53
Class	SAMPLE_DATE 2/9/17		2/9/17	2/9/17 2/8/17		2/8/17	2/8/17		2/8/17					
	Parameter	Units	Result		Result		Result		Result		Result		Result	
	Chloride	mg/L	678	D	492	D	443	D	348	D	661	D	288	D
Anions	Nitrate as N	mg/L	0.2	U	0.2	U	0.2	U	0.2	U	0.13	J	0.88	
AHIOHS	Phosphate as P, ortho	mg/L	0.2	U	1.51		1.60		1.76		0.27		1.54	
	Sulfate as SO4	mg/L	51.2		33.9		22.9		202	D	123	D	64.2	D
	del13C cDCE	ppt	-1.16		NA		-24.07		-16.18		8.79		-22.26	
CSIA	del13C TCE	ppt	-29.52		NA		-29.10		NA		NA		-24.43	
	del13C VC	ppt	-12.93		-1.12		-38.89		-25.19		NA		-28.13	
	DO	mg/L	0.01		0.17		0.12		0.19		0.23		0.58	
Field	ORP	mV	-257.6		-115.7		-145		-63.9		-23.2		66.5	
	pН	su	6.23		6.9		6.42		6.98		6.71		6.76	
	Ethane	μg/L	10.8		67.9		27.4		4	-	4.59		9.32	
Gases	Ethene	μg/L	31.4		5	U	107		÷	U	,	U	5	U
00363	Hydrogen	μg/L	1.65	D	0.0046		0.121		0.0511		0.0559		0.0184	
	Methane	μg/L	1390		342		3920	D	96.5		10.0		345	
Metals	Dissolved Iron	μg/L	268000		7220		97100		1070		1650		70	U
	Dissolved Manganese	μg/L	7800		1170		4560		622		720		343	
TOC	Total Organic Carbon	mg/L	1520		10.9		121		9.17		8.90		7.95	
	Acetic Acid	mg/L	1450			U	544	-	1	-	-	U	1	-
	Butuyric Acid	mg/L	4150			U	85.4	_	1	-		U	1	-
VFA	Lactic Acid	mg/L	20		•	U	20	-		U	0.6	-	1	•
	Propionic Acid	mg/L	3490			-	527	D	-	U	1	U	1	1
	Valeric Acid	mg/L	1280			U	13.2			U	1	U		U
	1,1-dichloroethylene	μg/L	-	U	-	-	14.5		5	U	-	U	5	U
	2-Butanone (MEK)	μg/L	1090		10	U	280	D	10	U	10	U	10	U
	Acetone	μg/L	9.33		10	U	5.72		10	U	10	U	10	U
	carbon disulfide	μg/L	30.9		5	U	5	U	5	U	5	U	5	U
VOC	Cis 1,2- Dichloroethylene	μg/L	160	U	3.28	J	7840	D	15.4		2.70	J	12.2	
	tetrachloroethylene	μg/L	1.68	J	2.35	J	1190	D	5	U	5	U	6.78	
	trans-1,2-dichloroethylene	μg/L	5	U	5	U	52.3		5	U	5	U	5	U
	trichloroethylene	μg/L	9.01		5	U	779	D	5	U	5	U	3.17	J
	vinyl chloride	μg/L	204		1.46	J	937	D	2.04	J	5	U	1.09	J

# Table 5-26. Sampling Results Summary, Treasure Island Naval Station Site 24.

Notes:

NA - Not Analyzed

U - Compound not detected above method practical quantitation limit.

D - Sample was diluted prior to analysis

J - Estimated value above MDL and less than PQL

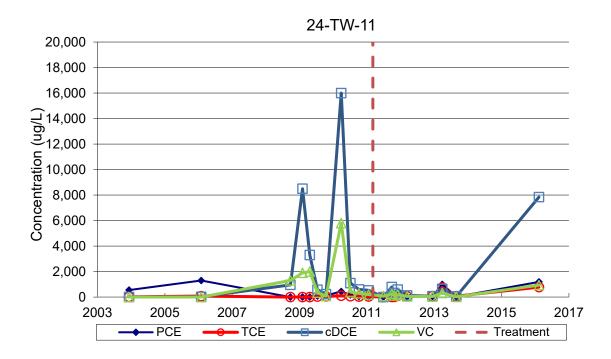


Figure 5-72. VOC Trends for 24-TW-11, Treasure Island Naval Station Site 24.

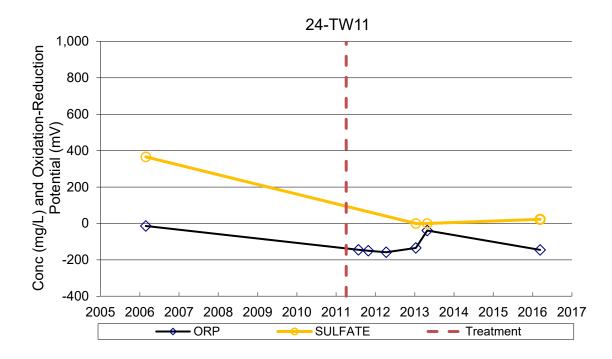


Figure 5-73.ORP and Sulfate Trends for 24-TW-11, Treasure Island Naval Station Site 24.

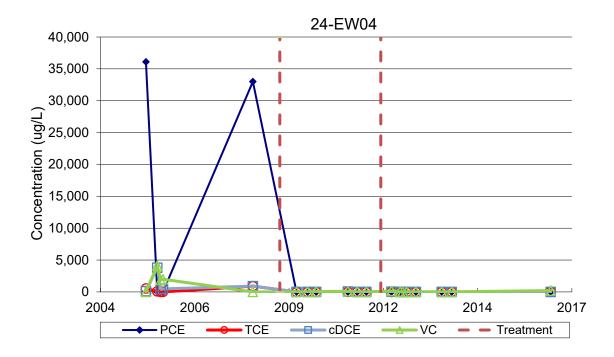


Figure 5-74. VOC Trends for 24-EW04, Treasure Island Naval Station Site 24.

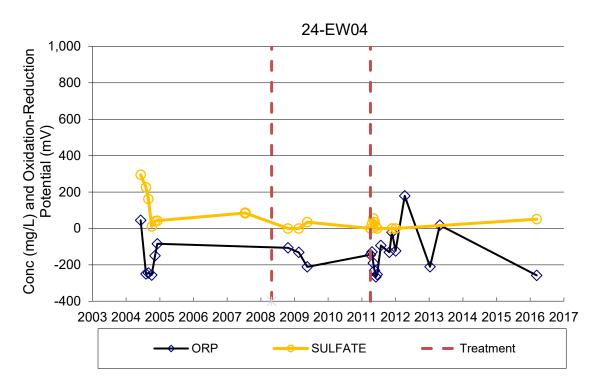


Figure 5-75. ORP and Sulfate Trends for 24-EW04, Treasure Island Naval Station Site 24.

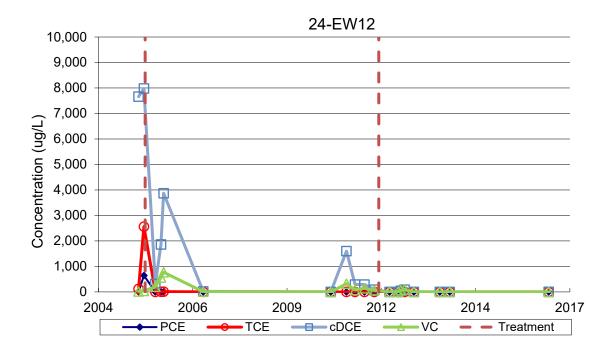


Figure 5-76. VOC Trends for 24-EW12, Treasure Island Naval Station Site 24.

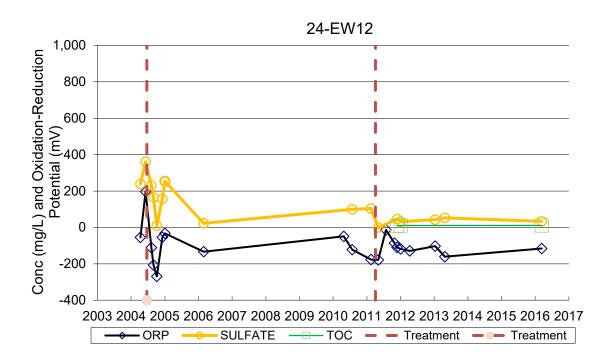


Figure 5-77.ORP and Sulfate Trends for 24-EW12, Treasure Island Naval Station Site 24.

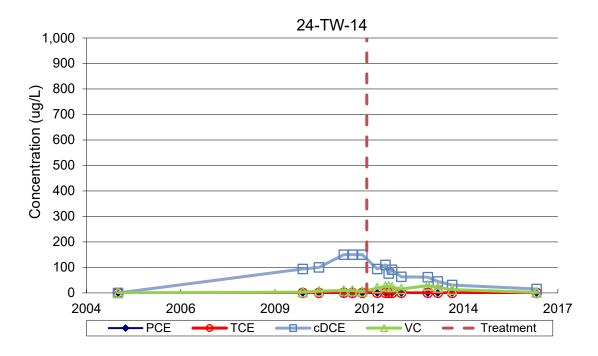
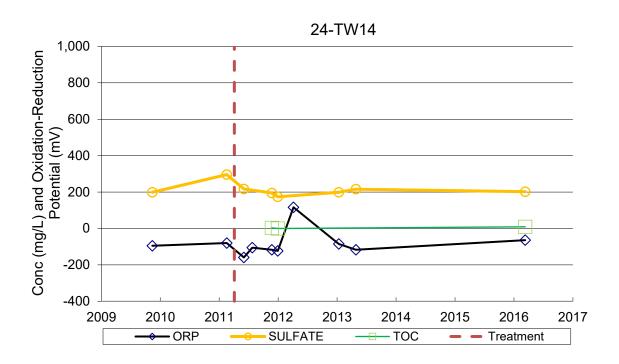
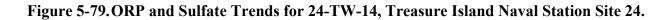


Figure 5-78. VOC Trends for 24-TW-14, Treasure Island Naval Station Site 24.





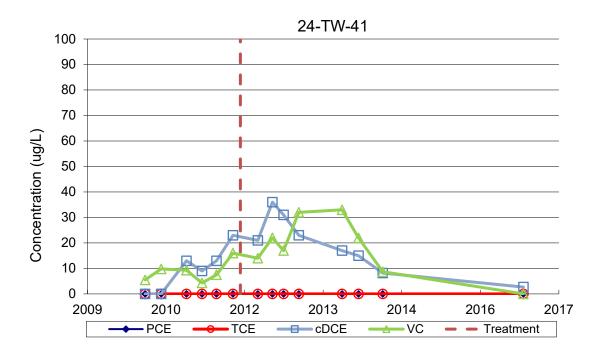
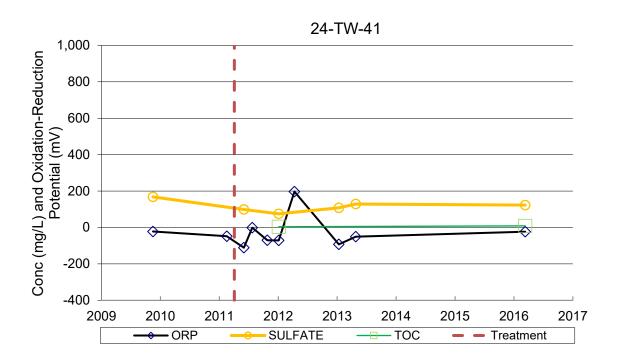
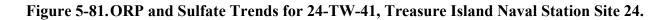


Figure 5-80. VOC Trends for 24-TW-41, Treasure Island Naval Station Site 24.





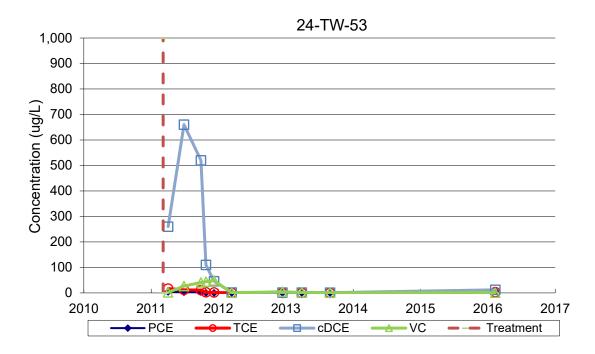


Figure 5-82. VOC Trends for 24-TW-53, Treasure Island Naval Station Site 24.

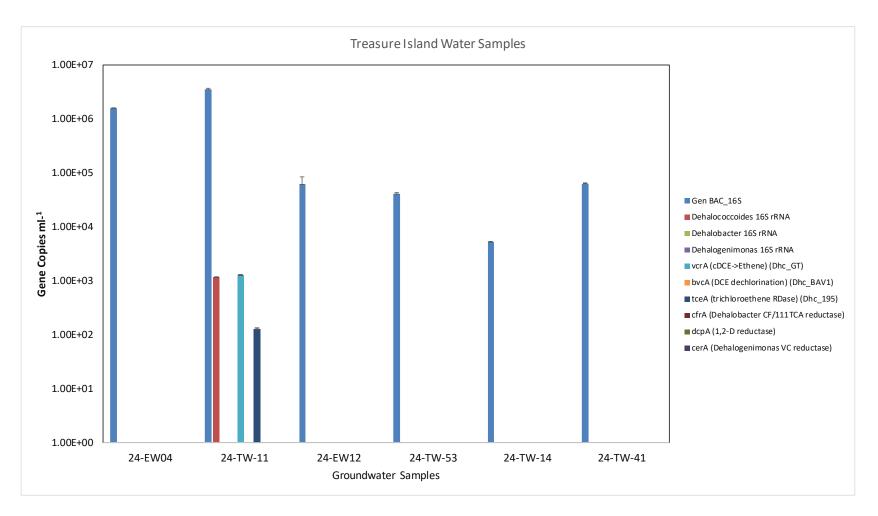


Figure 5-83. Organisms and Genes Associated with Reductive Dechlorination Detected in Wells at the Treasure Island Site.

Well	δ <sup>13</sup> C TCE (‰)	TCE (µg/L)	δ <sup>13</sup> C DCE (‰)	DCE (µg/L)	δ <sup>13</sup> C VC (‰)	VC (µg/L)	Location
24-EW04	-29.5	9.0	-1.2	160	-12.9	204	South Source
24-TW-11	-29.1	779	-24.1	7,840	-38.9	937	South Source
24-TW-14	NA	< 5	-16.2	15.4	-25.2	2.0	EW30
24-TW-41	NA	< 5	+ 8.8	2.7	NA	< 5	EW30
24-TW-53	-24.4	3.2	-22.3	12.2	-28.1	1.1	EW12
24-EW12	NA	< 5	NA	3.3	-1.1	1.5	EW12

 Table 5-27.
 CSIA Summary Treasure Island Site.

\*NA – Data not available because concentration was too low for CSIA analysis.

### 5.3.5.3.9. PFMs.

PFMs were deployed in EW4, EW12, and TW-41 (**Figure 5-84 & Table 5-28**). No chlorinated ethenes or ethene were detected in 24-TW41 and 24-EW12. No TCE or PCE were detected in 24-EW4, but an ethene flux of 14.1 mg/m<sup>2</sup>/day, VC flux of 11 mg/m<sup>2</sup>/day, and cis-DCE flux of 6.9 mg/m<sup>2</sup>/day were detected. These data suggest that complete dechlorination is occurring, but likely not at a rate fast enough to prevent the accumulation of VC and cis-DCE.

The flux-averaged equivalent TCE concentration (TCE+cis-DCE+VC), estimate from the PFM results and groundwater monitoring wells, are compared in **Table 5-29**. In well 24-TW41 groundwater concentration data detected very low cis-DCE while the PFMs found no chlorinated ethenes. For Well 24-EW4, groundwater and PFM data showed that TCE was below detection but that VC, ethene and cis-DCE were present. PFM data from Well 24-EW12 showed no detection of chlorinated ethenes whereas groundwater samples showed trace levels of (< 3.3  $\mu$ g/L) of PCE, cis-DCE, and VC. Overall, data were comparable. As noted previously, for areas where there was some disagreement in results, flux-averaged values (from PFMs) may better estimate concentrations than groundwater data because flux-average concentration is independent of divergence through the well and is both temporally and spatially average concentration as opposed to "instantaneous" measurements of groundwater data (Basu et al., 2006; Brooks et al., 2008). Nonetheless, both measures suggest that reductive dechlorination is still occurring at the site.

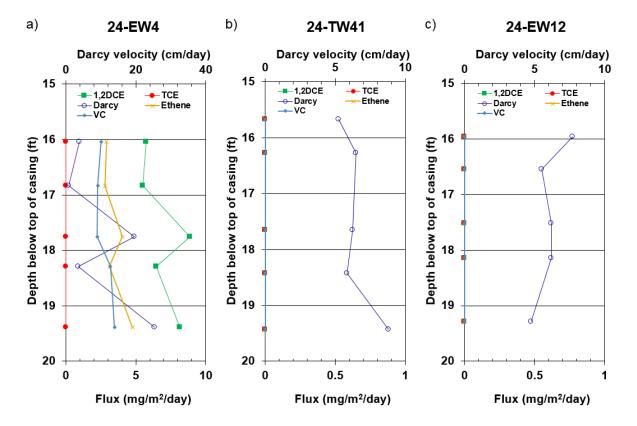


Figure 5-84. Mass Flux Profiles Measured in Select Wells Using the PFMs.

The yellow x's represents ethene, solid blue diamonds represent VC, the solid dark green squares represent DCE, the solid red circles represent TCE, and the open purple circles represent the Darcy Velocity. Panels a-e show flux values sampled on 02/24/2017. Note the changes in scale on both axes to accommodate the data.

Well_ID	Average Darcy Velocity (cm/day)	Average Ethene flux (mg/m²/day)	Average VC flux (mg/m²/day)	Average 1,2- DCE flux (mg/m²/day)	Average TCE flux (mg/m²/day)
24-EW4	10.7	14.1	11.0	6.9	0.00
24-TW41	6.5	0.00	0.00	0.00	0.00
24-EW12	6.1	0.00	0.00	0.00	0.00

 Table 5-28.
 Average Mass Discharge for Each Well.

Chemical	Туре	24-TW41	24-EW4	24-EW12
Equiv. TCE	GW	3.7	433	na
	PFM	<dl< td=""><td>3188</td><td><dl< td=""></dl<></td></dl<>	3188	<dl< td=""></dl<>
	% diff.	na	152	na
TCE	GW	<dl< td=""><td><dl< td=""><td><dl< td=""></dl<></td></dl<></td></dl<>	<dl< td=""><td><dl< td=""></dl<></td></dl<>	<dl< td=""></dl<>
	PFM	<dl< td=""><td>0</td><td><dl< td=""></dl<></td></dl<>	0	<dl< td=""></dl<>
	% diff.	na	na	na
cis-DCE	GW	2.7	3.3	3.3
	PFM	<dl< td=""><td>658</td><td><dl< td=""></dl<></td></dl<>	658	<dl< td=""></dl<>
	% diff.	na	198	na
VC	GW	<dl< td=""><td>204</td><td>1.5</td></dl<>	204	1.5
	PFM	<dl< td=""><td>1092</td><td><dl< td=""></dl<></td></dl<>	1092	<dl< td=""></dl<>
	% diff.	na	137	na
Ethene	GW	<dl< td=""><td>31</td><td><dl< td=""></dl<></td></dl<>	31	<dl< td=""></dl<>
	PFM	<dl< td=""><td>1336</td><td><dl< td=""></dl<></td></dl<>	1336	<dl< td=""></dl<>
	% diff.	na	191	na
Ethane	GW	4.6	11	68
	PFM	na	na	na
	% diff.	na	na	na

Table 5-29.VOC Concentration and GW and PFM Concentration Comparisons for<br/>Wells in Treasure Island.

Note: na = not available; <dl -= below detection

### 5.3.5.3.10. Site Summary

Results showed that biogeochemical conditions remained generally favorable for reductive dechlorination of chlorinated ethenes. Residual PCE was detected in 24-TW-11, in the South Source Treatment Area at 1,190 µg/L. This well also had 779 µg/L of TCE, 7,840 µg/L of cis-DCE and 937 µg/L of VC, respectively. This well showed rebound of PCE after various treatment phases with levels as high as 16,000 µg/L in 2011, but minimal rebound was observed after treatment in 2012 until this sampling event approximately 5 years later. The other well in this same treatment area, 24-EW04 had much lower levels of VOCs than 24-TW-11, with PCE at 1.7 µg/L, TCE at 9.0  $\mu$ g/L, and VC at 204  $\mu$ g/L. This well had concentrations as high as 36,000  $\mu$ g/L in 2005. The other 4 wells that were sampled at Site 24 had residual PCE < 7  $\mu$ g/L, cis-DCE < 16  $\mu$ g/L, and VC < 3  $\mu$ g/L. Thus, overall treatment effectiveness in the EW-12 and EW-30 areas showed little to no rebound after treatment. These wells were primarily contaminated with cis-DCE and VC during previous years. Both groundwater and the limited amount of PFM data detected the presence of ethene and ethane at the site, suggesting that complete biological reductive chlorination was still occurring ~5 years after the cessation of active treatment. The microbial community analysis indicated the presence of quantifiable dehalogenation-associated organisms/genes only in well 24-TW-11, which is the well with highest residual VOCs. It is likely that dehalogenators are not present in the vicinity of other wells due to the exceedingly low residual VOC concentrations. Overall, the CSIA data from this site suggest a rather complex scenario.

Wells in the EW30 area, where concentrations of VOCs are very low, each show isotopic evidence of ongoing *cis*-DCE degradation. One of the two wells in the South Source Zone Treatment Area (24-EW04) showed clear evidence of both cis-DCE and VC degradation, but no measurable numbers of relevant dehalogenating organisms/genes. The other well in this region (24-TW-11) showed some evidence of ongoing cis-DCE biodegradation, and very light VC (-38.9‰) indicating that VC was being formed from cis-DCE (i.e., very light daughter product is expected initially as a parent VOC degrades), but probably not biodegrading further. This well had detectable dehalogenating organisms/genes and the highest residual VOC concentrations. For the remining wells in the EW12 area, VC is clearly still degrading in 24-EW12 but not necessarily in 24-EW-53 (although residuals are exceedingly low in both wells).

# 5.4 STATISTICAL ANALYSIS

The historical data analysis included a statistical evaluation comparing measures of effectiveness to several possible factors for success. The evaluation was conducted on a well-by-well basis for each of the treatment zones and treatment phases at the 15 evaluation sites.

### Measures of Effectiveness

Measures of effectiveness were developed to provide quantitative metrics for the statistical evaluation as described below:

### Mass Reduction:

Mass reduction is a key measure of effectiveness and was calculated as a percent reduction of the mass concentrations of VOCs (typically presented in  $\mu g/L$ ) from the initial to final conditions.

### Dechlorination:

Dechlorination is a measure of the extent to which the dechlorination process has proceeded and for this application is equivalent to the number of chlorine atoms removed from the VOC parent molecules. Dechlorination was calculated by comparing the molar concentrations of chlorine atoms from the initial to final conditions.

### Risk Reduction:

The MCL for each VOC was used as an estimator of regulatory risk. The sum of the MCL multipliers (concentration/MCL) for each VOC was compared in the initial and final conditions.

### Rebound:

Rebound refers to the increase in concentrations after an initial decrease. Rebound was calculated by comparing the minimum concentration attained to the initial and final concentrations.

### DCE Accumulation:

DCE accumulation was calculated by comparing the ratio of DCE to Total VOC from initial to final conditions.

## VC Accumulation:

Similar to as described above for DCE Accumulation, VC Accumulation was calculated by comparing the ratio of VC to total VOC from Initial to Final conditions.

#### Factors for Success

The Measures of Effectiveness described above were compared to a number of potential factors for success, including those listed below.

### *Hydrogeologic factors:*

- Percent fines
- Heterogeneity
- Hydraulic conductivity
- Hydraulic gradient
- Seepage velocity

### Treatment Approach factors:

- Substrate used
- Substrate Loading (mass per treatment area volume)
- Substrate Dosing (injection fluid concentration)
- Nutrient amendments
- Treatment approach (for example recirculation, biobarrier, or direct push)
- Spacing of injection points

### *Geochemical factors:*

- DO
- ORP
- pH
- alkalinity
- iron
- manganese
- sulfate
- sulfide
- TOC
- temperature

## Biological factors:

- Abundance of DHC
- Abundance of reductase genetic markers

Data Field	Definition
Bioaug	An indication of whether bioaugmentation was used and, if so, what culture was employed.
Bioaug Rank	Y = Yes $N = No$
cis-DCE Accumulation	The difference between the Initial cis-DCE and Final cis-DCE concentrations as ratios for total CE at each individual well.
	Calculated as: (Final cis-DCE/ Final CE) – (Initial cis-DCE/ Initial CE)
	Where:Initial cis-DCE = Initial cis-1,2-dichloroetheneconcentration ( $\mu$ g/L);Initial CE = Initial total CE concentration ( $\mu$ g/L);Final cis-DCE = Final cis-1,2-dichloroethene concentration ( $\mu$ g/L);Final CE = Final total CE concentration ( $\mu$ g/L);Final CE = Final total CE concentration ( $\mu$ g/L);the Initial event was pre-treatment baseline monitoring andthe Final event is the latest post-treatment monitoring event.
	Results range from - 1.00 to + 1.00 A positive result indicates an increase in the ratio. A negative result indicates a decrease in the ratio.
cis-DCE Accumulation Rank	None = $-1.0 - 0.0$ Low = $0.001 - 0.1$ Med = $0.1 - 0.3$ High = $>0.3$
Conc Reduction	The greatest reduction in total chlorinated ethene (CE) concentration for each well during the treatment period prior to rebound. The concentration reduction between the <i>Initial CE</i> and <i>Min CE</i> expressed as a ratio of the <i>Initial CE</i> . Calculated as:
	<ul> <li>(Initial CE – Min CE)/Initial CE</li> <li>Where: Min CE = Minimum total chlorinated ethene (CE) concentration (μg/L) during the treatment period; and Initial CE = Initial total CE concentration (μg/L).</li> </ul>
	Results range from 0.00 to 1.00 (0 to 100% reduction).
Conc Reduction Rank	None = 0.0 Low = 0.01 to median Mid = median to $75^{th}$ percentile High = $< 75^{th}$ percentile

# Table 5-30.Data Dictionary.

Data Field	Definition
Dechlorination	The number of chlorine atoms removed from the chlorinated ethene (CE) molecules as indicated by the change of the molar concentration of chlorine atoms. The greatest reduction is calculated for each well during the treatment period before rebound. The calculation is based on the Chlorine Number for each CE where $PCE = 4$ , $TCE = 3$ , cis-DCE and $tDCE = 2$ , and $VC = 1$ .
	The molar concentration of chlorine atoms [Cl] is calculated as: [Cl] = 4 x (PCE) / 166 + 3 x (TCE) / 131 + 2 x (cis-DCE) / 97 + 2 x (tDCE) / 97 + 1 x (VC) / 62
	Where: molecular weights for individual CE's are: PCE = 166, TCE = 131, cis-DCE = 97, trans-DCE = 97, and VC = 62.
	Brackets indicate molar concentration in micromoles per liter and parentheses indicate mass concentration in ug/L.
	Dechlorination is calculated as: (Initial [Cl] – Min [Cl])/ Initial [Cl] A positive result indicates a decrease in [Cl]. Results range from 0.00 to 1.00
Dechlorination Rank	None = 0.0 Low = 0.01 to median Mid = median to $75^{th}$ percentile High = $< 75^{th}$ percentile
DNAPL	A Yes/No indication of whether DNAPL is present in the treatment area during the treatment period. Total chlorinated ethene (CE) concentrations greater than 10,000 $\mu$ g/L are used as a presumptive indication of the presence of DNAPL.
	Calculated as:: IF Initial CE >10,000, THEN "Y", ELSE "N"
	Where: CE = the sum of the concentrations of PCE, TCE, cis-DCE, tDCE, and VC for each individual sampling event at a specific well; and
	The Initial event was pre-treatment baseline monitoring.
	Calculation output: Y = yes N = no
Dosing	The concentration of the substrate injection fluid. The substrate product diluted in water as a percent by volume.
Final CE	Total chlorinated ethene (CE) concentration for each well at the Final post-treatment monitoring event ( $\mu$ g/L), where the Final event is the latest post-treatment monitoring event.

Data Field	Definition
Fines	The percentage of fines (silt plus clay) within the Treatment Zone based on representative soil boring logs for each site.
	ASTM D2488 classifies soils based on the percentage of fines, sand, gravel, and organic matter. Each soil type contains by definition a percentage range of fines. The midpoint of the range was selected as the average. For example, silty sand contains by definition between 15% and 50% silt by definition with a midpoint mean value of 32.5%.
	Table 5-26 provides a listing if USCS soil names, the percentage range of fines for each soil classification, and the mean percentage of fines.
	Representative soil boring logs were evaluated to determine the total percent fines using a weighted average of the percent fines for each distinct soil layer.
	Representative boring logs are provided in Appendix F.
Fines Rank	$\begin{array}{ll} H &> 0.50 \\ M &> 0.15 \mbox{ and } < 0.50 \\ L &< 0.15 \end{array}$
Gradient	The typical hydraulic gradient (ft/foot) at the site estimated from water level data reported in site documents.
	<i>Gradient</i> was determined for each site by examining a typical water level contour map and measuring the change in water level elevation or hydraulic head along a typical flow path.
	Calculated as: dh/dl
	Where: dh = change in hydraulic head and dl = horizontal length along flow path.
Heterogeneity	A measure of the number of distinct soil layers within the saturated thickness of the treatment zone with units of layers per 10 ft.
	A representative soil boring log was evaluated for each site to determine the number of distinct layers based on the Unified Soil Classification System Standard Practice for Description and Identification of Soils Visual-Manual Procedures (ASTM D2488).
	Calculated as: Number of Layers/ (Measured Thickness x 10).
	Calculations for <i>Heterogeneity</i> for the individual sites are provided in <b>Appendix E</b> .
Heterogeneity Rank	$\begin{array}{ll} H &> 5\\ M &> 2 \text{ and } < 5\\ L &< 2 \end{array}$
Hydraulic Conductivity	The average hydraulic conductivity (ft/day) of the aquifer within the treatment zone as reported in site documents.
Initial	Initial refers to the most recent pre-treatment monitoring event.
Initial Alk	Initial alkalinity concentration (mg/L as CaCO <sub>3</sub> )

Data Field	Definition
Initial CE	Initial total chlorinated ethene (CE) concentration (μg/L). Calculated as: <i>Initial CE</i> = Initial PCE + Initial TCE + Initial cis-DCE + Initial trans-DCE + Initial VC Where: PCE = tetrachloroethene TCE = trichloroethene cis-DCE = cis-1,2-dichlorethene trans-DCE = trans-1,2-dichlorethene VC = vinyl chloride
Initial DHC	Initial Dehalococcoides concentration (copies/L)
Initial DO	Initial dissolved oxygen concentration (mg/L)
Initial Fe	Initial iron concentration (mg/L)
Initial Methane	Initial methane concentration ( $\mu$ g/L)
Initial ORP	Initial oxidation-reduction potential (millivolts)
Initial PH	Initial pH value (standard pH units)
Initial Sulfide	Initial sulfide concentration (mg/L)
Initial Sulfate	Initial sulfate concentration (mg/L)
Initial TOC	Initial total organic carbon concentration (mg/L)
Loading Ratio	The Actual Loading of substrate volume compared to the Recommended Loading. Calculated as: Actual Loading/Recommended Loading Where: Actual Loading is the volume of substrate product added during the treatment period as reported in site documents. Recommended Loading is the volume of substrate product calculated using the Substrate Loading Tool developed for ESTCP Project ER-0627 "Loading Ratios and Impacts of Substrate Delivery for Enhanced Anaerobic Bioremediation" (Parsons, 2010).
	Inputs to the Substrate Loading Tool include the following: Treatment Zone – length, width, and thickness; Hydrogeologic – aquifer hydraulic conductivity, hydraulic gradient, and porosity; Geochemical – competing electron acceptors, DO, sulfate, and nitrate; Contaminant concentrations – average initial concentrations of PCE, TCE, cis-DCE, trans-DCE, and VC. The calculations employed in the Substrate Loading Tool use a hydrogen equivalent approach to calculate the substrate demand. A safety factor of 3 was used.
	The Substrate Loading Tool calculations are provided in Appendix G.

Data Field	Definition
Max	<i>Max</i> refers to the maximum (greatest) concentration observed during the treatment period extending from the <i>Initial</i> to <i>Final</i> monitoring events for each individual monitoring well.
Max Alk	Maximum alkalinity concentration (mg/L as CaCO <sub>3</sub> ).
Max CE	Maximum chlorinated ethenes (CE) = the sum of the concentrations of PCE, TCE, cis-DCE, trans-DCE, and VC for each individual sampling event at a specific well.
Max DHC	Maximum Dehalococcoides concentration (copies/L) observed during the treatment period.
Max DO	Maximum dissolved oxygen concentration (mg/L) observed during the treatment period.
Max Fe	Maximum iron concentration (mg/L) observed during the treatment period.
Max Methane	Maximum methane concentration ( $\mu$ g/L) observed during the treatment period.
Max ORP	Maximum oxidation-reduction potential (millivolts) observed during the treatment period.
Max PH	Maximum pH value (standard pH units) observed during the treatment period.
Max Sulfate	Maximum sulfate concentration (mg/L) observed during the treatment period.
Max TOC	Maximum total organic carbon concentration (mg/L) observed during the treatment period.
MCL Attained	A Yes/No indication of whether the Maximum Contaminant Level (MCL) was attained during the treatment period at each specific well.
	Calculated as: IF Final PCE < MCL AND Final TCE < MCL AND Final tDCE < MCL AND Final VC < MCL, THEN "Y", ELSE "N"
	Where: The Final event is the latest post-treatment monitoring event.
Method of	The method substrate was introduced into groundwater, including:
Injection	Recirc = recirculation with a combination of extraction and injection wells;
	IW = injection wells without active recirculation; and
	DPT = direct-push technology involving insertion of rods into the aquifer, injection of substrate through the rods followed by removal of the rods resulting in no permanent injection wells.
Method of Injection Cat	General categories: Direct or Recirc
Min	<i>Min</i> refers to the minimum (lowest) concentration observed during the treatment period extending from the <i>Initial</i> to <i>Final</i> monitoring events for each individual monitoring well.
Min Alk	Minimum alkalinity concentration (mg/L as CaCO <sub>3</sub> ) observed during the treatment period.
Min CE	Minimum total chlorinated ethene (CE) concentration ( $\mu$ g/L) during the treatment period.
Min DHC	Minimum Dehalococcoides concentration (copies/L) observed during the treatment period.

Data Field	Definition
Min DO	Minimum dissolved oxygen concentration (mg/L) observed during the treatment period.
Min Fe	Minimum iron concentration (mg/L) observed during the treatment period.
Min Methane	Minimum methane concentration ( $\mu$ g/L) observed during the treatment period.
Min ORP	Minimum oxidation-reduction potential (millivolts) observed during the treatment period.
Min PH	Minimum pH value (standard pH units) observed during the treatment period.
Min S	Minimum sulfide concentration (mg/L) observed during the treatment period.
Min Sulfate	Minimum sulfate concentration (mg/L) observed during the treatment period.
Min TOC	Minimum total organic carbon concentration (mg/L) observed during the treatment.
Nutrients	Yes/No indication of whether nutrients were added to the substrate injection. TRUE = yes; FALSE = no.
Previous Remediation	Remediation conducted in the treatment area before bioremediation. None Bio = bioremediation Bioventing Excavation ERH = electrical resistive heating SPH = six-phase heating ISCO = in situ chemical oxidation P&T = pump and treat SPH and ISCO = six phase heating combined with in situ chemical oxidation ZVI = zero valent iron
Previous Remediation Rank	Y = Yes $N = No$
Rebound	The difference between the Initial Decrease in concentration and Subsequent Increase in concentration expressed as a ratio of the Initial Decrease. If the Final CE concentration is greater than the Initial CE concentration, then the Rebound is set to 1.00. Calculated as: If (Final CE > Initial CE), then1.00, else ((Initial Decrease – Subsequent Increase)/Initial Decrease) Where: Initial Decrease = Initial CE – Min CE and Subsequent Increase = Final CE – Min CE.
Rebound Rank	None = 0.0
	Low= 0.01 – 0.11 (median)
	Med = 0.11 - 0.99
	High= 1.00

Data Field	Definition
Seepage Velocity	The groundwater seepage velocity (ft/year). Calculated as: ( <i>Hydraulic Conductivity</i> x <i>Gradient</i> )/ (Effective Porosity x 365.25) Where: Effective Porosity is assumed to be 0.20 and days per year = 365.25.
Seepage Velocity Rank	H > 100 M > 20 and < 100 L < 20
Spacing	The horizontal distance in ft between injection points.
Substrate	The organic carbon substrate or mixture of substrates used in the Treatment. Substrates used at the various sites included: Lactate = sodium lactate EVO = emulsified vegetable oil ESO = emulsified soybean oil EOS = emulsified oil substrate HRC = hydrogen releasing compound Lactic Acid Lactoil <sup>TM</sup> = sodium lactate and emulsified oil mixture Molasses ZVI = zero valent iron
Substrate Rank	H = high viscosity substrates including molasses, EOS, EVO, Lactoil <sup>TM</sup> and HRC $L$ = low viscosity substrates such as sodium lactate, lactic acid, and hydrogen
Sulfate Depletion	The greatest reduction in sulfate concentration for each well during the treatment period prior to rebound. The calculation is the reduction between <i>Initial Sulfate</i> and <i>Min Sulfate</i> expressed as a ratio of the <i>Initial Sulfate</i> . If the <i>Min Sulfate</i> is greater than the <i>Initial Sulfate</i> , then a value of zero is assigned (no depletion). Calculated as:
	<pre>If Min Sulfate &gt;= Initial Sulfate, then 0, else (Initial Sulfate – Min Sulfate)/Initial Sulfate Where:     Min Sulfate = Minimum sulfate concentration (μg/L) during the treatment period; and     Initial Sulfate = Initial sulfate concentration (μg/L); Results range from 0.00 to 1.00 (0 to 100% depletion).</pre>
Тетр	The average annual air temperature (°C) for the Installation as reported by NOAA for the closest city to the Installation. The average annual air temperature was assumed to approximate the average annual groundwater temperature.
Treatment ID	Identification number unique to a specific Installation, Site, Area, Depth Zone, and Phase of treatment operations.

Data Field	Definition
VC Accumulation	The difference between the Initial VC and Final VC concentrations at each individual well as ratios to the Initial CE and Final CE concentrations, respectively.
	Calculated as: (Final VC/ Final CE) – (Initial VC/ Initial CE)
	<pre>Where: Final VC = Final vinyl chloride concentration (μg/L); Final CE = Final total CE concentration (μg/L); Initial VC = Initial vinyl chloride concentration (μg/L); Initial CE = Initial total CE concentration (μg/L);</pre>
	and: the Initial event was pre-treatment baseline monitoring, and the Final event is the latest post-treatment monitoring event.
	Results range from - 1.00 to + 1.00 A positive result indicates an increase in the ratio. A negative result indicates a decrease in the ratio.
VC Accumulation Rank	None = $-1.0 - 0.0$ Low = $0.001 - 0.10$ Mid = $0.10 - 0.4$ High = $>0.4$

Table 3-51. USOS Son Classification with reference rines.	Table 5-31.	<b>USGS Soil</b>	Classification	with	<b>Percent Fines.</b>
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USCS Soil Name	Minimum % Fines	Maximum % Fines	Mean % Fines
Clay	85	100	92.5
Clay with sand	75	85	80
Clay with gravel	75	85	80
Sandy clay	50	70	60
Sandy clay with gravel	50	70	60
Gravelly clay	50	70	60
Gravelly clay with sand	50	70	60
Silt	85	100	92.5
Silt with sand	75	85	80
Silt with gravel	75	85	80
Sandy silt	50	70	60
Sandy silt with gravel	50	70	60
Gravelly silt	50	70	60
Gravelly silt with sand	50	70	60
Organic soil	85	100	92.5
Organic soil with sand	75	85	80
Organic soil with gravel	75	85	80
Sandy organic soil	50	70	60
Sandy organic soil with gravel	50	70	60
Gravelly organic soil	50	70	60
Gravelly organic soil with sand	50	70	60

# 6.0 PERFORMANCE ASSESSMENT

The information compiled in the database (version Matrix Bio 21) is used here to perform a statistical evaluation of the correlations observed between the effectiveness measures versus site characteristics, and effectiveness measures versus remediation methods (**Table 5-25**). Such correlations, if statistically significant, may indicate specific site characteristics or remediation methods that help or hinder successful remediation. Parameters that affect measures of rebound are also evaluated here. Definitions of the parameters in the database are provided in **Section 5.4**. The methodology and results of the performance assessment are provided in the following sections.

## 6.1 STATISTICAL METHOD

Correlations between selected pairs of numerical parameters were evaluated using the Spearman Rank Order (SRO) procedure. This procedure calculates a Spearman correlation coefficient R (rho), which is a nonparametric measure of statistical dependence between two variables. It assesses how well the relationship between two variables can be described using a monotonic function. The coefficient R ranges from -1.0 to +1.0. The value of R is positive if Y tends to increase as X increases, and the value is negative if Y tends to decrease when X increases. A value of zero indicates that there is no tendency for Y to either increase or decrease when X increases. A Spearman correlation coefficient of +1.0 or -1.0 occurs when each of the variables is a perfect monotonic function of the other.

The SRO procedure also returns a p-value, which allows the results to be interpreted in terms of the statistical confidence in any apparent correlation. The correlations are conservatively evaluated with the null hypothesis that no correlation exists. The p-value, which ranges from 0.0 to 1.0, can be thought of as the probability of the null hypothesis (no correlation) being true. For this report, the test is evaluated at a 95 percent confidence level, at which there is only a five percent chance of incorrectly accepting the null hypothesis (no correlation) when a true correlation actually exists. At a confidence level of 95 percent, a correlation is considered to have statistical significance if the p-value is equal to or less than 0.05.

The SRO correlation method is described as being "nonparametric," meaning that a perfect SRO correlation results, when X and Y are related by any monotonic function. This is different from a linear correlation (such as Pearson product-moment), which only gives a high R when X and Y are related by a linear function. The nonparametric SRO approach was selected here because there is no expectation that any of the parameters considered in this evaluation will be linearly related to each other, but at least some monotonic relationships are expected.

Many of the parameters in the database are categorical (non-numeric), so their effects on performance cannot be evaluated with the SRO method. For instance, the *DNAPL Y/N*, and *Bioaug Rank* parameters contain either a Y (yes) or N (no) value for each case. Likewise, the *Seepage Velocity Rank*, *Heterogeneity Rank*, and *Fines Rank* parameters contain either an H (high), M (medium), or L (low) value for each case. The relationships between these parameters versus performance parameters are evaluated using box plots for visual comparisons of the distributions within each individual category versus a performance parameter, and the quantitative Wilcoxon Rank Sum (WRS) test or the Kruskal-Wallis (KW) test. The KW test is a multiple sample

comparison test that is equivalent to the two-sample WRS test. Both are nonparametric procedures that test the null hypothesis that the sub-groups of results that are being compared are drawn from the same population. Both tests return a *p*-value that indicates the probability of the null hypothesis (no difference between sub-groups) being true. The KW test also returns *p*-values for the individual sub-groups, which indicate which sub-group is the most different from the others.

The software package *Statistica* (version 12) was used to evaluate the correlations between selected parameters in the database using SRO correlations, KW tests, and WRS tests; and was also used to generate the box plots.

# 6.2 STATISTICAL EVALUATION RESULTS

The key parameters in the database are grouped as performance measures, site characteristics, and remediation methods. An Excel spreadsheet that includes the entire matrix accompanies this report. Evaluations of SRO correlations between performance measures versus site characteristics and performance measures versus remediation methods were performed. Identification of positive or negative correlations that are statistically significant may indicate possible factors that affect performance. A table of the SRO correlation coefficients between all pairs of numerical parameters is provided in **Table 6-1**. Statistically significant correlations ( $\alpha$ = 0.05) are shown in red. Non-numeric (categorical) parameters were visually compared to performance measures using box plots and quantitatively compared using either the KW test or the WRS test.

Table 6-1. SRO Results for all Pairs of Numerical Parameters SRO Correlation Coefficients.

Variable	Initial CE	Min CE	Max CE	Final CE	DNAPL 1/0	Conc Reductio n	Dechlor- ination	Rebound	cDCE Accum	VC Accum	Initial pH	Max pH	Initial DO	Min DO	Initial ORP	Min ORP	Intitial SO4	Min Sulfate	Sulfate Depletion	Initial TOC	Max TOC	Initial Sulfide	Max Sulfide	Initial Fe	Max Fe	Intial DHC	Max DHC	Max Alkalinity	Max Methane	Temp	Seepage Velocity	Spacing	Dosing	Heteroge neity	Fines
Initial CE	1	0.523	0.853	0.477	0.648	0.167	0.171	-0.204	-0.095	0.193	0.124	0.088	-0.002	0.031	0.105	-0.016	0.148	0.057	0.080	0.062	0.234	-0.145	-0.232	0.072	0.222	-0.143	-0.192	0.054	0.101	0.013	-0.032	0.079	0.137	-0.086	0.304
Min CE	0.523	1	0.560	0.847	0.272	-0.655	-0.651	0.199	0.087	-0.003	-0.017	-0.055	-0.017	0.151	0.288	0.110	0.169	0.294	-0.247	-0.017	0.001	-0.192	-0.134	0.042	-0.008	-0.257	-0.570	-0.146	-0.285	0.029	-0.110	-0.092	0.220	0.051	0.409
Max CE	0.853	0.560	1	0.626	0.633	-0.030	-0.025	0.054	0.000	0.169	0.092	0.039	0.037	-0.014	0.219	-0.054	0.112	0.067	0.057	0.098	0.252	-0.103	-0.147	0.035	0.185	-0.109	-0.089	0.044	-0.048	-0.043	0.042	0.130	0.048	-0.196	0.319
Final CE	0.477	0.847	0.626	1	0.275	-0.552	-0.548	0.474	0.131	-0.032	-0.021	-0.054	0.043	0.078	0.338	0.094	0.127	0.270	-0.210	0.000	-0.002	-0.094	-0.142	0.019	0.000	-0.171	-0.484	-0.179	-0.386	0.025	-0.069	-0.023	0.199	0.017	0.408
DNAPL 1/0	0.648	0.272	0.633	0.275	1	0.130	0.132	-0.120	-0.014	0.094	0.114	0.023	-0.092	0.057	0.015	-0.018	0.076	0.017	0.007	0.155	0.193	-0.148	-0.183	0.031	0.089	0.120	-0.021	0.063	0.122	0.037	-0.022	0.106	-0.044	-0.108	0.139
Conc Reduction	0.167	-0.655	-0.030	-0.552	0.130	1	0.998	-0.515	-0.148	0.131	0.106	0.103	0.030	-0.115	-0.214	-0.148	-0.170	-0.301	0.264	0.013	0.111	0.111	-0.010	0.071	0.204	0.403	0.282	0.125	0.273	-0.026	0.118	0.116	-0.072	-0.092	-0.176
Dechlorination	0.171	-0.651	-0.025	-0.548	0.132	0.998	1	-0.513	-0.158	0.154	0.101	0.099	0.031	-0.119	-0.204	-0.153	-0.169	-0.304	0.270	0.006	0.108	0.114	-0.018	0.068	0.217	0.403	0.283	0.124	0.275	-0.029	0.124	0.117	-0.071	-0.098	-0.172
Rebound	-0.204	0.199	0.054	0.474	-0.120	-0.515	-0.513	1	0.131	-0.109	-0.025	0.017	0.044	-0.088	0.149	0.031	0.077	0.150	-0.062	-0.014	-0.059	0.112	0.063	-0.124	-0.058	-0.057	0.026	-0.055	-0.177	-0.028	-0.023	0.012	-0.001	0.003	0.107
cDCE Accum	-0.095	0.087	0.000	0.131	-0.014	-0.148	-0.158	0.131	1	-0.343	0.100	0.035	0.062	-0.057	0.151	0.007	0.172	0.208	-0.102	-0.061	0.062	0.217	0.071	-0.188	-0.088	-0.120	0.186	-0.016	-0.349	-0.067	0.048	-0.081	-0.002	-0.155	-0.091
VC Accum	0.193	-0.003	0.169	-0.032	0.094	0.131	0.154	-0.109	-0.343	1	-0.015	0.024	0.064	-0.069	0.025	-0.112	-0.011	-0.126	0.172	-0.141	0.074	-0.050	-0.022	-0.029	0.124	-0.237	0.299	0.063	0.083	-0.018	0.132	0.052	-0.082	-0.120	0.011
Initial pH	0.124	-0.017	0.092	-0.021	0.114	0.106	0.101	-0.025	0.100	-0.015	1	0.629	-0.046	-0.129	-0.291	-0.105	0.617	0.404	-0.023	0.176	0.064	0.365	0.095	-0.162	0.040	0.186	0.105	0.602	0.208	-0.279	-0.173	0.158	0.037	-0.281	-0.035
Max pH	0.088	-0.055	0.039	-0.054	0.023	0.103	0.099	0.017	0.035	0.024	0.629	1	0.044	-0.159	-0.175	-0.181	0.468	0.305	-0.073	0.095	0.121	0.230	0.200	-0.105	0.216	0.121	0.271	0.534	0.310	-0.280	-0.085	0.059	0.186	-0.208	0.122
Initial DO	-0.002	-0.017	0.037	0.043	-0.092	0.030	0.031	0.044	0.062	0.064	-0.046	0.044	1	0.056	0.274	-0.065	-0.070	-0.162	0.068	-0.146	0.004	0.000	0.213	-0.037	-0.179	-0.148	-0.313	-0.017	-0.246	-0.128	0.175	-0.051	-0.088	-0.098	0.000
Min DO	0.031	0.151	-0.014	0.078	0.057	-0.115	-0.119	-0.088	-0.057	-0.069	-0.129	-0.159	0.056	1	0.012	0.422	-0.030	0.049	-0.136	-0.175	-0.004	-0.150	-0.328	0.100	-0.317	0.127	0.187	-0.048	-0.291	0.125	-0.218	-0.056	-0.162	0.298	-0.080
Initial ORP	0.105	0.288	0.219	0.338	0.015	-0.214	-0.204	0.149	0.151	0.025	-0.291	-0.175	0.274	0.012	1	0.184	0.005	0.126	-0.079	-0.404	-0.059	-0.035	0.007	-0.250	0.046	-0.321	-0.311	-0.197	-0.349	-0.165	0.148	0.016	0.129	0.009	0.273
Min ORP	-0.016	0.110	-0.054	0.094	-0.018	-0.148	-0.153	0.031	0.007	-0.112	-0.105	-0.181	-0.065	0.422	0.184	1	-0.133	0.129	-0.304	-0.303	-0.259	-0.192	-0.466	0.133	-0.391	-0.013	-0.450	-0.154	-0.031	-0.035	0.006	0.142	-0.017	0.021	-0.001
Intitial SO4	0.148	0.169	0.112	0.127	0.076	-0.170	-0.169	0.077	0.172	-0.011	0.617	0.468	-0.070	-0.030	0.005	-0.133	1	0.702	0.098	-0.094	-0.020	0.270	0.171	-0.398	-0.031	-0.589	-0.501	0.441	-0.155	-0.211	-0.336	0.081	-0.015	-0.060	-0.091
Min Sulfate	0.057	0.294	0.067	0.270	0.017	-0.301	-0.304	0.150	0.208	-0.126	0.404	0.305	-0.162	0.049	0.126	0.129	0.702	1	-0.455	-0.144	-0.327	0.281	0.026	-0.213	-0.166	-0.650	-0.677	0.127	-0.381	-0.148	-0.213	0.103	0.048	0.097	-0.158
Sulfate Depletion	0.080	-0.247	0.057	-0.210	0.007	0.264	0.270	-0.062	-0.102	0.172	-0.023	-0.073	0.068	-0.136	-0.079	-0.304	0.098	-0.455	1	-0.006	0.343	0.046	0.240	-0.190	0.154	0.194	0.577	0.147	0.173	-0.031	0.153	0.016	-0.049	-0.177	-0.068
Initial TOC	0.062	-0.017	0.098	0.000	0.155	0.013	0.006	-0.014	-0.061	-0.141	0.176	0.095	-0.146	-0.175	-0.404	-0.303	-0.094	-0.144	-0.006	1	0.477	0.069	0.237	0.314	0.293	0.116	-0.556	0.156	-0.154	0.244	-0.286	-0.220	0.076	0.177	0.267
Max TOC	0.234	0.001	0.252	-0.002	0.193	0.111	0.108	-0.059	0.062	0.074	0.064	0.121	0.004	-0.004	-0.059	-0.259	-0.020	-0.327	0.343	0.477	1	0.003	0.450	0.146	0.399		0.179	0.413	0.315	0.135	-0.125	-0.085	0.073	0.085	0.073
Initial Sulfide	-0.145	-0.192	-0.103	-0.094	-0.148	0.111	0.114	0.112	0.217	-0.050	0.365	0.230	0.000	-0.150	-0.035	-0.192	0.270	0.281	0.046	0.069	0.003	1	0.318	-0.427	-0.310	0.400	-0.816	0.111	0.174	0.001	-0.238	-0.021	0.145	0.034	-0.280
Max Sulfide	-0.232	-0.134	-0.147	-0.142	-0.183	-0.010	-0.018	0.063	0.071	-0.022	0.095	0.200	0.213	-0.328	0.007	-0.466	0.171	0.026	0.240	0.237	0.450	0.318	1	-0.118	-0.024	-0.056	0.089	0.205	-0.260	-0.013	-0.156	-0.114	0.154	-0.144	-0.127
Initial Fe	0.072	0.042	0.035	0.019	0.031	0.071	0.068	-0.124	-0.188	-0.029	-0.162	-0.105	-0.037	0.100	-0.250	0.133	-0.398	-0.213	-0.190	0.314	0.146	-0.427	-0.118	1	0.149	-0.949		-0.140	0.135	0.274	-0.112	-0.014	-0.024	0.253	0.151
Max Fe	0.222	-0.008	0.185	0.000	0.089	0.204	0.217	-0.058	-0.088	0.124	0.040	0.216	-0.179	-0.317	0.046	-0.391	-0.031	-0.166	0.154	0.293	0.399	-0.310	-0.024	0.149	1		-0.400	0.183	0.297	-0.115	0.147	-0.180	0.184	0.048	0.203
Intial DHC	-0.143	-0.257	-0.109	-0.171	0.120	0.403	0.403	-0.057	-0.120	-0.237	0.186	0.121	-0.148	0.127	-0.321	-0.013	-0.589	-0.650	0.194	0.116		0.400	-0.056	-0.949		1	0.211	-0.050		-0.478	0.421	-0.375	0.196	-0.421	-0.478
Max DHC	-0.192	-0.570	-0.089	-0.484	-0.021	0.282	0.283	0.026	0.186	0.299	0.105	0.271	-0.313	0.187	-0.311	-0.450	-0.501	-0.677	0.577	-0.556	0.179	-0.816	0.089		-0.400	0.211	1	0.041	-0.632	-0.475	0.613	-0.639	-0.517	-0.545	0.407
Max Alkalinity	0.054	-0.146	0.044	-0.179	0.063	0.125	0.124	-0.055	-0.016	0.063	0.602	0.534	-0.017	-0.048	-0.197	-0.154	0.441	0.127	0.147	0.156	0.413	0.111	0.205	-0.140	0.183	-0.050	0.041	1	0.211	-0.596	0.006	0.491	-0.400	-0.213	-0.397
Max Methane	0.101	-0.285	-0.048	-0.386	0.122	0.273	0.275	-0.177	-0.349	0.083	0.208	0.310	-0.246	-0.291	-0.349	-0.031	-0.155	-0.381	0.173	-0.154	0.315	0.174	-0.260	0.135	0.297		-0.632	0.211	1	0.131	-0.029	0.447	-0.022	-0.317	-0.274
Temp	0.013	0.029	-0.043	0.025	0.037	-0.026	-0.029	-0.028	-0.067	-0.018	-0.279	-0.280	-0.128	0.125	-0.165	-0.035	-0.211	-0.148	-0.031	0.244	0.135	0.001	-0.013	0.274	-0.115	-0.478	-0.475	-0.596	0.131	1	-0.526	-0.349	0.295	0.511	0.085
Seepage Velocity	-0.032	-0.110	0.042	-0.069	-0.022	0.118	0.124	-0.023	0.048	0.132	-0.173	-0.085	0.175	-0.218	0.148	0.006	-0.336	-0.213	0.153	-0.286	-0.125	-0.238	-0.156	-0.112	0.147	0.421	0.613	0.006	-0.029	-0.526	1	0.227	-0.247	-0.598	-0.149
Spacing	0.079	-0.092	0.130	-0.023	0.106	0.116	0.117	0.012	-0.081	0.052	0.158	0.059	-0.051	-0.056	0.016	0.142	0.081	0.103	0.016	-0.220	-0.085	-0.021	-0.114	-0.014	-0.180	-0.375	-0.639	0.491	0.447	-0.349	0.227	1	-0.477	-0.394	-0.349
Dosing	0.137	0.220	0.048	0.199	-0.044	-0.072	-0.071	-0.001	-0.002	-0.082	0.037	0.186	-0.088	-0.162	0.129	-0.017	-0.015	0.048	-0.049	0.076	0.073	0.145	0.154	-0.024	0.184	0.196	-0.517	-0.400	-0.022	0.295	-0.247	-0.477	1	0.129	0.589
Heterogeneity	-0.086	0.051	-0.196	0.017	-0.108	-0.092	-0.098	0.003	-0.155	-0.120	-0.281	-0.208	-0.098	0.298	0.009	0.021	-0.060	0.097	-0.177	0.177	0.085	0.034	-0.144	0.253	0.048	-0.421	-0.545	-0.213	-0.317	0.511	-0.598	-0.394	0.129	1	0.133
Fines	0.304	0.409	0.319	0.408	0.139	-0.176	-0.172	0.107	-0.091	0.011	-0.035	0.122	0.000	-0.080	0.273	-0.001	-0.091	-0.158	-0.068	0.267	0.073	-0.280	-0.127	0.151	0.203	-0.478	0.407	-0.397	-0.274	0.085	-0.149	-0.349	0.589	0.133	1

# 6.2.1 Evaluation of Performance Measures

The first step was to evaluate correlations between the various performance measures themselves to select key performance measures that can be evaluated with respect to site characteristics and remediation methods. Formal database parameter names are shown in italics in the following text.

The *R* values for the SRO correlations between performance measures are provided in **Table 6-2**. All of the correlations considered in the Table are statistically significant at a 95 percent confidence level (shown in red) except for the correlation between *Rebound* versus *VC Accumulation*. Definitions of the parameters in the database are provided in **Section 5.4**.

Variable	Conc	Dechlor-	Rebound	cDCE	VC
valiable	Reduction	ination	Rebound	Accum	Accum
Conc Reduction	1	0.998	-0.515	-0.148	0.131
Dechlorination	0.998	1	-0.513	-0.158	0.154
Rebound	-0.515	-0.513	1	0.131	-0.109
cDCE Accum	-0.148	-0.158	0.131	1	-0.343
VC Accum	0.131	0.154	-0.109	-0.343	1

Table 6-2. SRO Correlation Coefficients for Performance Measures.

It is important to note that the *Concentration Reduction* performance parameter is based on the difference between the initial and minimum chlorinated ethene (CE) concentrations divided by the initial CE concentration and is unaffected by any rebound that may occur after the minimum CE concentrations are reached. Likewise, the *Dechlorination* parameter is based on the extent of dechlorination that occurs between the initial and minimum CE concentration and is also unaffected by any subsequent rebound that may occur after the minimum is reached. The *Rebound* performance parameter is based on the difference between the initial decrease in CE concentration and any subsequent increase in CE concentrations after the minimum is reached.

Concentration Reduction has significant Spearman Rank Order (SRO) correlations with:

**Dechlorination** – A very high positive correlation (R= +0.998) is observed, indicating that *Dechlorination* and *Concentration Reduction* are equivalent measures of CE reduction, but both parameters do not consider any subsequent rebound in concentrations if it occurs.

*Rebound* – A negative correlation (R= –0.515) is observed, suggesting that better concentration reduction is a predictor of less rebound after the minimum CE concentration is reached.

*cis-DCE Accumulation* – A negative correlation (R= –0.148) suggests that better concentration reduction is a predictor of less cis-DCE accumulation after the minimum CE concentration is reached.

*VC Accumulation* – A low but positive correlation (R=+0.131) suggests that large initial reductions in CE concentrations are not a good predictor of VC accumulation after the minimum CE concentrations are reached. In addition, the negative correlation between *VC Accumulation* and *cis*-*DCE Accumulation* suggests that these two parameters may be controlled by different processes. The parameters *cis-DCE Accumulation* and *Rebound* were positively correlated with each other because cis-DCE is often the main contributor to rebound. However, *VC Accumulation* is not significantly correlated with *Rebound*, and is negatively correlated with *cis-DCE Accumulation*, suggesting that factors that affect VC accumulation are different from those affecting rebound. These results show that the *Concentration Reduction*, *Rebound*, and *cis-DCE Accumulation* performance parameters are useful indicators of performance and can be used to determine the effects of site characteristics and remediation methods on treatment system performance.

# 6.2.2 Numerical Site Characteristics

**Table 6-3** provides the SRO correlation coefficients for the 22 numerical site characteristics parameters versus the performance parameter *Concentration Reduction*, and **Figure 6-1** shows these results graphically. The relationships between categorical parameters versus *Concentration Reduction* are shown with box plots and Kruskal Wallis test results. *Concentration Reduction* has statistically significant SRO correlations with the following numerical parameters:

*Initial ORP* – Negatively correlated with *Concentration Reduction*. Low initial ORP is a favorable condition for CE degradation. If ORP is low at the start of treatment, then it is likely that redox conditions are favorable for degradation during the treatment period.

*Minimum ORP* – Negatively correlated with *Concentration Reduction*. Low minimum ORP is a favorable condition for CE degradation. If ORP reached low values during treatment, then CE degradation is maximized. This parameter is also negatively correlated with Sulfate Depletion, indicating that low redox potentials are required for sulfate reduction, which is a key factor in CE degradation.

*Initial Sulfate* – Negatively correlated with *Concentration Reduction*. High initial sulfate interferes with degradation by providing competition for hydrogen between anaerobes that are involved in sulfate reduction and dechlorination. High sulfate also acts as a redox buffer that maintains Eh-pH conditions at or above the sulfate/sulfide equilibrium, which is above the optimal region of Eh-pH space for reductive dechlorination processes.

*Min Sulfate* – Negatively correlated with *Concentration Reduction*. Low or non-detectable sulfate during treatment indicates the establishment of sulfate-reducing conditions under which CE degradation rates are maximized. Persistently high sulfate concentrations can cause a "stall" as the redox conditions are prevented from entering the sulfide Eh-pH stability field, which is a field that is favorable for reductive dechlorination.

**Sulfate Depletion** – Positively correlated with *Concentration Reduction*. Sulfate Depletion is also significantly correlated with *Min CE* (-) and *Final CE* (-). The depletion of sulfate is a key step for the successful degradation of CEs. The inability to reduce sulfate concentrations can be caused by insufficiently reducing conditions, or by a continual source of sulfate, which can occur in coastal sites, arid sites with high-sulfate groundwater, or the presence of gypsum layers in the treatment zone.

Maximum Iron – Positively correlated with Concentration Reduction. High dissolved iron concentration is an indicator of iron-reducing conditions, which is a favorable condition for CE degradation. Max Fe is also significantly correlated with Min DO (-) and Min ORP (-), which are also indicators of favorable conditions for CE degradation.

**Fines** – Negatively correlated with *Concentration Reduction*. This may be due to the diffusion of CEs into and out of the fine-grained layers within the treatment zone, which can limit the effectiveness and/or extend the time required for *in situ* treatment. Fines are also positively correlated with *Min CE* and *Final CE*, also demonstrating that the presence of fines can limit the effectiveness of remediation.

*Max DHC* – This parameter is not correlated with *Concentration Reduction* or any of the rebound parameters, but it is significantly correlated with *Min CE* (-) and *Final CE* (-), indicating that elevated DHC counts contribute to lowered CE concentrations during remediation. In addition, *Max DHC* is significantly correlated with *Min ORP* (-), *Min Sulfate* (-), and *Sulfate Depletion* (+), which are all reliable indicators of performance. It should be noted that only 26 of the 256 cases have values for *Max DHC*, so these results may not be representative of its overall effect.

Table 6-3. SRO Correlation Coefficients for Site Characteristics.

Variable	Conc Reduction	Dechlor- ination	Rebound	cDCE Accum	VC Accum	Initial pH	Max pH	Initial DO	Min DO	Initial ORP	Min ORP	Intitial SO4	Min Sulfate	Sulfate Depletion	Initial TOC	Max TOC	Initial Sulfide	Max Sulfide	Initial Fe	Max Fe	Intial DHC	Max DHC	Max Alkalinit y	Max Methane	Temp	Seepage Velocity	_	Fines
Conc Reduction	1	0.998	-0.515	-0.148	0.131	0.106	0.103	0.030	-0.115	-0.214	-0.148	-0.170	-0.301	0.264	0.013	0.111	0.111	-0.010	0.071	0.204	0.403	0.282	0.125	0.273	-0.026	0.118	-0.092	-0.176
Dechlorination	0.998	1	-0.513	-0.158	0.154	0.101	0.099	0.031	-0.119	-0.204	-0.153	-0.169	-0.304	0.270	0.006	0.108	0.114	-0.018	0.068	0.217	0.403	0.283	0.124	0.275	-0.029	0.124	-0.098	-0.172
Rebound	-0.515	-0.513	1	0.131	-0.109	-0.025	0.017	0.044	-0.088	0.149	0.031	0.077	0.150	-0.062	-0.014	-0.059	0.112	0.063	-0.124	-0.058	-0.057	0.026	-0.055	-0.177	-0.028	-0.023	0.003	0.107
cDCE Accum	-0.148	-0.158	0.131	1	-0.343	0.100	0.035	0.062	-0.057	0.151	0.007	0.172	0.208	-0.102	-0.061	0.062	0.217	0.071	-0.188	-0.088	-0.120	0.186	-0.016	-0.349	-0.067	0.048	-0.155	-0.091
VC Accum	0.131	0.154	-0.109	-0.343	1	-0.015	0.024	0.064	-0.069	0.025	-0.112	-0.011	-0.126	0.172	-0.141	0.074	-0.050	-0.022	-0.029	0.124	-0.237	0.299	0.063	0.083	-0.018	0.132	-0.120	0.011
Initial pH	0.106	0.101	-0.025	0.100	-0.015	1	0.629	-0.046	-0.129	-0.291	-0.105	0.617	0.404	-0.023	0.176	0.064	0.365	0.095	-0.162	0.040	0.186	0.105	0.602	0.208	-0.279	-0.173	-0.281	-0.035
Max pH	0.103	0.099	0.017	0.035	0.024	0.629	1	0.044	-0.159	-0.175	-0.181	0.468	0.305	-0.073	0.095	0.121	0.230	0.200	-0.105	0.216	0.121	0.271	0.534	0.310	-0.280	-0.085	-0.208	0.122
Initial DO	0.030	0.031	0.044	0.062	0.064	-0.046	0.044	1	0.056	0.274	-0.065	-0.070	-0.162	0.068	-0.146	0.004	0.000	0.213	-0.037	-0.179	-0.148	-0.313	-0.017	-0.246	-0.128	0.175	-0.098	0.000
Min DO	-0.115	-0.119	-0.088	-0.057	-0.069	-0.129	-0.159	0.056	1	0.012	0.422	-0.030	0.049	-0.136	-0.175	-0.004	-0.150	-0.328	0.100	-0.317	0.127	0.187	-0.048	-0.291	0.125	-0.218	0.298	-0.080
Initial ORP	-0.214	-0.204	0.149	0.151	0.025	-0.291	-0.175	0.274	0.012	1	0.184	0.005	0.126	-0.079	-0.404	-0.059	-0.035	0.007	-0.250	0.046	-0.321	-0.311	-0.197	-0.349	-0.165	0.148	0.009	0.273
Min ORP	-0.148	-0.153	0.031	0.007	-0.112	-0.105	-0.181	-0.065	0.422	0.184	1	-0.133	0.129	-0.304	-0.303	-0.259	-0.192	-0.466	0.133	-0.391	-0.013	-0.450	-0.154	-0.031	-0.035	0.006	0.021	-0.001
Intitial SO4	-0.170	-0.169	0.077	0.172	-0.011	0.617	0.468	-0.070	-0.030	0.005	-0.133	1	0.702	0.098	-0.094	-0.020	0.270	0.171	-0.398	-0.031	-0.589	-0.501	0.441	-0.155	-0.211	-0.336	-0.060	-0.091
Min Sulfate	-0.301	-0.304	0.150	0.208	-0.126	0.404	0.305	-0.162	0.049	0.126	0.129	0.702	1	-0.455	-0.144	-0.327	0.281	0.026	-0.213	-0.166	-0.650	-0.677	0.127	-0.381	-0.148	-0.213	0.097	-0.158
Sulfate Depletion	0.264	0.270	-0.062	-0.102	0.172	-0.023	-0.073	0.068	-0.136	-0.079	-0.304	0.098	-0.455	1	-0.006	0.343	0.046	0.240	-0.190	0.154	0.194	0.577	0.147	0.173	-0.031	0.153	-0.177	-0.068
Initial TOC	0.013	0.006	-0.014	-0.061	-0.141	0.176	0.095	-0.146	-0.175	-0.404	-0.303	-0.094	-0.144	-0.006	1	0.477	0.069	0.237	0.314	0.293	0.116	-0.556	0.156	-0.154	0.244	-0.286	0.177	0.267
Max TOC	0.111	0.108	-0.059	0.062	0.074	0.064	0.121	0.004	-0.004	-0.059	-0.259	-0.020	-0.327	0.343	0.477	1	0.003	0.450	0.146	0.399		0.179	0.413	0.315	0.135	-0.125	0.085	0.073
Initial Sulfide	0.111	0.114	0.112	0.217	-0.050	0.365	0.230	0.000	-0.150	-0.035	-0.192	0.270	0.281	0.046	0.069	0.003	1	0.318	-0.427	-0.310	0.400	-0.816	0.111	0.174	0.001	-0.238	0.034	-0.280
Max Sulfide	-0.010	-0.018	0.063	0.071	-0.022	0.095	0.200	0.213	-0.328	0.007	-0.466	0.171	0.026	0.240	0.237	0.450	0.318	1	-0.118	-0.024	-0.056	0.089	0.205	-0.260	-0.013	-0.156	-0.144	-0.127
Initial Fe	0.071	0.068	-0.124	-0.188	-0.029	-0.162	-0.105	-0.037	0.100	-0.250	0.133	-0.398	-0.213	-0.190	0.314	0.146	-0.427	-0.118	1	0.149	-0.949		-0.140	0.135	0.274	-0.112	0.253	0.151
Max Fe	0.204	0.217	-0.058	-0.088	0.124	0.040	0.216	-0.179	-0.317	0.046	-0.391	-0.031	-0.166	0.154	0.293	0.399	-0.310	-0.024	0.149	1		-0.400	0.183	0.297	-0.115	0.147	0.048	0.203
Intial DHC	0.403	0.403	-0.057	-0.120	-0.237	0.186	0.121	-0.148	0.127	-0.321	-0.013	-0.589	-0.650	0.194	0.116		0.400	-0.056	-0.949		1	0.211	-0.050		-0.478	0.421	-0.421	-0.478
Max DHC	0.282	0.283	0.026	0.186	0.299	0.105	0.271	-0.313	0.187	-0.311	-0.450	-0.501	-0.677	0.577	-0.556	0.179	-0.816	0.089		-0.400	0.211	1	0.041	-0.632	-0.475	0.613	-0.545	0.407
Max Alkalinity	0.125	0.124	-0.055	-0.016	0.063	0.602	0.534	-0.017	-0.048	-0.197	-0.154	0.441	0.127	0.147	0.156	0.413	0.111	0.205	-0.140	0.183	-0.050	0.041	1	0.211	-0.596	0.006	-0.213	-0.397
Max Methane	0.273	0.275	-0.177	-0.349	0.083	0.208	0.310	-0.246	-0.291	-0.349	-0.031	-0.155	-0.381	0.173	-0.154	0.315	0.174	-0.260	0.135	0.297		-0.632	0.211	1	0.131	-0.029	-0.317	-0.274
Temp	-0.026	-0.029	-0.028	-0.067	-0.018	-0.279	-0.280	-0.128	0.125	-0.165	-0.035	-0.211	-0.148	-0.031	0.244	0.135	0.001	-0.013	0.274	-0.115	-0.478	-0.475	-0.596	0.131	1	-0.526	0.511	0.085
Seepage Velocity	0.118	0.124	-0.023	0.048	0.132	-0.173	-0.085	0.175	-0.218	0.148	0.006	-0.336	-0.213	0.153	-0.286	-0.125	-0.238	-0.156	-0.112	0.147	0.421	0.613	0.006	-0.029	-0.526	1	-0.598	-0.149
Heterogeneity	-0.092	-0.098	0.003	-0.155	-0.120	-0.281	-0.208	-0.098	0.298	0.009	0.021	-0.060	0.097	-0.177	0.177	0.085	0.034	-0.144	0.253	0.048	-0.421	-0.545	-0.213	-0.317	0.511	-0.598	1	0.133
Fines	-0.176	-0.172	0.107	-0.091	0.011	-0.035	0.122	0.000	-0.080	0.273	-0.001	-0.091	-0.158	-0.068	0.267	0.073	-0.280	-0.127	0.151	0.203	-0.478	0.407	-0.397	-0.274	0.085	-0.149	0.133	1

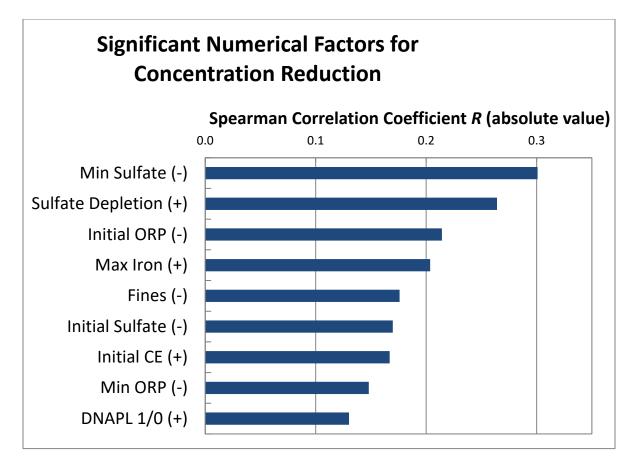


Figure 6-1. Significant Numerical Factors for Predicting Concentration Reduction.

## 6.2.3 Categorical Site Characteristics Parameters

**Presence of DNAPL** – Cases where DNAPL was present were assigned a "Yes" and sites without DNAPL were assigned a "No" for the parameter DNAPL Y/N. **Figure 6-2** shows a pair of box plots that compare the distributions of Concentration Reduction for the two categories of cases. These plots provide a summary view of the entire data set, including the overall location and degree of symmetry. The box encloses the central 50 percent of the data points so that the top of the box represents the 75<sup>th</sup> percentile, and the bottom of the box represents the 25<sup>th</sup> percentile. The median of the data set is represented by a small box within the larger box. The upper whisker extends outward from the box to the maximum point, and the lower whisker extends to the minimum point. **Figure 6-2** shows similar distribution ranges for the two categories, but the median and the 25<sup>th</sup> and 75<sup>th</sup> percentiles of the distribution with DNAPL (n= 43) are shifted higher relative to the distribution without DNAPL (n= 213).

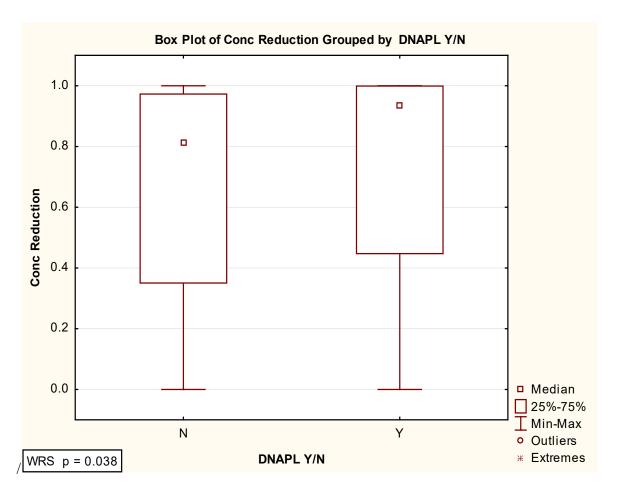


Figure 6-2. Box Plots of Concentration Reduction by the Presence or Absence of DNAPL.

Cases with and without DNAPL (parameter *DNAPL Y/N*) were also compared with respect to *Concentration Reduction* using the Wilcoxon rank sum test. The test result *p*-level, shown on the lower left corner of the Figure of 0.038 is less than the critical value of 0.05, indicating that there is a statistically significant difference between the two categories with respect to *Concentration Reduction*. The upward shift in the cases with DNAPL is likely due to the fact that sites with DNAPL usually have higher initial CE concentrations relative to sites without DNAPL, so the sites with DNAPL tend to show a greater decrease during remediation.

The parameter *DNAPL 1/0* assigns a "1" if DNAPL is present and a "0" if it is absent. This allows the SRO procedure to be used to evaluate correlations with other numerical parameters. *DNAPL* 1/0 has the highest significant positive SRO correlations with *Initial CE* and *Max CE* (**Table 6-2**), supporting the theory that cases where DNAPL is present have higher initial concentrations that allow greater decreases during remediation. These results also suggest that the presence of DNAPL at a site does not interfere with reductions in CE concentrations during remediation. The *DNAPL 1/0* parameter is not correlated with *Rebound*, *cis-DCE Accumulation*, or *VC Accumulation*, so its presence is not a predictor of rebound, which is somewhat surprising. **Fines Rank** – Cases were assigned a high (H), medium (M), or low (L) based on the percent fines (silt plus clay) present in the treated zone. **Figure 6-3** shows box plot comparisons of these three categories of cases as a function of *Concentration Reduction*. The box plots show a clear progression of medians, with the low fines cases (n=102) performing better than the medium fines cases (n=115), which in turn, perform better than the high fines cases (n=39). The KW test plevel of 0.035 indicates that there is a statistically significant difference between one or more of the three categories. High fines in the treatment zone can absorb CE and then slowly release it, and it can also limit the ability to adequately distribute reagents in the subsurface.

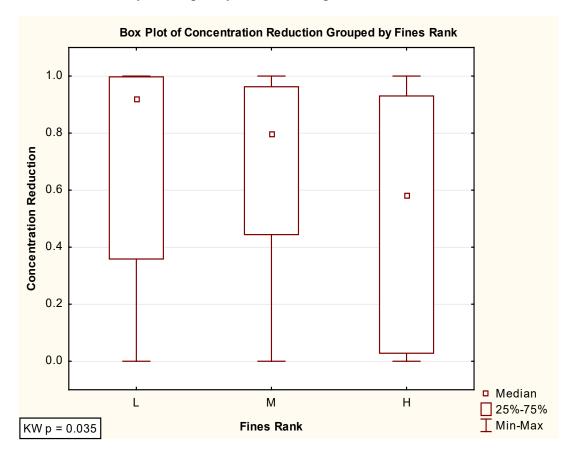


Figure 6-3. Box Plots of Concentration Reduction by Fines Rank.

**Previous Remediation Rank** – Cases were assigned a Yes or No, depending on whether remediation had previously been performed at that location. A box plot comparing these two categories is shown in **Figure 6-4.** The Yes category (n=191) has a higher median and percentiles relative to the No category (n= 57). The very low WRS test *p*-level (0.0008) indicates a statistically significant difference between the categories. A history of previous remediation at a site is a strong predictor of successful CE concentration reduction.

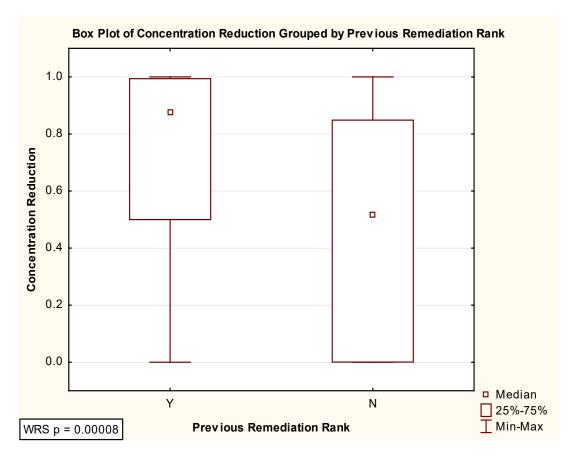


Figure 6-4. Box Plots of Concentration Reduction by Previous Remediation Rank.

*Initial pH* – It is recognized that the groundwater pH needs to be in the circum-neutral range for optimal degradation. Sites with pH conditions that are either too high or too low may, therefore, have lower amounts of CE reduction. The relationship between *Concentration Reduction* versus a broad range of pH values is thus expected to be a peaking function rather than a monotonic function, so a SRO correlation approach would not be appropriate. To address the dependence of *Concentration Reduction* on pH, the initial pH conditions at 229 cases where results are available were divided into four categories (<5, 5 to 6, 6 to 7, and >7). The *Concentration Reduction* results for these four categories were then compared using box plots (**Figure 6-5**) and the KW test. The KW *p*-value of 0.042 indicates that one or more of the four groups of pH ranges are likely drawn from a different distribution of *Concentration Reduction* values. The box plot shows a gradual increase in the *Concentration Reduction* medians as the pH increases. The lowest median, 25<sup>th</sup>, and 75<sup>th</sup> percentiles are observed in the lowest pH (<5.0) category. The 23 pH values in this group range from 3.67 to 4.96, with a mean of 4.45. These cases are from five different installations, so the results do not appear to be driven by location. Most of the cases in the low pH category likely had inhibited the degradation of CEs.

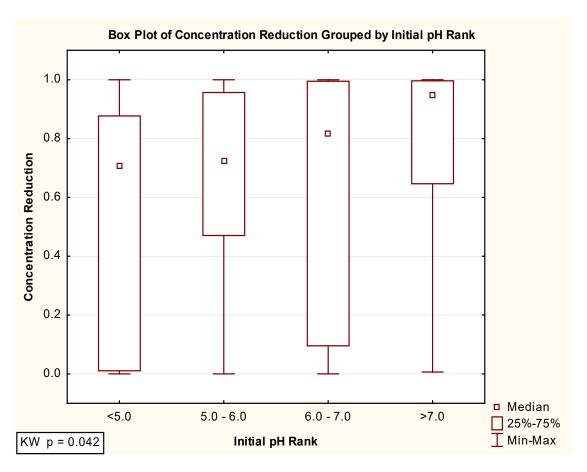


Figure 6-5. Box Plots of Performance by Initial pH.

## 6.2.4 Remediation Methods

**Table 6-4** provides the SRO correlation coefficients for the two numerical site characteristics parameters *Spacing* and *Dosing* versus the performance parameter *Concentration Reduction*. Neither of these parameters is significantly correlated with *Concentration Reduction*. The relationships between categorical remediation methods parameters versus *Concentration Reduction* are shown with box plots and WRS or KW test results and are discussed below.

**Bioaug Rank** – Bioaugmentation methods involve the injection of specific strains or consortia of live microbes such as KB-1 or SDC-9. Cases where bioaugmentation was used (n= 122) were compared with cases where it was not used (n= 134) in **Figure 6-6**. The Figure shows little difference in *Concentration Reduction* between the two groups, and the WRS *p*-value of 0.467 confirms that there is no significant difference. However, a likely confounding factor is that bioassessments (e.g., qPCR and/or treatability studies) may have been performed at some of these sites, resulting in the use of bioaugmentation primarily at sites where it was considered to be needed due to the lack of indigenous microbes and/or treatability study evidence that CE degradation was likely to occur via biostimulation alone.

Variable	Conc Reduction	Dechlor- ination	Rebound	cDCE Accum	VC Accum	Spacing	Dosing
Conc Reduction	1	0.998	-0.515	-0.148	0.131	0.116	-0.072
Dechlorination	0.998	1	-0.513	-0.158	0.154	0.117	-0.071
Rebound	-0.515	-0.513	1	0.131	-0.109	0.012	-0.001
cDCE Accum	-0.148	-0.158	0.131	1	-0.343	-0.081	-0.002
VC Accum	0.131	0.154	-0.109	-0.343	1	0.052	-0.082
Spacing	0.116	0.117	0.012	-0.081	0.052	1	-0.477
Dosing	-0.072	-0.071	-0.001	-0.002	-0.082	-0.477	1

Table 6-4. SRO Correlation Coefficients for Remediation Methods.

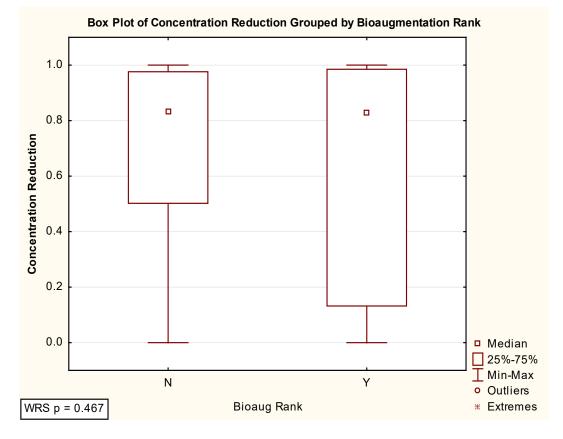


Figure 6-6. Box Plots of Performance by Bioaugmentation Rank.

**Substrate Rank** – This parameter ranks high viscosity carbon substrates, including molasses, EOS, EVO, Lactoil<sup>TM</sup>, and HRC as "H" (n= 148), and low viscosity substrates such as sodium lactate, lactic acid, and hydrogen as "L" (n= 106). The box plot comparison (**Figure 6-7**) shows very similar distributions of Concentration Reduction in the two categories, and the WRS *p*-value test result of 0.743 indicates no significant differences between the two categories. These results indicate that the viscosity of the substrate is not a key factor in predicting performance.

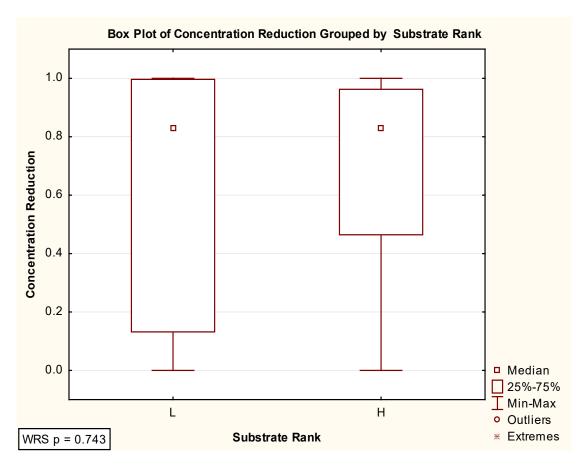


Figure 6-7. Box Plots of Performance by Carbon Substrate Rank.

**Method of Injection Category** – Cases where direct application of amendments via injection wells or direct push (n=170) was used versus recirculation loops (n=86) were compared with respect to *Concentration Reduction*. A box plot comparison of the two groups is shown in **Figure 6-8**. The WRS *p*-value of 0.276 indicates that there are no significant differences in *Concentration Reduction* between the two groups. A confounding factor in this comparison may be that sites with permeable treatment zones do not need recirculation, which is an added expense and is a regulatory issue in some states. If recirculation is only used at sites where it is required for adequate amendment delivery, then these sites would be expected to have similar performance as the sites with more permeable treatment zones where recirculation is not required and was not used.

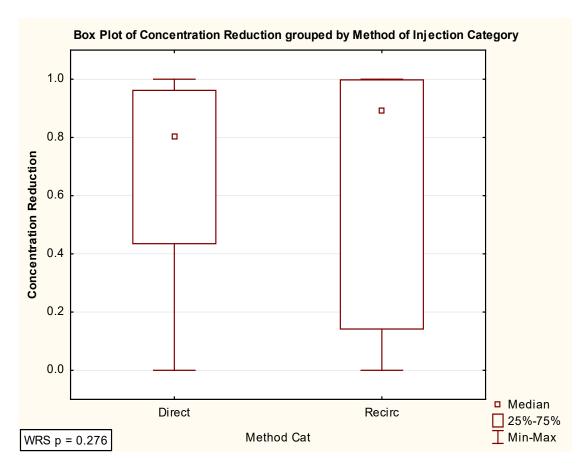


Figure 6-8. Box Plots of Performance by Method of Injection Category.

Use of Nutrients – Sites where inorganic nutrients were used were assigned a "Yes" (n=25), and sites where nutrients were not used were assigned a "No" (n=231) for this parameter. A box plot comparing the distributions of *Concentration Reduction* for each group is provided in Figure 6-9. The plot shows little difference in performance between the two groups, which is confirmed by the high WRS test *p*-value of 0.816. These results suggest that the addition of nutrients has little effect on *Concentration Reduction*. However, as was the case with bioaugmentation, laboratory or field assessments are performed at some sites to determine if nutrients are needed. If nutrients were only used where needed, then these results would be expected.

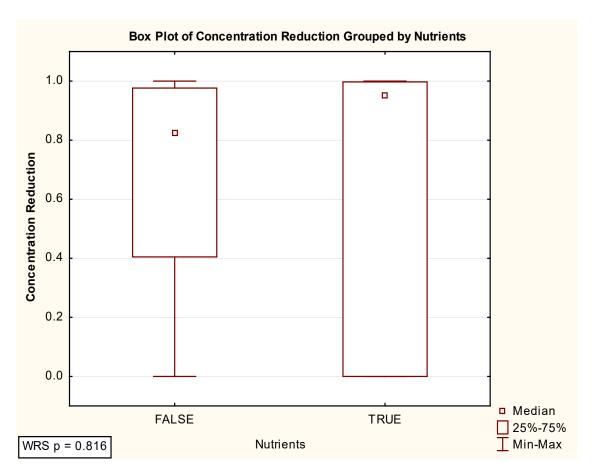


Figure 6-9. Box Plots of Performance by Use of Nutrients.

## 6.2.5 Factors Affecting Rebound, cis-DCE Accumulation, and VC Accumulation

The *Concentration Reduction* and *Dechlorination* parameters are measures of the initial decrease in CE concentrations to the minimum values, irrespective of any subsequent increases. The rebound parameters *Rebound*, *cis-DCE Accumulation*, and *VC Accumulation* are measures of any increases after the minimum CE concentrations are reached during and after remediation activities. Factors that affect these numerical and categorical parameters are discussed in this section.

**Rebound** – This parameter is negatively correlated (R= -0.515) with Concentration Reduction, in part because a greater initial decrease in CE concentrations allows a greater subsequent increase. It is also negatively correlated with Dechlorination, which is an alternate measure of initial concentration reduction. *Rebound* is positively correlated with *cis-DCE Accumulation* because cis-DCE is often a major contributor to rebound.

Numerical site characteristics that have significant correlations with *Rebound* include *Min Sulfate* (+) and *Initial ORP* (+) (**Figure 6-10**). These correlations suggest that high initial ORP is a predictor of rebound. In addition, failure to adequately reduce sulfate to low concentrations in the groundwater may also lead to rebound.

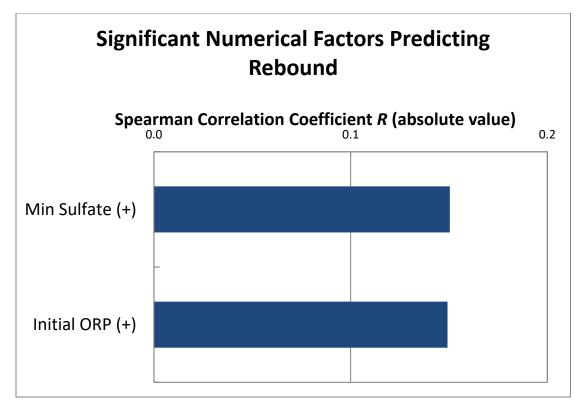


Figure 6-10. Significant Numerical Factors Predicting Rebound.

The proportion of fines in the treated zone may also be an important parameter. For the categorical *Fines Rank* parameter, cases were assigned a high (H), medium (M), or low (L) based on the percent fines (silt plus clay) present in the treatment zone. **Figure 6-11** shows box plot comparisons of these three categories of cases as a function of *Rebound*. The box plots show a clear progression of increasing medians, with the low fines cases (n=102) showing less rebound than the medium fines cases (n=115), which in turn, have less rebound than the high fines cases (n=39). The KW test *p*-level of 0.160 indicates that there is not a statistically significant difference between the three categories. However, the increasing progression of the medians in the figure suggests that a high proportion of fines may be a predictor of *Rebound*. The mass of CE absorbed by the fine layers can then slowly diffuse out of the layers after CE concentrations have decreased during remediation, leading to a reversal of local CE concentration gradients. In addition, a large proportion of fines in the treatment zone can limit the ability to distribute amendments. Note that **Figure 6-11** is a near mirror image of **Figure 6-3**, which shows *Fine Rank* versus *Concentration Reduction*.

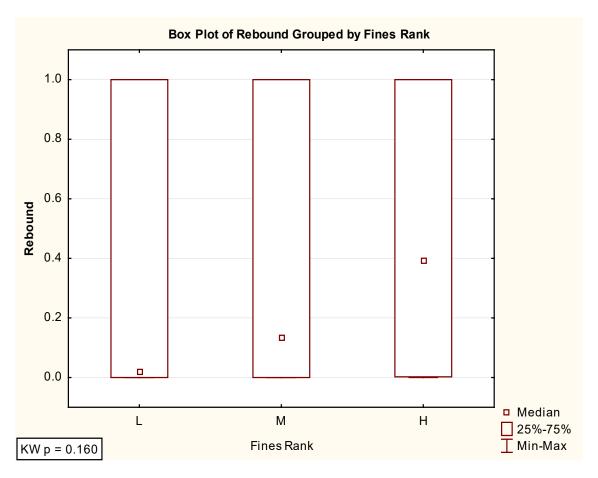


Figure 6-11. Box Plots of Rebound by Fines Rank.

Sites that have undergone previous remediation show significantly less rebound than sites that have not had previous remediation. Figure 6-12 shows *Previous Remediation Rank* versus *Rebound*. Sites without previous remediation (n= 57) have a much higher median rebound than sites with previous remediation (n= 191). The WRS p-value of 0.018 indicates that the differences between the two *Previous Remediation Rank* categories have a statistically significant effect on *Rebound*.

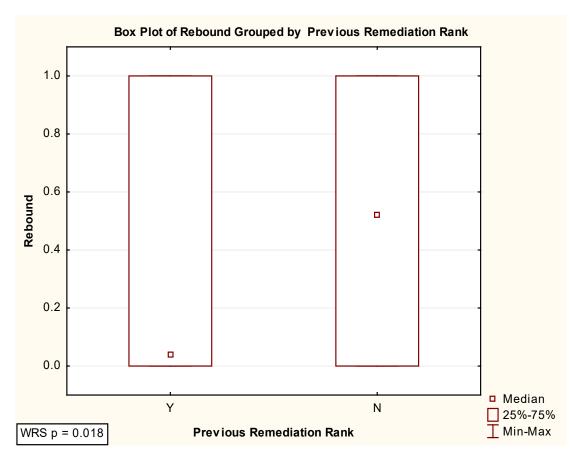


Figure 6-12. Box Plots of Rebound by Previous Remediation Rank.

*cis-DCE* Accumulation – This performance parameter is positively correlated with *Rebound* because cis-DCE is a frequent contributor to rebound. Numerical site characteristics that have significant correlations with *cis-DCE* Accumulation are Minimum Sulfate (+), Initial Fe (-), Initial Sulfate (+), Heterogeneity (-), and Initial ORP (+) in decreasing order (Figure 6-13). These results suggest that high initial redox conditions, as indicated by high initial ORP and low initial iron, are predictors of cis-DCE accumulation.

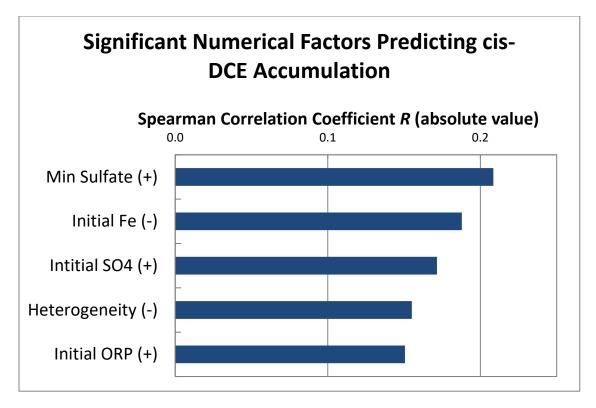


Figure 6-13. Significant Numerical Factors for Predicting cis-DCE Accumulation.

A significant positive correlation with *Min Sulfate* suggests that sulfate concentrations need to be lowered during remediation to avoid accumulation of cis-DCE. This can be a challenge if the site is in a coastal area (such as Seal Beach) or an arid region where naturally high-sulfate groundwater is common. These areas may have on-going sources of sulfate (such as the presence of gypsum layers) that can make it difficult to effectively lower sulfate concentrations, thus leading to cis-DCE accumulation.

A box plot comparison of cases with recirculation loops (n=86) versus direct injection (n=170) with respect to *cis-DCE Accumulation* is provided in **Figure 6-14**. The accumulation of cis-DCE at sites where recirculation loops were used was minor, but less than 25 percent of the sites where direct injection was used experienced a large amount of cis-DCE accumulation. The WRS p-value of 0.022 indicates that these differences are statistically significant. These results may be driven by a small fraction of sites where the permeability of the treatment zone was low and/or the well spacing was high, and recirculation was not used.

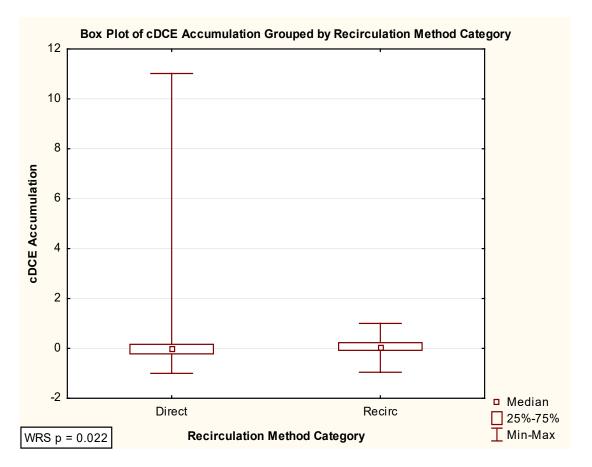


Figure 6-14. Box Plots of cis-DCE Accumulation by Recirculation Method Category.

*Initial pH Rank* also appears to be an important factor at pH conditions below pH 5.0. **Figure 6-15** shows a box plot of *cis-DCE Accumulation* for the four pH ranges. The initial pH does not seem to be a critical factor controlling *cis-DCE Accumulation* except in a small fraction (less than 25 percent) of the cases that have an initial pH below 5.0. These are the only cases of the 229 cases with available initial pH that showed a large degree of *cis-DCE Accumulation*.

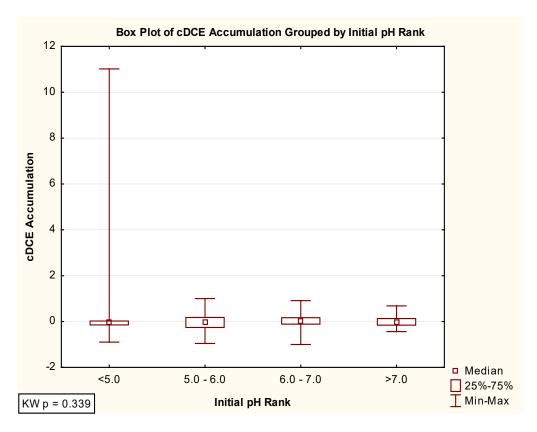


Figure 6-15. Box Plots of cis-DCE Accumulation by Initial pH Rank.

### 6.3 SUMMARY AND CONCLUSIONS

The parameter *Concentration Reduction*, equal to the greatest reduction in total CE concentration for each well during the treatment period before any rebound, was selected as the performance parameter for the majority of the evaluations. Rebound parameters (*Rebound* and *cis-DCE Accumulation*) were also evaluated.

Numerical site characteristics that showed statistically significant ( $\alpha$ = 0.05) positive (+) or negative (-) correlations with *Concentration Reduction* include *Min Sulfate* (-), *Sulfate Depletion* (+), *Initial ORP* (-), *Max Iron* (+), *Fines* (-), *Initial Sulfate* (-), *Initial CE* (+), *Min ORP* (-), and *DNAPL 1/0* (+) in decreasing order (**Figure 6-1**). The strongest correlations were with *Min Sulfate* (-) and *Sulfate Depletion* (+), suggesting that creating strongly reducing redox conditions (that reduce sulfate concentrations as low as possible) will contribute to successful remediation.

The third and fourth most significant correlations were with *Initial ORP* (-) and *Max Iron* (+), demonstrating that either initial or minimum redox potentials are strong predictors of successful CE concentration reductions. These first four strongest correlations provide independent evidence that creating redox conditions corresponding to sulfate-reducing (or lower) redox potentials is beneficial for reducing CE concentrations.

*Initial Sulfate* had a significant negative correlation with *Concentration Reduction*, indicating that high initial sulfate is a predictor of poor performance. High initial sulfate may be present at coastal sites or at arid sites where gypsum layers are present in the treatment zone.

*Fines* have a significant negative correlation with *Concentration Reduction*. This observation suggests that it may be more challenging to lower CE concentrations in treatment zones that have a high proportion of fine-grained sediments. Diffusion of CEs into the fine-grained units over time within the treatment zone may extend the time required for treatment, and may also increase the likelihood of rebound as the CE compounds slowly diffuse back out of the fine-grained units after treatment has stopped. The presence of fine-grained layers can also prevent the adequate distribution of amendments in the subsurface.

Cases where DNAPL is present show significantly greater concentration reduction than cases where DNAPL is not present. This is likely caused by higher initial CE concentrations were DNAPL is present. The significant correlations between *Concentration Reduction* versus *Initial CE* (+) and *DNAPL* 1/0 (+) likely reflect the fact that CE concentrations can only show strong decreases if they are initially high.

Numerical parameters that contribute to *Rebound* include *Min Sulfate* (+) and *Initial ORP* (+), suggesting that high initial ORP is a predictor of rebound. In addition, failure to adequately reduce sulfate to low concentrations in the groundwater during treatment may also lead to rebound.

A positive correlation with the categorical *Fines Rank* parameter indicates that a high proportion of fines in the treatment zone is a predictor of rebound, as well as poor reduction of CE concentrations, as noted above. Also, sites that have undergone previous remediation show significantly less rebound than sites that have not had previous remediation.

Numerical site characteristics that have significant correlations with *cis-DCE Accumulation* are *Minimum Sulfate* (+), *Initial Iron* (-), *Initial Sulfate* (+), Heterogeneity (-), and *Initial ORP* (+) in decreasing order. These results suggest that high initial redox conditions, as indicated by high initial ORP and low initial iron, are predictors of cis-DCE accumulation.

The use of recirculation loops avoids accumulation of cis-DCE, most likely by allowing more effective mixing of amendments in the subsurface. Initial pH below 5.0 is a predictor of cis-DCE accumulation by inhibiting microbial degradation rates, especially the final stages of cis-DCE reduction.

In summary, the key parameters that control CE reduction are similar to the parameters that control rebound. These parameters are:

• **Redox conditions** – The strongest predictors of CE concentration reduction and rebound are *Initial Sulfate*, *Min Sulfate*, and *Sulfate Depletion*. These results demonstrate that it is essential to establish sulfate-reducing (or lower) redox potentials for successful remediation. Correlations of these performance measures with other redox-related parameters such as *Initial ORP*, *Min ORP*, and *Initial Iron* provide independent evidence for these predictors. High initial sulfate or lack of sufficient sulfate depletion may prevent successful remediation if there is an ongoing source of sulfate during the treatment period that prevents redox conditions from falling below sulfate-reducing potentials. Sources of sulfate may include seawater intrusion, arid climate (where sulfate groundwater concentrations are typically high), or the presence of gypsum in the treatment zone.

- **Fines** The presence of fines in the treatment zone is correlated with poor concentration reduction, as well as rebound and DCE accumulation. These results provide independent evidence that the presence of fine-grained layers in the treatment zone interfere with performance, likely by allowing diffusion of CE into the fine layers over time. During remediation, when the CE concentrations decrease in groundwater along interconnected flow paths, the local concentration gradients reverse, and CE that had diffused into the fine layers slowly diffuses back out, thus contributing to rebound. Also, the presence of fine-grained layers can interfere with the distribution of amendments in the subsurface.
- Initial pH Initial pH below 5.0 is a predictor of poor CE concentration reduction and is also a predictor of cis-DCE accumulation. These effects are likely caused by inhibition of microbial activity.
- **Previous remediation** Cases that have undergone previous remediation have significantly higher CE reduction and significantly less rebound and cis-DCE accumulation. These results highlight the need for multiple rounds of remediation at some locations.
- **Presence of DNAPL** One surprising result was that cases where DNAPL was present had significantly greater reductions in CE concentrations than cases where DNAPL was absent. This result is likely due to the fact that sites with DNAPL usually have higher initial CE concentrations relative to sites without DNAPL, so the sites with DNAPL tend to show a greater decrease during remediation. (The presence of DNAPL as the parameter *DNAPL 1/0* is positively correlated with *Initial CE, Min CE, Max CE,* and *Final CE.*) These results also suggest that the presence of DNAPL at a site does not interfere with reductions in CE concentrations during remediation. The degree of rebound and cis-DCE accumulation were insensitive to the presence or absence of DNAPL.

# 7.0 COST ASSESSMENT

The nature of this project does not allow for a traditional ESTCP cost assessment, where one remediation or other technology is evaluated against traditional alternatives. However, this section is intended to provide a reasonable cost estimate for the different assessment technologies utilized during this demonstration. The cost includes labor and per diem for field sampling and PFM installation personnel, materials for sample collection, CSIA, molecular analysis, PFM analysis, and basic geochemistry and VOC analysis. The assessment does not include monies spent on the development of the site database and statistical analysis of the database results.

For the cost assessment, we assume that a total of 8 groundwater wells will be sampled by a single field technician using low-flow sampling and that the technician can sample 4 wells per day. For each well, basic field parameters (dissolved oxygen, oxidation-reduction potential, pH, conductivity) will be determined using a field meter, and samples will subsequently be collected for (1) VOC concentrations (EPA Method 8260); (2) anions (EPA Method 300); (3) C stable isotope analysis of TCE, cis-DCE and VC (4) molecular analysis of important dehalogenating organisms and genes. For the PFM analysis, it is assumed that pricing includes installation, removal, and data analysis on a per PFM basis (Dr. Mike Annable; University of Florida). Shipping of coolers will occur at the end of the second day of sample collection. Based on this scenario, the following cost assumptions were made:

- 1) Rental of required sampling pumps and meters: 1 week (\$400);
- 2) Field labor \$70 per hr x 32 hrs (including travel time) for sampling (\$2,240);
- 3) Vehicle rental: 1 week (\$375);
- 4) Hotel (3 nights) and per diem (2 full days; 2 travel days), Virginia default rate \$96/\$55/\$41 (\$480);
- 5) Shipping samples to laboratories: \$50 ea x 4 (\$200);
- 6) Other miscellaneous materials including ice, coolers, etc (\$300);
- 7) EPA Method 300 anions (5): 8 x \$55 (\$440);
- 8) EPA Method 8260 VOCs: 8 x \$131 (\$1,048);
- 9) C stable isotope analysis for TCE, cis-DCE and VC: 8 x \$600 (\$5,200);
- 10) PFM Analysis (EnviroFlux): 16 (2 PFMs per well) x \$1,700 (\$27,200).

Analytical costs for EPA Method 300 (\$55 per sample for analysis of 5 anions) and EPA Method 8260 (\$131 per sample) represent GSA pricing from a national analytical laboratory. Stable isotope analysis of C and Cl in VOCs is the analytical price provided by a vendor for this project, which for C is \$500 for the first compound in a sample and then \$50 each for every compound after that. For Cl, the price is \$600 for the first compound in a sample and then \$50 each for every compound after that. The price for PFM installation, sampling, and analysis (a service that can be provided commercially by EnviroFlux (Enviroflux.com), was provided by Dr. Mike Annable at the University of Florida, who was a founder of the company. The price assumes 2-5 ft PFMs installed in each well. Based on all assumptions provided above, the estimated cost of sampling and analysis of 8 wells in support of complete biodegradation evaluation of PCE/TCE and daughter products is \$37,883.

# 8.0 IMPLEMENTATION ISSUES

The primary end-users of these technologies (MBT, PFM, CSIA) are expected to be DoD site managers and their contractors, consultants, and engineers. The general concerns of these end users are likely to include the following: (1) technology availability and cost; (2) appropriate application of the technology at DoD sites; and (3) interpretation of CSIA, MBT, and PFM data. These implementation issues are addressed in the following sections. The database developed during this project will also be made available via the ESTCP website.

### 8.1 TECHNOLOGY AVAILABILITY

The C and Cl stable isotope analyses of VOCs described herein as well as the general qPCR analysis of important organisms and genes responsible for VOC biodegradation are commercially available and conducted in multiple university laboratories. Commercial laboratories conducting these analyses include Microbial Insights (Knoxville, TN) and Pace Analytical (Pittsburgh, PA). PFM installation, sampling, and analysis is also commercially available from EnviroFlux (Gainsville, Fl) as previously described in **Section 7**. Thus, the key technologies used in this ESTCP project are commercially available. The database developed during this project will also be made available via the ESTCP website.

### 8.2 APPROPRIATE APPLICATION OF THE TECHNOLOGY AT DOD SITES

Appropriate application of the technologies used in this project will vary by site depending on specific conditions and the questions to be answered. During this project, our primary question was the long-term effectiveness of bioremediation and the different tools were combined toward this end and to assess whether biodegradation was still occurring at select sites a few to several years after active treatment.

## 8.3 INTERPRETATION OF CSIA, MBT AND PFM DATA

## 8.3.1 CSIA

CSIA data gathered on environmental pollutants has been utilized to (1) document biological and abiotic contaminant degradation, (2) estimate or constrain rates of contaminant degradation; (3) identify dominant degradation mechanisms; and (4) forensically determine dominant sources of a specific contaminant in the environment, as well as various other specific applications for individual contaminants. The application and interpretation of CSIA data for the above purposes have been thoroughly reviewed in a US Environmental Protection Agency (EPA) document entitled "A Guide for Assessing Biodegradation and Source Identification of Organic Groundwater Contaminants Using Compound Specific Isotope Analysis (CSIA)" This document is available online through the EPA NEPIS Site (https://nepis.epa.gov/Exe/ZyPDF.cgi/P1002VAI.PDF?Dockey=P1002VAI.PDF). The readers of this ESTCP report are referred to Chapter 4 in this document entitled "Interpretation of Stable Isotope Data from Field Sites," which clearly describes and provides examples of how CSIA data can be utilized to document and quantify the biodegradation of organic contaminants in groundwater aquifers. The Interstate Technology and Regulatory Council (ITRC) Environmental Molecular Diagnostics (EMD) team has also developed online guidance and instruction on CSIA (https://www.itrcweb.org/Team/Public?teamID=3). The C isotope data for TCE, cis-DCE, and VC

gathered during this project provide information on ongoing degradation of these contaminants at 5 different sites and, for the Raritan Site, where samples were collected along a flow path (Section 5.3.1.3.8), an ability to estimate field rates.

# 8.3.2 MBT

The molecular analysis conducted during this project (qPCR of key dehalogenating organisms and genes) is now conducted routinely at VOC sites. Guidance concerning the application of this technique and interpretation of results is available at the ITRC EMD website (https://www.itrcweb.org/Team/Public?teamID=3) and via Microbial Insights, who provide information on the relative abundance of different gene markers at groundwater sites (https://microbe.com/quantarray/).

# 8.3.3 **PFMs**

The PFM analysis conducted was initially developed in 2004 as part of SERDP/ESTCP project ER-200114. Since then, PFM technology has been deployed at over 100+ contaminated sites and is commercially available from companies such as EnviroFlux® (Gainesville, FL, USA) and iFlux® (Boom, Belgium). Additionally, SERDP/ESTCP has funded various iterations of the passive sampling technology, which most recently include a PFM for measuring low partitioning organic contaminants such as 1,4-dioxane (ER-2304) and the colorimetric PFM which use a colorimetric response to assess Darcy fluxes (ER-2420). PFM data interpretation must be carefully performed, as the Darcy velocity varies spatially and temporally, which can have an impact on contaminant flux rates. Contaminant fluxes must be normalized to the groundwater flux in order to assess changes in contaminant mass discharge due to source removal.

Guidance concerning the application of this technology and interpretations is available in a ITRC document entitled "Use and Measurement of Mass Flux and Mass Discharge," which can be found at ITRC EMD website (https://www.itrcweb.org/Guidance/ListDocuments?TopicID=5&SubTopicID=11). Also, guidance document "ER-0144-PR: Demonstration and Validation of Water and Solute Flux Measuring Device" provides additional and more in-depth overview of the PFM technology employed in this project and the guidance document can be found on SERDP/ESTCP website (https://www.serdp-estcp.org/Program-Areas/Environmental-Restoration/Contaminated-Groundwater/Monitoring/ER-200114/ER-200114-TR).

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## **Appendix A: Points of Contact**

## **APPENDIX B: SITE DESCRIPTIONS**

Sites used for statistical data analysis but not described in the final report.

B.1 Former Alameda Naval Air Station, Site 4	2
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## **B.1 Former Alameda Naval Air Station, Site 4**

Former Alameda Naval Air Station is located on the northwestern tip of Alameda Island along the eastern margin of the San Francisco Bay and south of the City of Oakland, California. The northern portion of Alameda Island was formerly tidelands, marshlands, and sloughs adjacent to the historical San Antonio Channel, now known as the Oakland Inner Harbor. Alameda Point was created by filling sub tidal areas, natural tidelands, marshlands, and sloughs with dredge spoils from the surrounding San Francisco Bay, Seaplane Lagoon, and Oakland Inner Harbor.

Site 4 at former Alameda Naval Air Station has been the subject of environmental investigations due to historical documentation of fuel, oil, and solvent usage related to ship and aircraft maintenance. The site includes Building 360, a former Aircraft Engine Facility that contained process shops for blast testing, cleaning, painting, welding, and plating; various aircraft component repair rooms; and nondestructive testing facilities (Arcadis 2014).

Prior to 1975, wastes generated by plating operations in Building 360 were discharged directly to the Seaplane Lagoon via the industrial waste sewer system. Between 1975 to April 1997, plating wastes were treated at an industrial waste treatment facility located north of Building 360. Chemical contaminants from the various industrial processes inside Building 360 are believed to have been released to the subsurface at Site 4 by leaks in the sanitary, industrial waste, and storm sewer lines from the building (Shaw 2013).

Site 4 contains multiple hot sports of chlorinated volatile organic compound (VOC) contamination due to the presence of multiple contamination sources. Three of the hot spots are present and have been referred to as Hot Spot 4-1, Hot Spot 4-2 and Hot Spot 4-3. Two additional hot spots were identified at a utility corridor an at an oil/water separator.

Hot Spot 4-1 and associated Plume 4-1 contains VOCs at concentrations greater than 10,000  $\mu$ g/L and is a presumed dense non-aqueous phase liquid (DNAPL) site. Trichloroethene (TCE) concentrations up to 200,000  $\mu$ g/L were previously obtained from groundwater samples (Shaw 2010). Plume 4-1 extends from Hot Spot 4-1 over 1400 ft and discharges to Seaplane Lagoon in San Francisco Bay. Total VOC concentrations exceed 10,000  $\mu$ g/L throughout much of the plume. Plume 4-1 contains several hot spots of contamination, one of which is identified as Hot Spot 4-1. The predominant groundwater contaminants are TCE and vinyl chloride (VC).

The stratigraphy beneath Plume 4-1 consists of two geologic units: the Merritt Sand Formation and overlying artificial fill (Shaw 2013). The Merritt Sand Formation extends from roughly 10 to 70 ft below ground surface (bgs), and contains fine-grained, silty sand and fine-grained clayey sand. The top of the Merritt Sand is composed of dense, well-consolidated, clayey sand, between 1 and 5 ft thick and has a low hydraulic conductivity. Additionally, a contact zone divides the Merritt Sand into an upper eolian and a lower alluvial section. The eolian section of the Merritt Sand represents sediment deposition by airborne processes, and the alluvial section of the Merritt Sand and alluvial sand sections ranges from 5 to 15 ft thick, consisting of a dense to well-consolidated clayey sand. This clayey sand has a low hydraulic conductivity (Shaw 2013).

Overlying the Merritt Sand Formation is the artificial fill that extends from the surface down to 10 ft bgs. The artificial fill consists of a light to dark brown, fine-grained, silty sand with trace amounts of gravel and brick fragments. The fill is composed of dredge spoils from the San Francisco Bay. Percent fines are estimated to be 66% based on examination of representative soil boing and the treatment zone contains

1.8 layers per 10 ft. Groundwater at Plume 4-1 is encountered between 2 and 8 ft bgs within the artificial fill.

Aquifer pumping tests indicate an average hydraulic conductively of 4.8 ft per day (Arcadis 2014) and hydraulic gradient of 0.005 ft per ft. The groundwater flow directions are affected by local recharge from precipitation, seasonal variation in groundwater elevations, and tidal influences. For Alameda Point, the groundwater has been found to generally flow from southeast to northwest.

Previous pilot studies at Hot Spot 4-1 included bioremediation in 1999 to 2000, low temperature thermal treatment in 2002 (IT 2003) and *in situ* chemical oxidation in 2003 to 2004 (Shaw 2004).

A field-scale bioremediation treatability study was conducted between July 2011 and June 2013. The main objectives of the treatability study were to evaluate the effectiveness of enhanced bioremediation to reduce dense nonaqueous phase liquid in the groundwater within the target treatment zone and to provide a technical basis for expanding bioremediation to the extended plume (Shaw 2013).

The bioremediation treatability study included a pilot recirculation system that was operated for a 60-day period from September to November 2012. The bioremediation system included a groundwater recirculation system with four extraction wells and three injection wells. Each extraction well and injection well operated at a constant rate to distribute sodium lactate substrate across the treatment area. The injected groundwater was bioaugmented with SDC-9<sup>TM</sup> bacteria after reducing conditions were created. Approximately 730 gallons of sodium lactate product was introduced into the injection stream at a dosing concentration of approximately 1.5%.

Multi-level sampling wells and the extraction wells were used to monitor the performance of the system. Baseline sampling was performed during July 2012, prior to system operations to assess the existing groundwater concentrations. Treatment progress sampling occurred regularly over the lactate recirculation and fermentation periods and quarterly thereafter. Post-treatment sampling also was performed to assess contaminated treatment effectiveness and rebound. The final sampling event occurred June 12, 2013.

Multilevel sampling wells and the extraction wells were used to monitor the performance of the system. The wells were closely spaced, with six monitoring wells located within a 10 ft by 10 ft area. Each monitoring location had multiple depth zones (either three or four). The system of multilevel sampling points provides a means to observe the variability of treatment effects on a small scale.

The results of this study show that bioremediation at the site was highly effective. During active treatment and groundwater re-circulation, groundwater concentrations were reduced up to 99 percent in one location and on average by greater than 70 percent across all extraction and monitoring wells. Complete dechlorination of TCE, as evidenced by generation of ethene, was observed in all four extraction wells. Post-treatment monitoring initially showed very little rebound of chlorinated ethenes. While substantial decreases in the chlorinated ethene concentrations were observed, evaluation of the mass balance indicates that only a small fraction (less than 1 percent) of the estimated initial DNAPL mass present (11 to 55 kg) was removed. The range of the initial DNAPL mass present, as well as the uncertainty in correlating DNAPL mass removal to changes in groundwater quality, highlights the challenges in evaluating and treating DNAPL source areas (Shaw 2013).

Based on the results obtained from this pilot study, a full-scale bioremediation approach was selected as a remedy for the site (NAVFAC 2014). The design for the full-scale system included a grid of injection

wells in the source area with a spacing of 15 ft and 20 deep/shallow well couplets (Arcadis 2014). Further information on the effectiveness of the full-scale treatment was not available as of this writing.

## **B.2 Moody Air Force Base, Site FT-07**

Moody Air Force Base (MAFB) is located in central Georgia. FT-07 encompasses approximately 10 acres located in the central portion of MAFB, west of Grand Bay Swamp and east of the Flight line Area (SS-38). The FT-07 site has historically been separated into two areas. Area 1 is located west of the current fire training area and encompasses areas that are east and west of Perimeter Road. Area 2 is located east of the hot cargo pad and encompasses the area east of Perimeter Road. The data collected from Area 2 were used in the statistical analysis described in this report.

The site formerly included seven burn pits along with a building and mock aircraft for fire training exercises. Waste fuel, solvents, paints, and Jet Propellant-4 (JP-4) were used as fuel for fire training purposes. Currently, the majority of the land surface at FT-07 consists of cleared grassy areas and unimproved, piney woods and forested wetlands adjacent to Great Bay Swamp.

The FT-07 site contains groundwater impacted with dissolved-phase chlorinated solvents and hydrocarbon compounds, including 1,2-dichloroethane (DCA), TCE, cis-DCE, and VC. The plume of groundwater contamination at Source 1 extends approximately 600 ft from the source area to the wetland adjacent to Great Bay Swamp. The predominant contaminant is overwhelmingly cis-DCE rather than TCE, and this constituent exceeds 1 mg/L extends throughout the plume. The presence of high cis-DCE concentrations indicates resistance to dechlorinate to VC and then completely to ethene ("DCE Stall"). Concentrations of cis-DCE in excess of 1 mg/L extend over 500 ft from the source area. Source 2, located on the southeastern portion of the site, consisted primarily of cis-DCE contamination at somewhat lower concentrations.

Groundwater at the site has a low pH with pH values ranging between 3.8 and 5.3 prior to treatment, and the low pH is suspected to be the primary factor in the DCE Stall observed. The aquifer contains alternating layers of sand, silt and clay (CDM 2001). The treatment zone contains an estimated 58% fines and is highly heterogeneous with 18 layers per 10 ft. The core of the plume extends into a wetland area with low permeability soils, and the bioremediation injections were within the wetlands at this site.

Three major phases of bioremediation were implemented using three different substrates:

- Hydrogen-Releasing Compound<sup>®</sup> (HRC<sup>®</sup>) in 2002;
- Sodium Lactate in 2005; and
- EOS in 2008.

In each of the treatment phases, amendments were injected via either direct push or temporary well points. Sodium bicarbonate was added in 2008 for pH control.

## HRC Treatment

An interim measure (IM) involving the injection of HRC<sup>®</sup> to serve as an organic carbon electron donor was implemented within the Source Area 2 in November 2002 (Shaw 2003-10). A total of 18,600 pounds of HRC<sup>®</sup> was injected using Direct Push Technology (DPT) during three events (November 2002, November 2003 and March 2004). Injections were conducted in four areas (areas A, B, C, and D) with a total of 167 injection points. Spacing was 10 ft.

## Lactate Treatment

The second phase of bioremediation was implemented in June 2005 (Shaw 2003b). Injections were performed within an expanded area (Locations A and B) using sodium lactate as the primary carbon source to support reductive dechlorination. The injections were performed using a total of 57 injection

wells installed in Locations A and B. Spacing was 20 ft. A total of 1,078 gallons of lactate were injected at a dosing of 1 to 6% (Shaw 2005-11).

Lactate was delivered using a passive gravity-feed method rather than an active injection under pressure, a unique approach among the sites evaluated. Totes were set up at each injection well head and flexible hose extended from the bottom of the tote to the injection well head. The valve was opened at the bottom of the tote and the lactate solution drained by gravity into the well. Bioaugmentation was employed for the first time in this phase with injection of the SDC-9<sup>TM</sup> culture.

#### Emulsified Vegetable Oil (EVO) Treatment

A third phase of treatment was implemented in October 2008 with the installation of 73 additional injection wells in Area 2. Spacing was 40 ft. Injections were initiated in November 2008. A total of 19,800 pounds of sodium bicarbonate, 38,100 pounds of 50 percent sodium hydroxide, and 5,115 gallons of EVO product were injected.

The bioremediation at FT-07 had limited effectiveness. Overall decreasing trends were observed at many wells; however cis-DCE remained resistant to degradation. Additional injections were planned in 2015 but the status of this work is unknown at this writing.

## **B.3 Myrtle Beach Air Force Base, Site FT-11**

The former Myrtle Beach Air Force Base (MBAFB) is located in the Coastal Plain of South Carolina approximately 2.5 miles inland from the Atlantic Ocean near Myrtle Beach, South Carolina. FT-11 at MBAFB is located in the north-central portion of the base, approximately one-half mile west of the flight line. FT-11 was formerly used as a practice area to conduct fire training exercises from 1965 through 1969. The FT-11 area consisted of a 100-ft-diameter earthen burn ring with a low-relief berm. Fire training involved soaking the ground within the berm with jet petroleum grade-4 fuel, waste oil, and/or solvents, igniting the fuel, and then extinguishing the fire (CB&I 2014).

Topography at the FT-11 site is fairly flat and ranges from 18 to 20 ft above mean sea level. Surface water runoff from FT-11 flows east and southwest into two drainage ditches that carry the flow to the storm drainage system. The unconsolidated sediments underlying FT-11 consist of fine to medium-grained sand and silty sands with interbedded clay layers with the layers being relatively continuous across the site (Shaw 2008). Groundwater contamination is confined to two zones, a shallow zone (8 to approximately 15 ft bgs) and an intermediate zone (approximately 19 to 26 ft bgs). These zones correspond to coarser-grained layers in the sediment deposits at the site.

The historical area of contaminated groundwater is relatively small, being ~500 ft long and 200 ft wide. The original TCE was observed to have partially reduced to cis-DCE and VC, and these two compounds were observed in the highest concentrations. Concentrations of cis-DCE over 90  $\mu$ g/L were detected; VC was detected at over 5,000  $\mu$ g/L (Shaw 2008).

Hydraulic conductivities calculated from the site range from 0.5 to 21.7 ft per day with a geometric mean of 2.4 ft per day (Shaw 2008). The estimated percent fines is relatively low at 2.5%. The heterogeneity of the treatment zone is 13 layers per 10 ft of aquifer. Drainage ditches at the site control shallow groundwater flow, and groundwater flow directions fluctuate based on whether the ditches are full or empty.

Remediation at FT-07 has included a combination of technologies including excavation, *in situ* chemical oxidation, and groundwater extraction. Two source areas in soil have been identified and removed in two separate events. Soils were excavated in 2000 and 2006, and a potassium permanganate *In Situ* Chemical Oxidation (ISCO) pilot test was implemented in 2002.

Multiple phases of bioremediation treatment have been applied at the site including:

- Phase I and II using sodium lactate via injection wells;
- Phase III in 2007 using sodium lactate via injection wells;
- Phase IV in 2008 using sodium lactate via injection wells and direct push;
- Phase V in 2009 using LactOil<sup>®</sup>, via injection wells and direct push; and
- Phase VI in 2012 using sodium lactate via direct push.

The approach included injection of substrates and extraction of groundwater, however the extracted groundwater was not recirculated. Instead, the extracted groundwater was removed for *ex situ* treatment off-site. Each treatment phase is summarized below.

## Phase I and II (2006)

The initial two phases of bioremediation treatment occurred in two steps – Phase I in June 2006 and Phase II in November 2006. Sodium lactate was injected in 50 temporary injection wells laid out in an irregular grid. Approximately 2,350 gallons of sodium lactate product were injected during these two events (Shaw 2008).

## Phase III (2007)

A follow-up injection event was conducted in the same area in July 2007. Approximately 900 gallons of sodium lactate product were injected. Injections were conducted in existing injection wells and also via direct push (Shaw 2009).

## Phase IV (2008)

The next phase of injections occurred in March and April of 2008 with injection of an additional 670 gallons of sodium lactate product. The injections were all via DPT in this phase. Injections were conducted at ~ 280 locations. Bioaugmentation was used for the first time in this phase with introduction of the SDC-9<sup>TM</sup> culture (Shaw 2009).

## Phase V (2009)

LactOil<sup>®</sup>, rather than sodium lactate, was used in Phase V in 2009. Approximately 2,280 gallons of LactOil<sup>®</sup> were injected into 59 injection wells and 2333 DPT locations. The dosing concentration was approximately 1.6% (Shaw 2011).

## Phase VI (2012)

The injections conducted in 2012 targeted a smaller portion of the previous treatment areas. Sodium lactate was applied at 64 locations. Sodium bicarbonate was added for pH control (Shaw 2013). This series of treatments was highly successful at reducing the concentrations of VOCs. VC and cis-DCE were the most recalcitrant of the VOCs, but eventually these were successful degraded. Six phases of bioremediation were used at the site. Based on the results achieved, no further active bioremediation is planned just natural attenuation of remining  $\mu g/L$  concentrations of cis-DCE and VC.

## B.4 North Island Naval Air Station, OU 24

Naval Air Station North Island (NAS North Island) is located in San Diego County, California, on the tip of the Silver Strand peninsula that separates San Diego Bay and the Pacific Ocean. OU 24 at NAS North Island is identified as the groundwater plume near Building 653 in the northeastern part of the installation. It consists of shallow groundwater impacted with TCE, cis-DCE and VC, likely the result of leaks from the industrial wastewater lines located south of Building 653 (Trevet 2014). Other investigations indicated that releases were associated with an acid waste pump station located south of Building 653 (Geosyntec 2006). The plume of groundwater contamination historically extended approximately 500 ft where it discharged to San Diego Bay.

Groundwater is primarily anoxic with DO less than 1 mg/L and ORP is low or negative. Sulfate is elevated at the site with concentrations exceeding 1,000 mg/L in the source area. NAS North Island is located on generally flat land with an average elevation of approximately 20 ft above mean sea level. The shape and size of NAS North Island has been modified considerably, primarily as a result of the addition of artificial fill derived from the dredging of San Diego Bay. The uppermost 30 ft of sediments are designated "Zone A" and "Zone B" monitoring zones at the site. The uppermost unit, the hydraulic fill material, consists primarily of silty fine sand. The artificial fill material can be up to 20 ft thick and is discontinuous across the island. Beach sand is located below the artificial fill (Trevet 2014).

Water level monitoring data indicates the groundwater gradient is minimal at 0.0004 to 0.0007 ft per ft, and the flow direction is radial. The rate of groundwater movement was estimated to be between 19 and 33 ft per year (Trevet 2014). The hydraulic conductivities in the A and B monitoring zones are about 10 ft per day (Geosyntec 2006).

The bioremediation system at the site included a recirculation system in the source area and a series of six downgradient biobarriers aligned perpendicular to the plume (Geosyntec 2006). Remediation began in December 2006 with the installation of monitoring wells and recirculation wells for an active treatment system in the source area near Building 653. Full operation of the active system began in May 2007. The active system is combined with a series of biobarriers in the downgradient portion of the plume. Biobarrier installation was completed in November 2007. In 2008 and 2009, the source-area system was optimized to target a smaller area and the biobarriers were recommended to stimulate additional degradation.

The remedial design included sodium lactate introduced once per week for a 2-hour pulse at a minimum flow rate of 0.3 gpm to provide a minimum loading of 36 gallons of sodium lactate product per week. With a designed recirculation flow rate of 1.5 gpm, the dosing concentration for the sodium lactate solution was 20%.

The treatment resulted in substantial reduction in VOC concentrations in most of the monitoring wells. In January 2010, the active recirculation system was shut down because performance monitoring results demonstrated that concentrations in the source area had been reduced to below project action levels.

Additional upgradient contamination was subsequently discovered and the remediation was expanded in 2012 to the upgradient areas. The expansion was completed in December 2011 and treatment commenced in March 2012. In March 2013 the upgradient recirculation system was shut down because performance monitoring results demonstrated that concentrations in the upgradient area had been reduced to below project action levels. The remediation at the Building 653 source area is considered complete. However, additional bioremediation treatment has continued in the upgradient areas and the downgradient system of biobarriers.

## **B.5 Orlando National Training Center, Study Area 17**

The former Orlando National Training Center is a former U.S. Navy facility located in the city of Orlando in north-central Florida. Study Area 17 (SA 17) is a former motor pool area at the installation. Previous site activities related to a motor pool area have contributed to subsurface soil and groundwater contamination from TCE. The specific source of the contamination is unknown.

VOCs adversely impacted groundwater throughout the surficial aquifer and in isolated areas within the upper part of the intermediate aquifer of the Hawthorn Group sediments. Given the contaminant distribution pattern, the plume appeared to have originated from two release points at the surface located in the western and central parts of the former motor pool area. In the western source area, compounds detected at the highest concentrations were cis-DCE and VC, with a combined maximum concentration of  $400 \mu g/L$ . In the eastern source area, TCE was the predominant compound detected, with a maximum concentration of  $577,000 \mu g/L$ . The highest contaminant concentrations were detected at the water table interface in the source areas and along the upper surface of a silty sand layer that is located between 15 and 25 ft bgs. This layer and another somewhat deeper layer of silty sand act as apparent aquitards that divide the surficial aquifer into three units – shallow, intermediate, and deep.

The VOC contamination within the Target Treatment Zone (TTZ-1) extended vertically through the surficial aquifer from the water table (approximately 5 ft bgs) to the top of a confining layer at an approximate depth of 50 ft bgs. The historical lateral footprint of the source area is approximately 50 ft long and 50 ft wide. The vertical treatment zone is approximately 45 ft deep.

The plume dives downward at the site (Solutions 2011). The plume extends at the water table interface from both source areas for a distance of approximately 50 to 100 ft in the direction of groundwater flow (east-southeast). In the intermediate unit of the surficial aquifer, the plume extended to a distance of approximately 250 ft downgradient, and in the deep unit of the aquifer, the plume extended approximately 300 ft from the source areas.

The surficial aquifer at the site contains interbedded sand and silty sand to a depth of approximately 25 ft. The surficial aquifer is underlain by sediments of the Hawthorn Group consisting of interbedded sand, silty sand and clay. The highest contamination is in the surficial aquifer. The water table is at approximately 6 ft below ground surface (bgs) across the site, with a variation of 2 ft. The surficial aquifer extends to a depth of about 50 ft bgs with its lower extent defined by the uppermost Hawthorn clay layer. Zone A (5 to 15 ft bgs), Zone B (15 to 30 ft bgs), and Zone C (30 to 50 ft bgs) were monitored. Monitoring Zone C is within the Hawthorn Group sediments at a depth of approximate 45 to 50 ft (Solutions 2011).

The horizontal gradient ranges from 0.003 to 0.004 ft/ft. A downward vertical hydraulic gradient of 0.007 to 0.020 ft/ft exists within the surficial aquifer except near the drainage ditch, where groundwater discharges to the ditch and an upward gradient of approximately 0.25 ft/ft exists. Hydraulic conductivity was calculated based on the August 2005 aquifer pump test to be 4.7 ft/day in the surficial aquifer, and 6.9 ft/day in the deeper intervals of the surficial aquifer (Solutions 2011).

Prior to bioremediation, *In Situ* Chemical Oxidation (ISCO) with hydrogen peroxide was conducted from November 2000 through September 2002 (CH2M 2003). Several phases of ISCO injections occurred through that time period. A total of 69 injection wells were installed within the three vertical depth horizons (Zones A, B, and C) in the first phase. Additional direct-push injections were added in subsequent injection phases. An evaluation of the pre-and post-treatment monitoring data indicated an

average VOC concentration decrease of over 80%. However, VOC concentrations remained above 1,000  $\mu$ g/L in some areas (CH2M 2003).

Multiple phases of subsequent bioremediation treatment with different treatment approaches have included:

- Phase 1 in 2006 which included recirculation in Zones B and C with EOS injections;
- Phase 2 in 2008 which included direct push injections of EOS in Zone B only; and
- Phase 3 in 2012 which included injection into Zones B and C wells without circulation. These phases are briefly described below.

## Phase 1 Recirculation (2006)

The recirculation system included twelve injection wells arranged in a circular pattern around the target treatment zone. Two extraction wells were installed in the center of the circle in Zones B and C. Zone A was not targeted for treatment.

The recirculation equipment was housed in a mobile process trailer. The approach included recirculation of Emulsified Vegetable Oil (EVO), an approach not commonly employed with this substrate. The other evaluation sites with recirculation systems used high solubility substrates such as sodium lactate rather than EVO with a low solubility. EVO was metered into the recirculation stream at a concentration dosing of about 10%. The total loading was 990 gallons of EVO product. Bioaugmentation was not used.

## Phase 2 Direct Push Technology (DPT) (2008)

Based on post-treatment monitoring results for Phase 1, additional polishing injections were implemented for Zone B. The injections included 50 gallons of EVO product injected at six DPT locations. The injection water was diluted to a 6% dosing concentration prior to injection.

## Phase 3 Injections (2012)

A third phase of injections occurred in 2012. The twelve injection wells installed for Phase 1 were used for the injections. A total of 456 gallons of EVO product were injected. Phase 3 included the addition of a buffering agent for pH control.

The treatments resulted in a significant reduction in TCE in most of the monitoring wells, however, cis-DCE accumulation was significant in many wells. The site has switched to a long-term monitoring program with reliance on natural attenuation processes for further degradation.

## B.6 Naval Base Ventura County Point Mugu, Site 24

The Naval Base Ventura County (NBVC) Point Mugu facility is located in Ventura County, California, approximately 50 miles northwest of Los Angeles. Site 24 at NBVC Point Mugu consists of two former Underground Storage Tanks (USTs): UST 23 and UST 55. In 1970 a single 550-gallon concrete tank (UST Site 23) was installed at the Site and was used as an oil/water separator, and the suspected source of groundwater contamination is associated operations.

Much of the developed land of NBVC Point Mugu was formed from mechanically compacted fill material. Fill thickness and composition vary widely (TtEMI 2004).

Upon investigation of the site, VC was reported as the predominant VOC in the plume. The specific chlorinated solvent that constituted the original source is uncertain, but presumed to be TCE and/or PCE. The predominance of VC at this site is presumed to be due to extensive reductive dehalogenation with relatively little and or slow degradation to ethene. The treatments at this site provide information to evaluate bioremediation of a plume with VC as the principal initial constituent.

Three general lithologic units are present at the site: (1) a coarse to fine sand, mostly fill, from ground surface to about 5 ft bgs; (2) a clay layer from about 5 to 10 ft bgs; and (3) primarily sand with some silty sand and silt from about 10 to 90 ft bgs (AIS-TN 2014).

Two separate monitoring zones were established at the site, designated Zone A and Zone B. Zone A extends from 4 to 12 ft bgs and Zone B extends from 20 to 35 ft bgs. These two zones are within the same shallow unconfined aquifer. A laterally-continuous clay layer is encountered in Zone A at a depth of approximately 8 ft bgs. This clay layer acts as an aquitard locally, but its thickness is highly variable, pinching out totally in some areas; it has also been compromised by excavation and well installation activities. Zone B includes a relatively homogeneous sand/silt unit (AIS-TN 2012).

The Navy completed pilot testing at Site 24 in 2002 to evaluate the effectiveness of a bioremediation system to degrade chlorinated solvents in groundwater (Shaw 2003). In 2006, a phase I pilot test was designed and implemented to evaluate the effectiveness of substrate injection/delivery using Direct Push Technology (DPT). Phase 1 substrates included emulsified soybean oil and lactic acid. Phase 2 batch injections of soybean oil and lactic acid were conducted between September and December 2006 as detailed below.

## Phase 1 (2006)

The phase I pilot test was designed and implemented to evaluate effectiveness of substrate injection/delivery using direct-push technology and the use of emulsified soybean oil (EOS) as a longer-term electron donor (TN&A 2006). The approach used a substrate of mixture of EOS and lactic acid. A total of approximately 2,000 gallons of EOS and 1,000 gallons of lactic acid were injected. The injections were completed over a one-week period in July 2006. Twelve DPT locations were used with a spacing of 20 ft.

## Phase 2 (2006)

Phase 2 batch injections of EOS and lactic acid in Zone B were initiated on September 2006 and were completed in December 2006. Groundwater was extracted from a central extraction well, amended with substrate and re-injected to the surrounding injection wells. Approximately 8,000 gallons of EOS and 4,500 gallons of lactic acid were injected via 8 injection wells. The groundwater extraction and treatment system processed approximately 1.7 million gallons of groundwater from September 2006 to February 2007 (TN&A 2007).

## DPT Injections (2013)

EOS, sodium lactate, and microbes (BAC-9; same culture designated SDC-9) were injected into the Zone A and B groundwater through 18 DPT injection points covering an area of approximately 10,800 square ft using a portable mixing and injection system in January 2013. The system consisted of trailer mounted equipment and a substrate/groundwater mixing and storage tank. The total volumes of EOS and sodium lactate injected were 1,050 and 210 gallons, respectively.

The electron donor substrates, sodium lactate and EOS, were diluted approximately 10 to 1 with anoxic groundwater and injected in 5-ft intervals. The distance between the injection points was approximately 20 ft within each injection row and the injection point rows were approximately 20 ft apart. Substrate and microbes were delivered to the subsurface in Zones A and B between 5 and 35 ft bgs.

The results of the treatment were mixed. Substantial reduction of all VOC constituents were observed in many of the wells, however substantial VC accumulation remained in other monitoring wells. Upon completion of the bioremediation operations, the site entered into a long-term monitoring program to evaluate the long term effectiveness of the treatments.

## B.7 St Juliens Creek Annex, Site 21

The Navy's St Juliens Creek Annex (SJCA) is located near Chesapeake, VA. Site 21 is in the southcentral portion of the installation. Most of Site 21's ground surface, with the exception of a few small, unconnected grassy areas, is covered with asphalt. The general topography of the area is flat, with elevations ranging from 7 to 9 ft above msl.

Site 21 is associated with Building 187, a locomotive maintenance facility where TCE was used. The site encompasses a number of nearby industrial buildings, which historically were used as maintenance shops, electrical shops, and munitions-loading facilities. Waste oils and degreasers (including TCE) were reportedly disposed of on the ground surface and around the railroad tracks in this industrial area (CH2M 2015). Multiple isolated sources of contamination are suspected.

Historical contamination in groundwater was distributed unevenly due presumably due to the multiple sources of contamination. Contamination contour maps show an amoeba-like pattern to the outlines of the plume with higher concentrations scattered with the plume. The contaminant distribution maps lack the elongate plume pattern typical of sites with a relatively constant groundwater flow direction. Ten sub-areas of the site were treated, designated Group 1 through Group 10.

The subsurface geology at the site consists of the fine to coarse silty and clayey sands of the Columbia aquifer underlain by the high-plasticity clay of the Yorktown Formation. The Columbia aquifer extends to a depth of between 13.5 and 20 ft bgs. Shallow groundwater flow velocity has been calculated to be approximately 72 ft per year. Shallow groundwater at Site 21 is encountered from 2 to 7 ft bgs and flows southwest in the eastern portions of the site and southeast in the western portion of the site, toward the storm sewer system east of Building 1556 (CH2M 2015). Shallow groundwater is fresh, but becomes brackish at depths greater than 30 ft.

Groundwater contour maps show an irregular pattern of groundwater elevations. In general, groundwater elevations decrease in a southwestern direction toward St Juliens Creek (Tidewater 2014). Surface water drainage ditches may have a role in distorting groundwater flow along with tidal effects.

Previous remediation at Site 21 included a small-scale pilot study to evaluate the ability to enhance natural attenuation through bioaugmentation with aerobic bacteria that degrade cis-DCE (ESTCP 2010). Zero-valent iron (ZVI) injections were initiated in December 2010 with injection DPT at three areas of the site (designated North, South and East Areas).

EVO injections were initiated in April 2011 and completed in September 2011. A total of 10,758 gallons of EVO was injected into ten treatment areas of the site (designated Group 1 through Group 10). Nine of the ten EVO treatment areas received EVO only without ZVI. One of the ten areas received treatment with both ZVI and EVO.

The bioremediation at Site 21 was generally successful with substantial reduction of contamination. Arsenic was monitored throughout the treatment and arsenic concentrations substantially increased after treatment was initiated. Monitoring indicates several areas resistant to treatment and requiring follow-up action. Additional EVO injections were conducted in May 2014 (CB&I, 2014).

## B.8 Vandenberg Air Force Base (VAFB), Site 15A/15B

## VAFB Site 15A

Vandenberg Air Force Base is located near the Pacific Ocean near Oxford, California. Site 15 at was active from 1960 to 1967 and used to launch Atlas D and Atlas F missiles. Prior to launches, TCE was used to degrease rocket motors. As part of the launch process, TCE along with deluge water was discharged to the drainage channels and subsequently to the ground surface. Site 15A is associated with Launch Pad 1 and Channel A extends about 800 ft northwest from the launch pad.

Groundwater in the Site 15 area has been impacted by dissolved TCE and its degradation products. Two TCE source plumes with concentrations greater than 1,000  $\mu$ g/L have been identified (AECOM, 2009). These source areas are located at the discharge point of Channel A and near the 90-degree bend in Channel B. TCE concentrations greater than 5  $\mu$ g/L extend almost 2,000 to 3,000 ft toward San Antonio Creek.

Site 15 is located on Pleistocene sand dune deposits derived from the Pacific Ocean beaches located 1.5 miles west of the site. Surface topography of the site consists of sand dune hills with small closed depressions. The overall surface slopes southwest to San Antonio Creek, which flows west-northwest, passing about 3,000 ft southwest of the launch pads. The dune sand overlies the Orcutt Formation which consists of wind-deposited sand and local gravel deposits near the base. The combined deposits above the bedrock are typically 25 to 50 ft thick.

The plume of contamination occurs in the unconfined groundwater in the unconsolidated dune sand and underlying coarse sand and gravel and deposits. AECOM (2009) noted that the saturated thickness of the aquifer varied from approximately 4 to 45 ft. The depth to groundwater at the site varies from a few ft bgs in some areas to almost 50 ft bgs in places. Slug test and aquifer tests at the site provide highly variable results for hydraulic conductivity. A value of 50 ft per day is a representative value (Shaw 2009).

Active bioremediation was performed in several phases. Phase I of the recirculation operations extended from May 2011 to January 2012 and included operation with five extraction wells and four injection wells. Phase II included conversion of the five extraction wells to injection wells and installation of additional well points for both injection and extraction. Phase II operation extended from February 2012 to March 2013. The Phase II system included four injection wells and six extraction wells (Shaw 2014).

Alkalinity monitoring was used to evaluate distribution of amendments from the injection well to the extraction wells, a rarely used method to obtain real-time results.

The workplan called for injection of 12,140 gallons of sodium lactate product in Phase I and 3,000 gallons in Phase II. The SDC-9 bioaugmentation culture was introduced directly into injection wells during Phase II. The sodium lactate was diluted with water to provide a dosing concentration of approximately 0.2%.

The recirculation system was initially composed of five downgradient extraction wells and four upgradient injection wells. The system included an additional five extraction wells and conversion of the original extraction wells to injection wells, thereby "sweeping" across the source area westward. The average injection rate was estimated to be 9 gpm per well or 35 gpm total, with injections of sodium lactate at concentrations up to 10,000 mg/L. The average groundwater extraction rate was estimated to be 7 gpm per well or 35 gpm total. Hydrogen gas was also added as a gaseous electron donor. A total of 5,640 standard cubic ft (~29.5 pounds) was added during the two phases

Substantial declines in VOC concentrations were noted while recirculation was ongoing in Phase I and II. Many wells had sustained VOC declines post treatment; however substantial rebound was observed in some wells after the recirculation operations had ceased (CB&I 2015). Additional injections were planned for 2015 but the details concerning these injections and the results were not available at the time of this project.

## VAFB Site 15B

Similar to the previously described Site 15A, Site 15B at VAFB was active from 1960 to 1967 and used to launch missiles; prior to launches TCE was used to degrease rocket motors. Site 15B had a separate drainage channel for wash-down water - Channel B. Investigations at Channel B indicate the entry point for groundwater contamination was at a 90-degree bend mid-way in the channel. The plume of contamination at Site 15B is shorter that that at Site 15A and extends approximately 2,000 ft from the launch pad area to San Antonio Creek. TCE and it daughter products are the primary contaminants. (Shaw 2009). TCE concentrations greater than 5  $\mu$ g/L extend almost 2,000 to 3,000 ft toward San Antonio Creek.

As was the case previously described for Site 15A, the impacted shallow aquifer at Site 15B consists of unconfined groundwater in the unconsolidated dune sand and underlying coarse sand and gravel and deposits.

A groundwater recirculation system was initially planned for Site 15B, however the approach was modified due to the low expected extraction rates (Shaw 2013). The approach was modified to include 24 temporary injection well points. Twelve additional injection well points were added in the course of the operation. Injection and recirculation operations occurred in two phases, the first in July 2011 and the second in May 2012 as summarized below.

Initial remediation activities included injection of EVO and lactate (4,000 gals each) into 24 well points at Channel B in July 2011. Additional injections were conducted in May 2012. The bioremediation resulted in significant reduction of concentrations in the source area although the *cis*-1,2-dichloroethene (cis-DCE) concentrations remained in the downgradient areas at concentrations greater than 1 mg/L.

## **B.9 Vandenberg Air Force Base, Site 19**

Site 19 at Vandenberg Air Force Base is an active NASA facility and is largely occupied by buildings and parking lots. Improper disposal of waste oils and solvents reportedly occurred in the storm water drainage ditch located to the south of Building 836 during the years that the facility was operated. Reportedly waste solvents were also disposed of by pouring the liquids into a storm drainage grating at the southeast corner of the site. In the past, TCE was used as a degreasing agent for parts cleaning and was likely discharged to the ground surface (MWH 2008-02).

Investigations conducted since the mid-1980s have indicated that the primary groundwater contaminants at the site are PCE and TCE along with their daughter products. Other VOCs and petroleum-related compounds including benzene have also been detected at low levels. TCE was considered the main VOC of concern, because it was detected at substantially high concentrations in the subsurface media. TCE-impacted soil was found mainly between two buildings (830 and 836) at 10 to 11 ft bgs in soil near the groundwater table interface.

Two plumes of contamination have been identified – the "TCE Plume" and the "PCE Plume". The TCE Plume is approximately 250 ft in length and 110 ft in width. The highest TCE concentrations are observed in the Building 836 source area (3,300  $\mu$ g/L), and the contamination is mostly in the upper 35 ft at the site. However, contamination has been observed near monitoring wells at a depth of 70 to 85 ft (MWH 2008-02).

Site 19 is underlain by Quaternary alluvial deposits of the Santa Ynez River floodplain. The upper 5 ft of sediments consists of sand, silty sand, and sandy silt. Between 5 and 20 ft bgs, plastic clay and silty clay were encountered. These unconsolidated sediments are underlain by medium to coarse grained sand and gravelly sand to an approximate depth of 30 to 35 ft. Below this sand unit, the unconsolidated sediments consist of interbedded plastic clayey silt and silty sand. Treatments have occurred in both the upper Clay Zone and the lower Sand Zone. Based on groundwater elevation data collected since 1987, the average groundwater gradient is approximately 0.002 to 0.003; site groundwater flows north-northeast toward the Santa Ynez River.

Phases of bioremediation at the site included:

- Pilot Study with Hydrogen Release Compound<sup>®</sup> (HRC<sup>®</sup>) and dehalogenating culture BDI in 2006;
- Reinjection of HRC<sup>®</sup> and BDI in 2008;
- Injection of LactOil<sup>®</sup> in 2009;
- Injection of LactOil<sup>®</sup> in 2011; and
- Additional injections in 2014.

## HRC® Pilot Study 2006

The pilot study included the use of HRC<sup>®</sup> and BDI. Injections were performed in the TCE treatment area in May 2006 and again in the immediate vicinity of monitoring well 19-MW-17A in July 2008. The treatment included injection via DPT of HRC<sup>®</sup> and BDI in both the shallow Clay Zone and deeper Sand Zone. A total of 17,364 pounds of HRC<sup>®</sup> and 126 L of BDI were injected. Spacing was 5 ft in the Clay Zone and 10 ft in the Sand Zone. The HRC<sup>®</sup> and BDI injections had limited success in maintaining reducing conditions in the TCE treatment area.

## LactOil®/Lactate Injections 2009

ISB injections (8 ft spacing) occurred between December 2009 and January 2010 and consisted of injections of emulsified vegetable oil (LactOil<sup>®</sup>), lactate and the dechlorinating microbial culture SDC-9 in the TCE treatment areas (Shaw, 2011). The injections included a mixture of LactOil<sup>®</sup> (500 gallons) and lactate (20 gallons) delivered at a dose of about 8%. The injections targeted the shallow Clay Zone only. The bioaugmentation cultures BDI and SDC-9 are the same culture.

The data collected during subsequent groundwater monitoring events occurring over the 5 years following the ISB treatment injections showed significantly reduced TCE concentrations in some areas. Ethene and ethane were observed indicating complete reductive dechlorination in some locations, however, TCE continued to persist in other regions. Additional injections were conducted in 2011 and 2014.

## B.10 Vandenberg Air Force Base, Site 35

Site 35 at Vandenberg Air Force Base, also known as Missile Silo 576-G, is a rocket launch facility. Atlas F missiles, fueled with a combination of rocket propellant and liquid oxygen, were launched from the site in the 1960s. Mixed solvents, lubrication oils, hydraulic fluids and TCE were typically used at the dry pad launch facilities (Tetra Tech, 2005).

The plume is long and narrow with TCE concentrations above 1 mg/L extending over 0.5 mile from the source. Investigations at the site indicate that the plume occurs within unconsolidated sediments in a paleo-channel incised into the underlining bedrock. Groundwater is moderately aerobic with DO of 1 to 3 mg/L and a positive ORP.

The subsurface is composed of loosely consolidated deposits of the Orcutt Formation, consisting of beds of sand, gravel, and clay of predominantly continental origin, with the upper zone representing eolian and beach sand. Shale bedrock of the Sisquoc Formation underlies the Orcutt Formation sediments at a depth of approximately 40 ft.

Groundwater beneath the site is unconfined and the water table surface is approximately 20 ft below ground surface. Groundwater appears to flow following the bedrock topography to the southwest at a relatively high hydraulic gradient of 0.05 ft per ft (Tetra Tech, 2005). Aquifer tests estimated the hydraulic conductivity between 1.1 and 5.8 ft/day (Tetra Tech, 2008b).

Source area remediation included soil excavation prior to bioremediation activities. In April 2010, source area soils were excavated and in September 2011 additional source area treatment included injection of EHC, a combination of zero-valent-iron and carbon-substrate product.

Biobarriers have been employed as a bioremediation approach at two locations, one located near the source and one farther downgradient. These are named the ISB Biobarrier and the Distal Biobarrier, respectively. The near-source ISB Biobarrier is a subject of evaluation in the statistical analysis.

Initial pilot studies were conducted from 2001 to 2003 were followed by a full-scale application in 2009. The system included nine injection wells extending approximately 180 ft aligned perpendicular to the plume. The spacing was approximately 25 ft. A total of 1,600 gallons of LactOil<sup>®</sup> and 385 gallons of sodium lactate product were injected. Bioaugmentation included 1546 L of SDC-9 culture. The substrates were diluted prior to injection to provide a dosing concentration of approximately 1.5%.

Post injection monitoring indicated substantial decreases in VOC concentrations within the first 2 years in downgradient monitoring wells followed by a rebound in concentrations. Additional injections were planned.

Appendix C: Groundwater Sampling for Nucleic Acid Biomarker Analysis

# 2 Groundwater Sampling for Nucleic Acid Biomarker Analysis

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**Abstract:** Microbes catalyze relevant transformation processes in aquifers (e.g., contaminant detoxification) and monitoring microbial biomarkers in groundwater yields valuable information regarding their presence, abundance and activity as well as their spatial and temporal dynamics. This chapter provides protocols for *on-site* and *off-site* biomass collection from groundwater for biomarker analysis.

#### 1 Introduction

Freshwater-saturated subsurface environments (aquifers) play a pivotal role in supplying drinking water for a growing human population (Gleick, 1996). Unfortunately, anthropogenic contamination impairs groundwater quality, and supplying clean drinking water has become increasingly challenging. Common groundwater contaminants include organic compounds such as gasoline constituents and chlorinated solvents, as well as inorganic compounds including metals, radionuclides, metalloids, nitrate, and perchlorate, to name a few (www.epa.gov/ogwdw/hfacts.html).

A wealth of information has accrued over recent years about the microbiology contributing to the detoxification of common groundwater pollutants such as chlorinated ethenes (Bradley, 2003; Freedman and Gossett, 1989; He et al., 2003; Holliger et al., 1993; Maymó-Gatell et al., 1997; Sung et al., 2006). A productive pathway leading to detoxification is the stepwise reductive dechlorination of tetrachloroethene (PCE) to trichloroethene (TCE), dichloroethenes (DCEs), vinyl chloride (VC) and ethene catalyzed by bacteria capable of respiratory reductive dechlorination ([de]chlororespiration) (Futagami et al., 2008; Löffler and Edwards, 2006).

The key microbes with the ability to efficiently dechlorinate chlorinated ethenes, including DCEs and VC, to ethene are *Dehalococcoides* (*Dhc*) spp., and a link between *Dhc* presence and detoxification has been established (Ellis et al., 2000; He et al., 2003; Hendrickson et al., 2002; Lu et al., 2006). Bioremediation of contaminated aquifers using both biostimulation (i.e., electron donor additions) and bioaugmentation with Dhc-containing consortia or groundwater were successfully implemented (Lendvay et al., 2003; Lookman et al., 2007; Major et al., 2002; Ritalahti et al., 2005; Scheutz et al., 2008). A critical site assessment component prior to technology implementation is knowledge of the presence, abundance and potential for activity of the microbiology contributing to the process of interest (i.e., detoxification). Further, monitoring microbial abundance after technology implementation provides site managers with information on how to control (i.e., enhance) and sustain process performance (i.e., detoxification rates). A few Dhc biomarkers to indicate Dhc presence and abundance have been identified and *Dhc*-specific prognostic and diagnostic tool kits targeting the *Dhc* 16S rRNA gene and three reductive dehalogenase genes (i.e., tceA, vcrA, bvcA) (See O Chapter 26, Vol. 5, Part 2) have been designed (Cupples, 2008; Holmes et al., 2006; Löffler et al., 2000; Regeard et al., 2004; Ritalahti et al., 2006, 2005).

Sampling aquifers for microbial and geochemical analyses in support of bioremediation applications faces several constraints and the ideal sampling regime with regard to sampling locations, sampling frequency and sample type can typically not be achieved. For example, site access limitations, obstruction by existing infrastructure, lack of sampling wells, and shortage of funds for well installation, sample collection and analysis constrain sampling efforts. For practical reasons, sampling focuses on groundwater, and solid aquifer samples are typically not included in routine analyses. The focus on groundwater is justified for the analysis of targets like *Dhc* that distribute in the aqueous phase (e.g., planktonic microbial cells) (Lendvay et al., 2003; Major et al., 2002); however, the distribution of organisms between solid and aqueous phases in different aquifer matrices is poorly understood for most microbial populations (Amos et al., 2008). Hence, the factors controlling attached versus planktonic growth of target microbes warrants further investigation so that the value added by performing microbial analysis with aquifer solids can be evaluated.

This chapter provides protocols for groundwater collection and biomass recovery for subsequent nucleic acid extraction and biomarker analysis. These procedures are currently applied to *Dhc* biomarker analysis in support of bioremediation at chlorinated solvent sites but the protocols are universal and should be applicable for monitoring other planktonic populations of interest.

#### 2 Approach

#### 2.1 General Considerations

Prior to sampling, the expected outcomes (goals) of the analysis should be clearly defined. What information should the analysis provide? Does qualitative information (i.e., *Dhc* presence) suffice or is quantitative information (i.e., *Dhc* abundance) desired? Will samples be collected only once or will the site be monitored over time? A crucial consideration is that the sampling protocol for a given well (or site) is maintained for the duration of the monitoring efforts. Changes to the protocol during long-term monitoring will complicate data interpretation and should be avoided. Aseptic techniques should be employed to the extent possible when handling groundwater destined for laboratory analysis. An appropriate canopy or cover is recommended to protect sample equipment and samples from direct sunlight or rain, and samples should be packed in a cooler with ice without delay.

Since the filtration method is expected to trap the majority of microbial cells, the amount of groundwater collected determines the number of biomarker genes available for subsequent PCR analysis. Hence, filtering large volumes of groundwater seems beneficial, but for practical purposes (e.g., time constraints, slow well recharge), 0.5–2 L of water are typically collected; however, depending on groundwater characteristics (e.g., fines that clog the membrane filter), as little as 10 mL may be gathered, which may or may not be sufficient for subsequent biomarker analysis. Whether sufficient biomass was captured depends on the abundance of the target biomarker(s) and the types and concentrations of inhibitory compounds. A typical quantitative real-time PCR (qPCR) quantification threshold is 20 gene targets per qPCR reaction. Assuming 1,000 target gene copies in 1 L of groundwater were concentrated to a 100  $\mu$ L volume of DNA, then 10 target genes would be added to the qPCR reaction tube per  $\mu$ L of template. Consequently, as the volume of groundwater filtered decreases, the detection threshold increases, and a 100 mL groundwater sample will yield a positive result only if more than 1 × 10<sup>4</sup> target gene copies are present per liter of groundwater (Cupples, 2008; Rahm et al., 2006; Ritalahti et al., 2006).

#### 2.2 Selection of Sampling Wells

*Dhc* abundance data are valuable throughout the remedial investigation and feasibility study process, and samples for *Dhc* analysis should certainly be collected prior to designing a

bioremediation system (Stroo et al., 2006). Following implementation of remedial action, temporal *Dhc* data from a treated aquifer zone are critical for evaluating and managing the bioremediaton system's performance.

Groundwater samples for *Dhc* analysis should be collected from source area(s) and down gradient plume locations where biodegradation products may have been observed or are anticipated, and where geochemical conditions are favorable for anaerobic bioremediation. If available, samples from a well outside (i.e., upstream) of the contaminated zone(s) should be included, at least in the initial site assessment efforts. Wells should produce sufficient water for adequate purging and collecting at least a 500 mL groundwater sample. Well screen depths and lengths should be considered when selecting a sampling location. Ideally, wells installed for establishing vertical profiles of contaminant (i.e., chlorinated ethenes) concentrations are used to establish a similar profile for *Dhc* distribution. Discrete sampling zones in intervals with dechlorination daughter products (i.e., DCEs, VC) are preferred for *Dhc* analysis, although zones where primarily parent compounds (i.e., PCE, TCE) are present may provide additional information regarding the rate-limiting factors controlling contaminant transformation. Wells with extended screens used for injection of substrates (e.g., electron donor) or bioaugmentation culture should not be used for monitoring performance of the bioremediation system.

#### 2.3 Groundwater Sampling Procedures

Groundwater sampling for microbial analysis typically utilizes methods established for evaluating groundwater chemistry. Selecting a method depends on a number of site-specific conditions including sampling depth, well construction, and aquifer permeability, as well as historic site data and regulatory requirements. Groundwater sampling approaches employ a variety of purging and sampling devices and applicable procedures have been described (reviewed in Yeskis and Zavala, 2002). The goal of these procedures is to generate a sample representative of the formation groundwater of the well vicinity. However, geochemical variations in the water column within a well and geochemical and contaminant stratification within a screened interval are affected by seasonal changes (e.g., rain events, temperature changes) and can lead to variation in biomarker abundance and confound data interpretation (Stroo et al., 2006).

The traditional "well volume" groundwater sampling method involves bailers or high speed pumps (>500 mL min<sup>-1</sup>) to purge 3–5 well casing volumes prior to collecting groundwater samples. Alternatively, low-flow purging methods (100–500 mL min<sup>-1</sup>) with a peristaltic or submersible bladder pump are generally recommended to collect groundwater samples for volatile organic compounds (VOCs) and/or geochemical analysis (Puls and Barcelona, 1996). These approaches also apply to microbial (e.g., *Dhc*) analysis, which should occur after geochemical parameters have stabilized. Surging the monitoring well with a surge block or disposable bailer can increase particulate matter in the sample and recovery of associated (i.e., attached) biomass. Whether using traditional "well volume" methods or "low-flow" methods, it is imperative that the same protocol is applied for every sampling event.

This chapter describes two different procedures for sample collection following well preparation (i.e., purging and surging). Procedure 1 (see 3.5.1) relies on collection of ground-water for *off-site* laboratory filtration and biomass collection whereas Procedure 2 (see 3.5.2) collects biomass *on-site* by field filtration using Sterivex-GP 0.22 µm polyethersulfone membrane filter cartridges (Millipore Corporation, Billerica, MA). For analysis of DNA

biomarkers, both approaches provide valuable information (Ritalahti et al., 2009); however, *on-site* filtration using Sterivex cartridges has several advantages. The cartridges are easy to ship, the addition of nucleic acid preservatives to enhance biomarker stability is feasible, and, depending on the aquifer characteristics, larger volumes of groundwater can be collected. In addition, bottle breakage and the disposal of contaminated groundwater in the analytical laboratory are avoided. The primary advantages of laboratory filtration are that the method is commonly used, field personnel are accustomed to the procedure, and the analytical laboratory has greater control over the filtration process (e.g., can prepare replicate filters from the same groundwater sample).

## 3 Protocol for Groundwater Sampling

3.1. Connect a flow-through cell (e.g., the YSI 556 Handheld Multiparameter Instrument, www.ysi.com) to the tubing of the peristaltic pump. Record the sample start time and the field measurements for pH, oxidation-reduction potential (ORP), specific conductance, temperature, dissolved oxygen, and turbidity.

3.2. Disconnect the flow-through cell.

3.3. Lower a polyethylene disposable bailer into the well to the midpoint of the screen and move the bailer up and down in the water column to surge the well. It is important to agitate at the midpoint of the well screen as this step is not intended to stir up sediment in the sump and/or the bottom of the well.

3.4. While continuing to surge the well with the bailer, re-connect the flow-through cell and record the field measurements for pH, ORP, specific conductance, temperature, dissolved oxygen, and turbidity. Disconnect the flow-through cell but continue to surge the well with the bailer through the sample collection process. Surge the well with steady motion avoiding rigorous mixing of sediment from the bottom of the well.

3.5.1. In order to sample groundwater for *off-site* biomass collection, fill the appropriate sample containers (e.g., clean, sterile 1 L amber glass or plastic bottles with Teflon-lined caps) directly from the effluent end of the pump. No preservatives (e.g., acid) routinely used for stabilizing cations and anions in groundwater samples should be added. The bottles should be filled with groundwater from tubing that has already been used to withdraw one to two well volumes of groundwater to ensure that a representative sample of aquifer water, rather than well water, is collected. The bottles should be filled to capacity (i.e., minimal headspace) to minimize air exposure. Apply the Teflon-lined caps and ensure a tight seal.

3.5.2. For *on-site* biomass collection, use sterile Sterivex-GP 0.22  $\mu$ m membrane filter cartridges. Attach 1/4–5/16 in. polyethylene tubing to the inlet of the Sterivex cartridge and secure with a clamp. Place the cartridge over a graduated cylinder that can accurately measure the volume of water filtered. Using a 10 mL syringe filled with air, push any remaining liquid out of the Sterivex cartridge. Close the inlet and the outlet of the Sterivex cartridge with male and female Luer Lock plugs, respectively. Sterivex cartridges without a Luer Lock closure on the outlet end are not recommended. Replicate samples should be collected consecutively without flow interruption. Record the volume of filtered groundwater on the chain-of-custody form (see below) and on the Sterivex cartridge to a separate, new 50 mL Falcon conical plastic tube. The Sterivex cartridges may clog during field filtration before the target volume of

groundwater could be filtered. When membrane fouling occurs, record the volume of groundwater collected so that subsequent data normalization is possible.

#### 4 On-Site Sample Handling and Shipping

Since the stability of microbial biomarkers is of concern, the samples (i.e., bottles filled with groundwater and/or Sterivex cartridges) should be transferred to coolers with ice immediately after sampling and shipped using an overnight carrier to the analytical laboratory.

4.1. Chain-of-custody forms. Immediately following sample collection, record the sampling well location, the well ID, notes on individual samples (e.g., the volume of water that passed through each Sterivex cartridge), date and time of sampling, the handler's name and contact information, and the type of analyses requested. Chain-of-custody forms are either provided by the analytical laboratory or the consulting firm, and must accompany each sample shipment.

4.2. Sample shipping. The coolers with samples are shipped for next day delivery to the analytical laboratory. It is important to notify analytical laboratories when samples are shipped to avoid delays in handling and processing that could affect biomarker integrity.

4.2.1. Groundwater-filled containers: Apply several layers of bubble wrap to protect against breakage and double bag each container in separate Ziploc plastic bags. Transfer the bottles to a cooler and use ice packs and/or blue ice (in Ziploc bags) to ensure refrigeration until arrival at the analytical laboratory. Use additional packing material, as appropriate, to prevent movement and breakage during shipping.

4.2.2. Sterivex cartridges: Place each Falcon plastic tube with one enclosed cartridge into a separate Ziploc bag and transfer to an appropriately sized cooler with ice packs and/or blue ice (in Ziploc bags).

## 5 Sample Handling in the Analytical Laboratory

Upon arrival, transfer containers with groundwater immediately to a 4°C incubator and Sterivex cartridges to a -80°C freezer. Collect biomass from groundwater samples (see 5.1) as soon as possible to minimize biomarker loss and formation of iron precipitates, which interfere with the filtration process and downstream sample processing.

5.1. Biomass collection from groundwater.

5.1.1. Connect a 47 mm diameter, polyethersulfone membrane filter (0.22  $\mu$ m pore size) unit (MO BIO Ultraclean<sup>TM</sup> Water DNA Isolation Kit, MO BIO Laboratories Inc.) to a vacuum filtration system. Use a vacuum manifold for simultaneously filtering multiple samples.

5.1.2. Working aseptically inside a fume hood, filter 100 mL aliquots of the groundwater. Repeat with additional 100 mL aliquots until 1 L of groundwater has been filtered. If less than one liter passes through the filter, record the final filtered volume.

5.1.3. Use flame-sterilized forceps to transfer the membrane filter with the biomass cake to a sterile Petri dish. Following biomass collection, the membrane filters can be stored at  $-80^{\circ}$ C or processed immediately.

5.1.4. Cut the membrane filter into strips with a sterile razor blade. Remove the strips aseptically with sterile forceps and place them into a sterile tube appropriately sized for the DNA extraction method to be used.

5.2. Biomass collection for nucleic acid extraction from Sterivex cartridges.

5.2.1. If stored at  $-80^{\circ}$ C, thaw the Sterivex cartridges at room temperature.

5.2.2. Remove the cartridge from the Falcon plastic tube and record any information written on the cartridge barrel. Wipe the exterior of the cartridge with ethanol and place in a sterile Petri dish.

5.2.3. Remove the plugs from both ends of the cartridge and connect a 10 mL syringe to the cartridge outlet and remove any remaining liquid from inside the cartridge barrel.

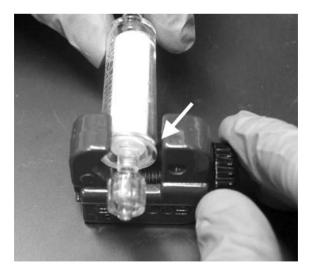
5.2.4. Place the cartridge in an appropriately sized pipe cutter that has been cleaned with 70% ethanol. Fit the blade of the pipe cutter into the seam near the outlet end of the filter cartridge ( $\bigcirc$  *Fig.* 1).

5.2.5. Remove the exterior portion of the cartridge housing by slowly tightening the blade and rotating the cartridge until the cartridge barrel holding the membrane and the housing come apart (**>** *Fig.* 1). The membrane filter remains attached to the interior barrel of the cartridge.

5.2.6. Use a sterile razor blade to cut the membrane filter into 4–12 strips while leaving one end of each strip attached to the housing ( $\bigcirc$  *Fig. 2*). The size, shape and the number of strips cut depends on the tube size (e.g., 2.0 or 15 mL) required for subsequent DNA extraction procedures.

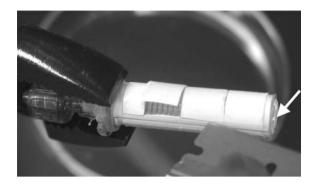
5.2.7. Remove the filter strips including the 0.1  $\mu$ m membrane filter at the end of the barrel (**>** *Fig. 2*) aseptically with sterile forceps and place them into an appropriate tube for DNA extraction. Add any liquid remaining in the cartridge to the same tube.

5.2.8. Standard protocols described elsewhere are applied to extract DNA from the biomass collected on the membrane filters. Protocols for RNA extraction from Sterivex cartridges are described in **O** Chapter 26, Vol. 5, Part 2.



#### Figure 1

The blade of the pipe cutter (white arrow) is fit into the seam near the output end of the Sterivex cartridge. By rotating the cartridge while slowly tightening the pipe cutter, the cartridge housing opens and the membrane filter is exposed.



#### Figure 2

A flame-sterilized razor blade is used to cut the membrane filter into 4–12 strips while leaving each strip attached to the housing. The arrow points to the 0.1  $\mu$ m membrane filter at the end of the barrel. The filter strips, including the round end piece, are removed aseptically with forceps, placed into sterile 2.0 or 15 mL plastic tubes and used for DNA extraction.

## 6 Equipment and Supplies

Microbial (e.g., *Dhc*) analysis requires the standard equipment and supplies used for ground-water sampling to perform VOC and geochemical analyses.

For on-site filtration:

- Sterivex-GP 0.22 µm polyethersulfone membrane filter cartridges (Catalog # SVGPL10RC), Millipore Corporation, Billerica, MA.
- Male and female Luer Lock plugs (Catalog # EW-45503–58 and EW-45500–28, respectively), Cole-Parmer, Vernon Hills, IL.
- Falcon tubes (50 mL) or equivalent plastic containers for protecting Sterivex cartridges during shipping and storage, standard laboratory suppliers such as Fisher Scientific, Pittsburgh, PA and VWR International, West Chester, PA.

Sample processing in the laboratory:

Standard Duty Dry Vacuum Pump (Catalog #2511B-01), Gardner Denver Thomas, Welch Vacuum Technology Inc., Nile, IL.

Filtering manifold (Catalog # 02924-20), Cole-Parmer, Vernon Hills, IL.

- MO BIO UltraClean<sup>TM</sup> Water DNA Isolation Kit (Catalog # 14880–10, 14880–25), MO BIO Laboratories, Inc., Solana Beach, CA for collecting and processing biomass from groundwater.
- MO BIO PowerSoil<sup>TM</sup> DNA Isolation Kit (Catalog # 12888–50, 12888–100), MO BIO Laboratories, Inc., Solana Beach, CA. for processing Sterivex Filters
- Vortex Genie 2 (Catalog # 13111-V), and adapters for 15 mL Falcon tubes (Catalog # 13000-V1–15), and/or 2 mL bead tubes (Catalog # 13000-V1), MO BIO Laboratories, Inc., Solana Beach, CA.

Small tubing cutter, local hardware store.

Petri dishes, forceps, razors, 10 mL syringes, and 2.0 or 15 mL plastic tubes from standard laboratory suppliers (Fisher Scientific, Pittsburgh, PA and VWR International, West Chester, PA)

## 7 Time Considerations

Low-flow purging and sampling methods typically can be completed in 1–2 h per well. Field filtration procedures (i.e., Sterivex cartridges) may add up to 30 min per well, along with additional training of personnel.

An important consideration is the shipment of samples to the analytical laboratory, which should occur without delay to minimize biomarker loss. Overnight carrier delivery is commonly used and the samples should be processed and/or preserved as soon as possible upon arrival. Biomass from groundwater should be harvested as soon as possible after receipt of samples. Filtration of groundwater samples greatly depends on the amount of particulates present and could require from a few min to several hours for each sample. Immediate filtration of groundwater samples is recommended because iron precipitates formed when dissolved ferrous iron is oxidized following air exposure can significantly reduce the volumes of groundwater that can be filtered in acceptable time periods. With practice, processing a Sterivex cartridge will require 5–10 min.

#### 8 Quality Assurance/Quality Control

A one-size-fits-all approach to biomarker analysis has not been adopted, and different analytical laboratories use different PCR primers, qPCR detection chemistries and analytical equipment. Since all of the factors that affect the analysis of environmental samples cannot be accounted for, and repeating the analysis of a particular sample may not be feasible (e.g., if only one Sterivex cartridge was available), good record keeping is key, and begins in the field. The analytical laboratory benefits from having as much information as possible available, in particular with regard to volume filtered, and whether problems were associated with collecting the sample. As with any analytical method, the laboratory must follow standard operating procedures (e.g., for DNA extraction and qPCR). Standard curves should be performed with each qPCR run to verify proper instrument performance and operation. Each sample should include at least three replicate qPCR reactions and multiple (e.g., at least two) dilutions of the extracted DNA solution should be analyzed. Preferably, two or more Sterivex cartridges or groundwater samples are available for each well, so that quantitative results from replicate DNA extractions can be compared. The skilled technical specialist must be able to reproducibly perform the assays, assess data quality and interpret the qPCR results by analyzing the amplification plots. If qPCR data suggest inhibition (e.g., deviation from exponential amplification or poor amplification in undiluted samples), this information should accompany target gene quantification estimates.

#### 9 Trouble Shooting

Negative results (i.e., no *Dhc* biomarker genes detected) suggest that the target microbes (e.g., *Dhc*) are absent and that reductive dechlorination of DCEs and VC to ethene cannot be

expected. Negative results should be interpreted cautiously because the target organisms may be heterogeneously distributed in aquifer formations or the analysis may erroneously fail to detect *Dhc* biomarker genes. For example, false negative results may be caused by several factors including:

- Very low numbers of *Dhc* bacteria are present.
- The characteristics of the groundwater or sampling well(s) limit the volume of groundwater that can be withdrawn or filtered (and hence the amount of biomass that can be analyzed).
- The presence of inhibitors that interfere with biomarker analysis (e.g., via qPCR).
- The loss of biomarkers during sample handling, transport and storage.

To minimize false negative results, samples from multiple wells in different areas within the contaminant zone(s) should be collected and analyzed for the presence and abundance of *Dhc* biomarker genes, in particular during initial site analysis.

DNA extracted from groundwater may contain substances that interfere with PCR analysis either by affecting PCR amplification or by impacting the intensity of the fluorescence signal in qPCR. This effect is not predictable and typically leads to underestimation of the actual biomarker gene copy numbers (Stults et al., 2001). To detect possible PCR inhibition, template DNA should be used undiluted and at a 1:10 or appropriate higher dilution. PCR inhibition is indicated when the undiluted DNA sample yields fewer copies per  $\mu$ l of template DNA than the dilutions, when adjusted for the dilution factor. Another indication of inhibition is obtained from the qPCR fluorescence plot on the real-time thermocycler. Inhibition is indicated when undiluted samples do not display the expected sigmoid, exponentially increasing fluorescence curve. Since dilution may result in too few copies to be detectable with qPCR, it may be prudent to perform a nested PCR that avoids inhibition and provides unsurpassed sensitivity (but quantitative information is lost) (Löffler et al., 2000).

A prerequisite for a meaningful application of the analytical approach is that the biomarkers of interest are stable from the time of sample collection until the actual analysis occurs. Biomarker stability cannot be predicted, varies between organisms, and is influenced by environmental conditions. *Dhc* biomarker genes are considered to be stable for at least 24 h at 4°C, and no differences have been observed between Sterivex cartridges prepared in the field and shipped to the analytical laboratory, as compared with groundwater shipped from the field with the filters prepared in the lab (Ritalahti et al., 2009). Hence, *Dhc* DNA biomarker loss should be negligible with proper sample handling procedures and short-term refrigeration. RNA is generally less stable than DNA and prone to degradation and integrity of biomarker transcripts during sample shipping is of concern. Procedures to obtain RNA from Sterivex membrane filters are presented in **O** Chapter 26, Vol. 5, Part 2.

#### 10 Research Needs

To advance microbial groundwater monitoring, the integrity of the target biomarkers (DNA, RNA, proteins, fatty acids) and preservation methods should be explored in detail so that the loss of analyte(s) during sampling and sample handling can be quantified and appropriately considered for data interpretation. The search for process-specific biomarkers, including DNA (genes) RNA (transcripts), proteins, and fatty acids, must continue to enhance information richness and resolution of the analytical process. To reduce analysis costs, even with a broader

suite of biomarker targets, site-tailored multiplex qPCR assays that only enumerate biomarkers that are informative for a given site should replace the current "shotgun" approaches that probe all available biomarkers. Further, the contemporary biomass collection and sample handling procedures should be compared, and if needed, standardized to facilitate comparative analyses between sites. In addition, standard analytical procedures should be developed and applied so that data generated in different laboratories can be directly compared. Groundwater analysis for chemical analytes (i.e., contaminants) routinely uses internal standards to quantify losses and provide accurate quantitative measurements. Internal standards for microbial analysis are currently not available but the development of such internal standards would be a major advance for microbial biomarker quantification. Further, implementing standard analytical procedures would allow direct comparison of data generated in different laboratories. Since groundwater analysis for monitoring microbial processes in the subsurface requires that some fraction of the target microbes is planktonic, the factors controlling the switch from sessile (e.g., attached growth in biofilms) to a planktonic lifestyle must be explored. Unfortunately, the effects of changing geochemical conditions on the physiological status (ecophysiology) (e.g., attached vs. planktonic growth) of the target populations are poorly understood, which confounds quantitative assessment of groundwater samples and data interpretation. Addressing these research needs and technical challenges will advance microbial groundwater analysis to a routine procedure that informs about microbial processes of interest in aquifers.

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Appendix D: Passive Flux Meter Protocol Report

# Environmental Security Technology Certification Program (ESTCP)

# **Final Protocol Report**

# Field Demonstration and Validation of a New Device for Measuring Water and Solute Fluxes CU-0114



November 2005

University of Florida & Purdue University

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# List of Acronyms

ACGIHAmerican Conference of Governmental Industrial HygienistsCARcorrective action reportCFchloroformCMchloromethaneCPVCchlorinated poly (vinyl chloride)CVcoefficient of variationDMP2,4-dimethyl-3-pentanolDNAPLdense nonaqueous phase liquidDOdissolved oxygenDoDDepartment of DefenseEDelectrochemical detectorEPAEnvironmental Protection AgencyESTCPEnvironmental Security Technology Certification ProgramFIDflame-ionization detectorFRTRFederal Remediation Technology RoundtableFTLfield team leaderGCgas chromatographyHASPhealth and safety planHPLChigh pressure liquid chromatographyIBAisobutyl alcoholICion chromatographyI.D.inner diameterIDLinstrument detection limitIDLHimmediately dangerous to life or healthIPAisopropyl alcoholMDLminimum detection levelMCImethylene chlorideMLSmultilevel samplersMSmatrix spikeMSDmatrix spike duplicate
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ESTCP Environmental Security Technology Certification Program FID flame-ionization detector FRTR Federal Remediation Technology Roundtable FTL field team leader GC gas chromatography HASP health and safety plan HPLC high pressure liquid chromatography IBA isobutyl alcohol IC ion chromatography I.D. inner diameter IDL instrument detection limit IDLH immediately dangerous to life or health IPA isopropyl alcohol MDL minimum detection level MeCl methylene chloride MLS multilevel samplers MS matrix spike
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<ul> <li>HASP health and safety plan</li> <li>HPLC high pressure liquid chromatography</li> <li>IBA isobutyl alcohol</li> <li>IC ion chromatography</li> <li>I.D. inner diameter</li> <li>IDL instrument detection limit</li> <li>IDLH immediately dangerous to life or health</li> <li>IPA isopropyl alcohol</li> <li>MDL minimum detection level</li> <li>MeCl methylene chloride</li> <li>MLS multilevel samplers</li> <li>MS matrix spike</li> </ul>
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MLSmultilevel samplersMSmatrix spike
MS matrix spike
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MSD matrix spike duplicate
main spike aupieue
MSDS materials safety data sheets
MTBE methyl tertiary butyl ether
NAPL non-aqueous phase liquid
NBS national bureau of standards
NIOSH National Institute for Occupational Safety and Health
NITS National Institute of Standards and Testing
NITSNational Institute of Standards and TestingO.D.outer diameter
NITSNational Institute of Standards and TestingO.D.outer diameterOSHAOccupational Safety and Health Administration
NITSNational Institute of Standards and TestingO.D.outer diameter

PFM	Passive Fluxmeter
PPE	personal protective equipment
PSO	project safety officer
PVC	poly (vinyl chloride)
QAPP	quality assurance project plan
QA/QC	quality assurance/quality control
RCRA	Resource Conservation and Recovery Act
RPD	relative percent difference
RRF	relative response factors
RRT	relative retention times
SD	standard deviation
SOP	Standard operating procedure
SRM	Standard Reference Materials
SS	stainless steel
SSO	site safety officer
TBA	tert-butyl alcohol
TCE	trichloroethylene
TLV	threshold limit value
TWA	time weighted averages
VOA	volatile organic acid

# 1.0. PFM Construction, Storage, and Transport

## **1.1. Description of PFM**

The PFM is a self-contained permeable unit that is inserted into a well or boring such that it intercepts groundwater flow but does not retain it (See Figure 1-1).

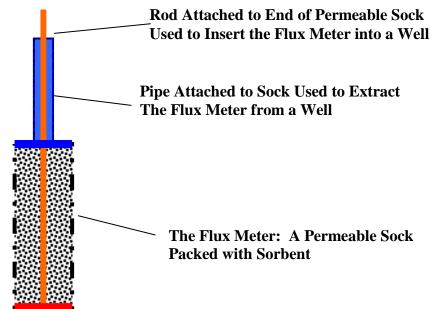


Figure 1-1. Schematic of a Flux meter comprised of a permeable sock filled with a selected sorbent.

The interior composition of the flux meter is a matrix of hydrophobic and hydrophilic permeable sorbents that retain dissolved organic and/or inorganic contaminants present in fluid intercepted by the unit. The sorbent matrix is also impregnated with known amounts of one or more fluid soluble 'resident tracers'. These tracers are leached from the sorbent at rates proportional to fluid flux.

After a specified period of exposure to groundwater flow, the flux meter is removed from the well or boring. Next, the sorbent is carefully extracted to quantify the mass of all contaminants intercepted by the flux meter and the residual masses of all resident tracers. The contaminants masses are used to calculate cumulative and time-averaged contaminant mass fluxes, while residual resident tracer masses are used to calculate cumulative or time- average fluid flux. Depth variations of both water and contaminant fluxes can be measured in an aquifer from a single flux meter by vertically segmenting the exposed sorbent packing, and analyzing for resident tracers and contaminants. Thus, at any specific well depth, an extraction from the locally exposed sorbent yields the mass of resident tracer remaining and the mass of contaminant intercepted. Note that multiple tracers with a range of partitioning coefficients are used to determine variability in groundwater flow with depth that could range over orders of magnitude. This data is used to estimate local cumulative water and contaminant fluxes.

#### 1.2. Preparation of Sorbent (Activated Carbon) and Tracers

Table 1-1 provides a complete list of equipment items need to prepare the PFM sorbent, while Table 1-2 lists all required parts. A tracer mixture is prepared by combining appropriate ratios of all tracers used for the test. Tracer volumes are measured in graduated cylinders and transferred to a volumetric flask for mixing. The flask is manually mixed. The sample mixture includes 100 ml of methanol, 100 ml of ethanol, 200 ml of isopropylalcohol (IPA), 200 ml of tert-butyl alcohol (TBA) and 66 ml of 2, 4-dimethyl-3-pentanol. A volume of the tracer mixture is transferred to a 22L plastic jar containing water (190 ml of tracer solution is added to 15 L of water). The jar cap is tightened with several layers of Teflon tape to provide a seal. This is mixed manually agitating occasionally over a period of a few hours until all immiscible liquids have dissolved.

Dry activated carbon (AC) is added to the aqueous solution containing tracers. 15 L of AC is slowly added to the 22L jar. Each jar is weighted before transfer. Significant gas is generated by adding the AC to water so this process requires addition of AC with gentle agitation until the AC is completely wet. Note a dust mask should be worn during this process. After all AC is added the jar is again sealed using several layers of Teflon tape.

The jars containing AC, water and tracers are mixed by rotating. The jars are placed in a 55 gallon drum and secured using foam packing. The 55 gallon drum is rotated or rolled for a period of 6 hours to homogenize the AC tracer mixture. Following mixing, the 22 L jars are placed in a cooler until shipping or packing.

## **1.3. Assembly of PFMs**

Table 1-1 provides a complete list of equipment items need to prepare for PFM assembly, while Table 1-2 lists all required parts. The passive flux meters are constructed in a pipe having the same diameter as the well screen. The exception is for stainless steel well screen nylon for which a red mesh material is added. In this case the pipe used for construction should be slightly (0.1 inch) smaller diameter than the well screen. The pipe length is 5 feet. A table is used as a work area for packing of the flux meters. All materials needed for packing the PFMs are listed in Section 3-0. Nitrile protective gloves and protective clothing are worn during construction.

Prior to packing the PFMs with AC, the sock is attached to the center tube of the PMF. The center tube for 2 inch wells consists of 1/2 inch CPVC pipe cut to 5 feet lengths. The bottom of the sock is clamped to the CPVC pipe using a SS worm drive clam (pipe band clamp). The sock is protected from the clamp by wrapping the CPVC pipe with electrical tape (4 wraps) prior to attaching the sock and between the sock and the SS clamp. This is important to avoid sharp edges of the clamp tearing the sock material. The sock and tube are then placed in the packing pipe and the top of the sock is pulled back over the outside of the packing pipe (note the ends of the packing pipe should be sanded to remove sharp edges capable of tearing the sock material). Prior to adding AC to the PFM a thick (1/8 inch) viton washer is inserted to the bottom of the sock. The viton washers used should have an OD of the packing pipe and a center hole the same

as the center tube (for 2 inch wells, 2 inch OD with a 5/8 inch hole). The viton washers are pushed to the bottom of the PFM using a 3/4 inch CPVC pipe.

Prior to packing, the top of the center tube must be plugged to avoid AC entering the center tube. Any method of plugging is appropriate (cork, rubber stopper, cap, electrical tape). At this point the PFM is ready for packing with AC. A funnel is used to drop the AC lifts into the PFM. The funnel should be cut to have an opening slightly smaller than the packing pipe. AC is transferred to a 400 ml beaker using a large spoon and the AC poured into the funnel attached to the top of the packing pipe. The funnel is tapped to slide the AC into the sock and packing pipe. A vibrator (or manual tapping) is applied to the packing pipe to help settle the AC to the bottom of the sock. After adding the required amount of AC (this will depend on the desired sampling interval), a thin (1/16 inch) viton washer is pushed down the sock to pack the AC in place. This process is continued until the sock is filled. During this period, an initial AC sample is collected and placed in a 40ml vial containing 20 ml of extraction solvent (IBA). For 5 foot long socks typically 4 to 6 lifts are packed. At the top of the PFM a thick viton washer is added followed by a sponge cut to the same size as the viton washer. The sponge is used to minimize AC loss from the top of the PFM since the connection between the top of the sock and the center tube must be loose.

At this point the top of the sock is attached to the retrieval wire and short section of PVC pipe. The outside of the PVC pipe section should be wrapped with electrical tape to protect the sock from the wire rope. The PVC pipe section is then slid over the center tube down to the position of the sponge. The sock is pulled up over this section. Electrical tape is applied over the sock attaching it to the PVC pipe section (4 or 5 wraps). This is critical to protect the sock from the SS clamp. A SS clamp is attached to the sock ensuring that the clamp is securely over the PVC pipe section and electrical tape is protecting the sock. This step is very critical since failure of this attachment will make PFM retrieval extremely difficult. This clamp should be well tightened but not stripping the worm drive clamp. Electrical tape is then wrapped around the outside of the clamp to prevent the clamp from catching on joints in the well screen and casing.

1/16 inch stainless steal wire rope is used to retrieve the PFMs. These need to be constructed prior to PFM assembly. The wires are fastened using wire crimps with a crimping tool. The wires are connected to the sock at the top of the PFM using a short section of PVC pipe (approximately 1.5 inch section of 3/4 inch PVC) with the wire looped through holes drilled in the short section of pipe. Four holes (1/8 inch) are used to thread the wire though the pipe and is crimped to the entry wire using two crimps. The other end of the wire is formed into a loop at the end of the wire either the length needed for deep deployment or a short section (6 inches) that is attached to a second wire rope using a coupler. The short wire is required if the PFM will be stored in tubes for storage and shipping.

The PFM construction process is visualized in pictures below.

- 1) Place empty PFM sock into PVC packing tube.

2) Take a  $C_0$  sample of the sorbent for each PFM created.





3) Scoop sorbent mixture into 400mL beaker. Pour sorbent mixture from beaker into PFM sock using a funnel placed inside the packing tube.





4) Slide a neoprene washer onto center tube after each one-foot segment of sorbent is poured into sock (for vertical separation of segments). Use pipe with a larger I.D. than the O.D. of the PFM center pipe to push the washer into the mesh sock.



- 5) Repeat steps 3 and 4 until all of the sorbent has been loaded into the PFM.
- 6) Attach retrieval mechanism to top of PFM



7) Push newly constructed PFM out of packing tube and into transport tube.



8) <u>Screw on Teflon lined end</u> caps to both ends of transport tube.



9) PFM is ready for transport.



10) Cut retrieval cable for each PFM to correct length. Use stainless steel compression sleeves to form loops at both ends. Label the wire with the appropriate PFM ID. For detailed description of wires construction see paragraph 1.3 below.



Equipment items and materials needed to construct and sample PFM are summarized in Table 1-1.

Preparing Activated Carbon and Tracers
Prepare tracer solution in water:
Graduated cylinders (50, 100, 500 ml, 1, 2 L)
1L volumetric flask
Glass funnel
Stir bars
Stir plate
Pipettes (10, 25, 50, 100 ml)
Pipette bulb
22L plastic jars (Cole-Parmer)
Adding AC to 22L jars:
2L plastic jars
Balance (2 decimal place up to 4Kg)
Dust masks
Teflon Tape (3/4 inch heavy duty)
Mixing AC:
55 gallon drum roller
55 gallon drum with removable top
Dense foam pieces to hold 22L jars in place
Preparing wire lines for PFMs:
Wire cutters
PVC pipe cutter
Pliers
Wire Crimper (McMaster Carr 3582T5 Multi Groove Hand Tool For All 3/64" Sleeves
& 3/32" Alum Oval Sleeves)

Table 1-1. Equipment Items Needed to Construct and Sample PFM(task based).

200 ft tape measure
Drill with 1/8in bit
Dremel tool
Constructing PFMs:
Nut drivers (5/16")
400 ml beakers
Work table
Constructing PFM carrying tubes:
Hack Saw
Sand paper
PVC glue
Teflon tape
Sampling PFMs
Preparing sample vials:
Balance
Syringe dispenser (for IBA) (Fisher item}
Sampling in the field:
Spatulas
Scissors
Mixing bowls
Buckets

All parts used for PFM construction and their costs are listed in Table 1-2.

Part	Supplier	Part	Cost
		Number	
Silver Impregnated Granular Activated Carbon	Eric R. Hasis Inside Sales Representative Site Services/Remediation Calgon Carbon Corporation PH: 800-422-7266 x 4770		50 – 200 lbs\$7.22/lb
	Fax 412-787-4523 website: www.calgoncarbon.com	15020 0025	<b>0</b>
IBA 2.5L	Fisher Scientific	15828-0025	\$95
40 ml VOA vials Max=300 (432) Ordered direct to site (222 left)	Fisher Scientific	03-339-14A	
60+ 5.5 ft long socks	Lili's Alterations		\$4/sock
330 feet of red mesh (only needed for SS wells)	Cole-Parmer	U-09405-30	\$95/164ft
<sup>1</sup> / <sub>2</sub> inch CPVC tube	Hardware Store		
Rubber Washers	Servalite	RT258	\$0.52 each

Table 1-2	PFM Parts List	10/3/05
1 abic 1-2.	I I WI I alto List	10/5/05

	1-800-477-6760	RM258	
Hose clamps	McMaster-Carr	54155K15	\$0.56
1	www.mcmaster.com	5415K32	\$0.39
Threaded rod for pushing	GeoProbe		\$500
wells in			
Wire lines.	McMaster-Carr	3461T9	Wire \$0.08/ft
	www.mcmaster.com	3883T39	
PFM Parts List (Need to have on site)			1
1.8" PVC pipe (to pack socks	Hardware Store		
in)			
3/4 in PVC pipe for packing	Hardware Store		
Funnels	Hardware Store		
Vibrator			
Power cord	Hardware Store		
Bucket for fluid leakage	Hardware Store		
Large spoon for transferring			
Calibrated jar	Fisher Scientific		
Spatula for Co sampling	Fisher Scientific	14-357	\$6.06
Electrical tape <b>order more</b>	Hardware Store		
Wrenches	Hardware Store		
Pipe cutter	Hardware Store		
Hack-saw	Hardware Store		
Tape measure	Hardware Store		
Balance on-site	Fisher Scientific		
Cooler			
Blue Ice			
Syringe dispenser	Fisher Scientific	<u>13-689-</u>	<u>\$274</u>
		<u>135D</u>	
Field notebook	UF Bookstore		
Gloves	Fisher Scientific		
Rope	Hardware Store		
Hard Hats			
Copper caps			
Safety vests			
Steel toe boots			
Methanol	Fisher Scientific	A452-4	\$174.41
Ethanol	Fisher Scientific		
IPA	Fisher Scientific		
ТВА	Fisher Scientific		
2,4-dimethyl-3-pentanol	Fisher Scientific		

#### 1.4. PFM Storage

Table 1-1 provides a complete list of equipment items need to prepare PFMs for storage, while Table 1-2 lists all required parts. If the PFMs are constructed for transport to the field site, the PFMs will be stored in tubes and cooled. PFM storage tubes are constructed using PVC pipe the same diameter as the packing tube. The ends have threaded caps that are sealed using Teflon tape. In the bottom of the storage tube a spacer is place to stop the PFM from sliding past the end of the PVC pipe (usually a gap exists between the pipe and end cap in which the PFM can expand during transport). A rubber stopper (#10) works well. The PFM is then extruded from the packing tube into the storage tube. A section of threaded rod or PVC pipe is used to push the PFM out of the packing tube and into the storage tube. The top of the storage tube is then sealed. The PFMs are then placed in cold storage (5 °C) until transport.

#### **1.5. PFM Transport**

Table 1-1 provides a complete list of equipment items need to prepare for PFM transport, while Table 1-2 lists all required parts. The PFMs are transported in insulated containers to the site. Cardboard boxes (5ft by 8x8in) with foam insulation (1 inch) forming the walls has been used for FedEx shipments. Blue ice is added to the box for cooling. For vehicle transport appropriate insulation for the travel time can be constructed.

# 2.0. PFM Deployment

#### 2.1. PFM Insertion

Table 1-1 provides a complete list of equipment items need for PFM deployment, while Table 1-2 lists all required parts. At the field site the PFM in the packing tube or storage tube is prepared for PFM insertion into the well casing. A wire rope is attached to the top of the PFM using a safety carabineer (or if packing on-site the required wire has been attached to the PFM). The tube is lined up with the top of the well casing and a section of push rod is used to push the PFM from the tube into the top section of well casing. Additional push rods are attached to continue pushing the PFM to the screen interval. If multiple PFMs are deployed in a single well, the wires from PFMs currently in place are held taught to avoid the wire catching on the PFM being inserted. When inserting the PFM some back pressure may build since the water in the well casing must flow through the center tube as the PFM is inserted. Proceed slowly is pressure builds. The flux meter steel cable attached to the sock assembly is then secured to an exterior 2" segment of PVC pipe to ensure that it will not be lost to the well head.

Step-by-step instructions are provided below.

1) Transport PFMs to site using PVC transport tubes (either by vehicle or FedEx Overnight). Once onsite, organize the flux meters for deployment.



2) Setup a workstation at the first deployment location by laying the PFMs for the first well across two portable sawhorses.



3) <u>Remove end caps from PVC transport tubes</u>.



4) Remove well lid and cap.





5) Attach retrieval cable to the top of each PFM.



6) Install PFM by setting the transport tube on top of the monitoring well casing and using Geoprobe rods to push the PFM out of the transport tube and into the well.



7) Push PFM into position in the well using Geoprobe rods while maintaining tension on the retrieval cable.



- 8) Repeat steps 5 through 7 for each PFM that is to be installed in the well.
- 9) Replace well lid and cap (wire cables are cut to a length such that two feet of each retrieval cable will remain outside the well).
- 10) Repeat steps 2 through 9 for each well.

# 2.2. PFM Retrieval and Sampling

Table 1-1 provides a complete list of equipment items need to retrieve and sample PFM, while Table 1-2 lists all required parts. PFMs are retrieved using the wire rope. The top PFM in the well is extracted first by gently pulling up on the wire (heavy work gloves should be worn when pulling on 1/16 inch cable). The PFM should be pulled to the top of the well casing. The PFM will occasionally catch on joints in the well screen. Simply apply more pressure to overcome. If the PFM will not move look at troubleshooting options below. When the PFM is at the top of the well casing untangle any wires that are twisted at the well head. Thread the retrieval cable through a 5' x 2" I.D. PVC pipe and place the pipe over the well to guide and contain the extruded PFM. Move the PFM to the sampling work station.

A tarpaulin acts as a 'protective flooring' for the work zone. A portable table is used as a work zone for sampling the PFMs. All material listed in Table 1 will be contained in this area during the retrieval stage. Nitrile protective gloves and necessary other protective clothing will be worn by all samplers. A lined bucket is placed under the work area to capture un-sampled residual activated carbon from the retrieved PFM. The sock is extruded from the PVC pipe to the sampling interval extent. The flexible mesh and cotton packing materials are cut and the sorbent captured in plastic bowls for homogenization using a stainless steel spatula. A sub-sample is then transferred into 40 mL VOA vials each containing the extraction solvent. 10 grams of sample or 1.5 cm sample depth is added to the vials. The vials are stored in a cooler containing blue ice prior to transport back to the laboratory for analysis. The center tube and viton washers are measured to obtain the sample interval lengths in the PFM. Sampling materials, spatula, scissors, mixing bowls are wiped cleaned of AC grains prior to retrieval of the next PFM.

40 ml VOA vials are used for AC sampling. The vials are weighed empty (nearest 0.01 g) and recorded. The vials are then filled with IBA (extraction solvent) using a fixed volume dispenser (20 ml syringe dispenser) and sealed. The vials weights are then recorded. Following addition of AC (approximately 2cm depth) the vials are weighed.

The PFM retrieval and sampling procedure is visualized in pictures below.

1) Retrieve PFM from well by pulling up on the attached wire cable. The PFM is pulled from the well pipe directly into a PVC tube of the same diameter.



2) Place tube on table and expose the first segment by pulling on the bottom end of the PFM.



3) Using scissors, cut open the nylon mesh covering the first segment and pour the exposed sorbent into a bowl.



- 4) Stir the sorbent to homogenize
- 5) <u>Sub-sample the mixture and place into 40mL vial containing IBA</u>



- 6) Measure the interval length of the PFM segment
- 7) Repeat for steps 3-8 for remaining segments of PFM
- 8) After all PFMs are sampled, place 40mL vials into cooler(s) and ship back for analysis
- 9) Excess sorbent is collected in a plastic-lined container for proper hazardous waste disposal.

All equipment items and materials needed for deployment and retrieval of PFM in the field are listed in Table 2-1.

Item	Used During Deployment	Used During Retrieval
Field notebook	X	X
Cotton Flux Socks	Х	
Stainless Steel Rods	Х	
Flexible Plastic mesh	Х	
Plastic funnel	Х	
Tracer loaded Activated Carbon	Х	
Rubber washers	Х	
Steel-wire retrieval cable	Х	
Medium threaded stainless steel clamps	Х	
Small stainless steel threaded clamps	Х	
Electrical tape	Х	X
Spoon/scoop	Х	
Paper towels	Х	X
Clean cloth	Х	X
Garbage bags	Х	X
20L Carboy containing Deionized water	Х	X
Portable workbench	Х	X
Tarpaulin	Х	X
Latex gloves	Х	X
Protective work gloves	Х	X
Wire crimper	Х	
40 mL VOA vials containing 20mL solvent	Х	X
1" I.D. PVC packing rod	Х	
1.8" I.D. PVC packing pipe	Х	
2" I.D. PVC Transporting pipe	Х	X
4' Plastic attachable insertion rods	Х	
Tape measure		X
Stainless steel spatulas		X
Tool box	Х	X
Plastic homogenizing bowls		X
Scissors	Х	X
Bucket	Х	X
Alconox solution	Х	X
20L Carboy for liquid waste	Х	X
Nitrile protective gloves	Х	X

Table 2-1. Field Equipment and Materials

## 2.3. Troubleshooting PFM extraction

In the event that the PFM is difficult to remove from the well the following steps might be considered. Using the rods used to insert the PFM, push down to move the PFM below the obstruction. In this case it is useful to attach a viton 2 inch washer at the end of the push rod to center the rod in the well. Holding both the retrieval wire and the push rod, surge the PFM up and down to attempt to overcome the obstacle.

In the event that the wire breaks or becomes detached from the PFM, a corkscrew attachment can be added to the rod to attempt to "grab" the top of the PFM and advance it upwards. If this fails the corkscrew can be used to dig into the AC and viton washers again in an attempt to "grab" the PFM. Finally, a pump with tubing lowered to the top of the PFM can be used to extract the AC. This slow process obviously destroys the PFM, but can be successful in clearing the well.

# 3.0 Analytical Methods Supporting the Experimental Design and Sampling Plan

Details of, or references to, the analytical methods employed in sampling and analysis to determine the results of the application (i.e. performance) of the technology.

# **3.1.** Standard operating procedure for extraction of analytes from flux device sorbents (October 10, 2001)

# 3.1.1. Scope and application

1. This SOP describes the procedures used by the Department of Environmental Engineering Sciences, University of Florida, for extraction of target analytes (including tracers) from sorbents used in flux devices inserted in monitoring wells.

2. This SOP was written by M.D. Annable, Department of Environmental Engineering Sciences, University of Florida, Gainesville, FL.

3. The selected constituents are TCE, PCE, and alcohol tracers:

Methanol Ethanol 2-propanol (IPA) 2-methyl-1-propanol (IBA) 2-methyl-2-propanol (TBA) n-propanol n-butanol n-pentanol n-hexanol n-heptanol 3-heptanol n-octanol 2-octanol 2,4-dimethyl-3-pentanol 2-ethyl-1-hexanol 3,5,5-trimethyl-1-hexanol 6-methyl-2-heptanol 2,6-dimethyl-2-heptanol n-decane

Potential Sorbents include:

Liquid (mixed in a sand matrix at a pore volume saturation of 10%) Tetradecane Heptadecane Hexadecane

Solid Activated Carbon Surfactant modified zeolytes

4. The method involves liquid extraction in 20 or 40 ml VOA vials using organic solvents.

#### 3.1.2. Purpose

The purpose of this SOP is to insure reliable and reproducible analytical results. Extracted constituents will be quantified suing analytical methods described in other SOPs.

#### 3.1.3. Procedures

1. Sample Containers, Collection, Transportation and Storage

<u>Sample Containers</u>: Field samples will be collected in 20-mL or 40-ml glass sample vials (Fisher Catalog # 03-340-121) with teflon-faced rubber backed caps.

<u>Sample Collection</u>: Each field sample vial will be partially filled with the extraction solvent (alcohol IPA, IBA, etc. or Methylenechloride) using a pipet or repeating volume dispenser. Typically 10 or 20-ml of solvent will be used.

<u>Transportation and Storage</u>: Field samples will be stored in coolers containing "blue ice", and later stored in refrigerators in a trailer located on the site. Samples will be sent to UF labs packed in coolers and shipped via overnight air express (e.g., FedEx). The samples will be stored in the cold

storage room or refrigerator at 4C, until GC analysis. After sub-sampling, the samples are returned to cold storage.

For lab studies, samples will be collected directly in 20 mL Headspace vials whenever possible and stored in a refrigerator if analysis is expected to take more than a day.

2. In the laboratory, samples will be rotated for a minimum of 8 hours on a rotator (Glas-Col model RD 4512).

#### 3. Sub-sampling and Dilution

Field samples will be sub-sampled into 2 ml GC vials. Pipets will be used to transfer samples from 20-mL sample vials to the 2-mL GC vials.

3. Apparatus and Materials

Glassware: Glass pipets are required for sub-sampling.

4. Safety

Gloves and eye protection will be worn during all extraction activities.

Reference to the Materials Safety Data Sheets (MSDS) will be made for information on toxicity, flammability, and other hazard data.

# **3.2.** Standard operating procedure for analysis of alcohol tracers (November 15, 1995) utilized at the University of Florida

#### 3.2.1. Scope and application

1. This SOP describes the analytical procedures utilized by the Soil and Water Science Department, University of Florida, IFAS, for analysis of alcohols used as partitioning tracers in both lab and field studies in order to quantify the amount and distribution of residual non-aqueous phase liquids (NAPLs) present in the saturated zone.

2. This SOP was written by R.D. Rhue, Soil and Water Science Department, University of Florida, Gainesville, Fl. It is a modification of SOP-UF-Hill-95-07-0010-v.2, prepared by D.P. Dai, H.K. Kim, and P.S.C. Rao, Soil and Water Science Department, University of Florida. The SOP of Dai, Kim, and Rao was modified from a protocol provided to them by Professor Gary Pope at the University of Texas-Austin.

3. The alcohol tracers used in the UF lab and field studies are ethanol, n-butanol, n-pentanol, n-hexanol, n-heptanol, 2,2-dimethyl-3-pentanol, and 6-methyl-2-heptanol.

4. The method involves gas chromatography (GC) analysis for alcohol concentrations in aqueous samples. A flame-ionization detector (FID) is used to quantify the analyte concentrations in the sample. The method has been found to provide reliable and reproducible quantitation of alcohols for concentrations > 1 ug/mL. This value may be considered the minimum detection level (MDL). The standard calibration curve for FID response has been found to be linear up to 3,000 ug/mL for ethanol.

5. Samples selected for GC-FID analysis may be chosen on the basis of preliminary screening which will provide approximate concentration ranges and appropriate sample injection volumes, standard concentrations, etc.

# 3.2.2. Purpose

The purpose of this SOP is to insure reliable and reproducible analytical results for alcohols in aqueous samples for laboratory-based or on-site (field-based) GC-FID analyses, and to permit tracing sources of error in analytical results.

# 3.2.3. Procedures

1. Sample Containers, Collection, Transportation and Storage

<u>Sample Containers</u>: Field samples will be collected in 5-mL glass sample vials (Fisher Catalog # 06-406-19F) with teflon-faced septa caps. Glass vials and caps are not reused.

<u>Sample Collection</u>: Each field sample vial will be completely filled with liquid, such that no gas headspace exists, and capped. The vials will not be opened until the time for analysis.

<u>Transportation and Storage</u>: Field samples will be stored in coolers containing "blue ice", and later stored in refrigerators in a trailer located on the site. Samples may be subjected to on-site GC analysis, and/or shipped back to UF labs; samples will be packed in coolers and shipped via overnight air express (e.g., FedEx). The samples will be stored in the cold storage room or refrigerator at 4C, until GC analysis. After sub-sampling, the samples are returned to cold storage.

For lab studies, samples will be collected directly in 2 mL GC vials whenever possible and stored in a refrigerator if analysis is expected to take more than a day.

2. Sub-sampling and Dilution

Field samples will be sub-sampled into 2-ml vials for automated GC analysis. Disposable, Pasture glass pipets (Fisher Catalog # 13-678-20B) will be used to transfer samples from 5-mL sample vials to the 2-mL GC vials.

For samples needing dilution prior to GC analysis, a dilution of 1:10 should be sufficient. Dilutions will be made using double-distilled, deionized water.

3. Apparatus and Materials

<u>Glassware</u>: Disposable micro-pipets (100 uL; Fisher Catalog # 21-175B; 21-175F) and Class A volumetric pipets (1 or 2 mL) are required for sample dilution.

Disposable Pasteur glass pipets (Fisher Catalog # 13-678-20B) are required for sub-sampling.

GC vials (2-mL) with Teflon-faced caps (Fisher Catalog # 03-375-16A) are required for GC analysis.

Volumetric class A pipets and volumetric class A flasks are required for preparations of the calibration standards.

<u>Gas Chromatograph System</u>: An analytical GC system with a temperature-programmable oven, auto-injector capable of on-column injection, and either an integrator or a PC-based data acquisition/analysis software system are required. Also required are other accessories, including analytical columns and the gases required for GC-FID operation.

A Perkin Elmer Autosystem with an FID and an integrated autosampler are suitable for analysis of field and laboratory samples. The Perkin Elmer system is linked to an IBM-compatible PC loaded with Turbochrom (version 4.01) software.

A J&W Scientific DB-624 capillary column (30m x 0.53mm, 3um film thickness) are required. Zero-grade air and ultra-high purity hydrogen are required for the FID. Ultra-high purity nitrogen or helium are required for carrier gas.

## 4. Reagents

<u>Deionized</u>, <u>Double-Distilled Water</u>: Deionized, double distilled water is prepared by double distillation of deionized water in a quartz still. This water will be referred to as reagent water.

<u>Alcohols</u>: Certified ACS grade alcohols are required for analysis.

## 5. Standard Solutions

<u>Stock Standard Solution</u>: Analytical standards are prepared from reagent chemicals in the laboratory. Stock standards each contain a single alcohol dissolved in reagent water and stored in 20 mL glass vials (Fisher Catalog # 03-393-D) with teflon-lined caps. These stock solutions are kept in a refrigerator at 4 C. Fresh stock standards are prepared every six months. The procedure for making stock standard solutions is essentially that given in the Federal Register, Rules and Regulations, Thursday, November 29, 1979, Part III, Appendix C, Section 5.10,

"Standard Stock Solutions". The only modification of the procedure for the current study is that reagent water is used as the solvent in place of methanol.

<u>Calibration Standards</u>: Calibration standards are prepared by diluting the stock standards in reagent water. Each calibration standard contains each of the alcohols listed above. Five concentrations should be prepared that cover the approximate concentration range utilized in the partitioning tracer experiments.

6. QC blank Spike/Matrix Spike

Two 1 mL aliquots of the sample to be spiked are transferred to clean vials. To one vial, 1 mL of reagent water is added. To the second vial, 1 mL of a calibration standard is added. The spike recovery is calculated using the difference between the two measured concentrations and the known spike concentration.

7. Quality Control

GC injector septa should be changed every 80 to 100 injections, or sooner if any related problems occur.

Injector liner should be cleaned or changed every 80 to 100 injections or sooner if any related problems occur.

A method blank should be included in every 50 samples

A complete set of calibration standards (5) should be run at the beginning of each day and after every fiftieth sample.

One standard and a blank should be included in every 25 samples.

A sample spike and a blank spike should be included in every 50 samples.

8. Instrumental Procedures

Gas Chromatography: For J&W DB-624 Column:

Injection port temperature200CFID detector temperature225C

Temp Program: Isothermal at 60C for 0 min; Ramp to 120C at 5 C/min.

9. Sample Preparation

<u>Sub-sampling</u>: Field samples are transferred from the 5 mL sample vials to the 2 mL GC vials and capped with open-top, teflon-lined septa caps.

<u>Dilution</u>: Samples are diluted if chromatographic peak areas for any of the alcohols exceed those of the highest calibration standard. One mL of sample is added to an appropriate amount of reagent water to make the dilution.

10. Sample Analysis

Analysis: The samples should be allowed to reach ambient temperature prior to GC analysis.

Sample vials (2 mL) are loaded onto the Perking Elmer GC auto-injector. A one uL injection volume should be used for both samples and standards.

<u>Analyte Identification</u>: Analyte identification should be based on absolute retention times. The analytes of interest should elute at their characteristic retention times within 0.1 minute for the automated GC system.

<u>Analyte Quantitation</u>: When an analyte has been identified, the concentration should be based on the peak area, which is converted to concentration using a standard calibration curve.

## 11. Interferences

Contamination by carry-over can occur whenever high-level and low-level samples are sequentially analyzed. To reduce carry over, the injector syringe should rinsed with reagent water between samples.

Potential carry-over should be checked by running a highly concentrated sample, but one still within the standard concentration range, followed by a blank. A negligible reading for the blank will insure that carry-over has been minimized.

## 12. Safety

The main safety issue concerning the use of the GC at a field site relates to the compressed gases. The FID gases (hydrogen and air) form explosive mixtures. It is important to keep this in mind at all times, and be aware of the hazard potential in the event of an undetected hydrogen leak. All gas connections will be properly leak tested at installation.

High-pressure compressed-gas cylinders will be secured to a firm mounting point, whether they are located internally or externally.

Gas cylinders should preferably be located outside the trailer on a flat, level base, and the gas lines run inside through a duct or window opening. If the gases are located outside, then some form of weatherproofing for the gauges will be necessary. As a temporary measure, heavy-duty polyethylene bags, secured with tie-wraps, have been used successfully; this may not be very elegant but it is very effective for short-term use of the GC. A more permanent protective housing must be built if the GC is located at the trailer for an extended time period.

The main operating drawback to locating the gas cylinders externally is that it is not easy to monitor the cylinder contents from inside. The gas which could be used up most quickly is air for the FID, particularly if two instruments are hooked up to the same supply and they are running continuously. A reserve cylinder of air should be available at all times to prevent down time.

If it is not possible to arrange external citing easily, the gas cylinders should be secured to a wall inside the trailer.

It is a good laboratory operating practice to make sure the flame is attended at all times.

When it is necessary to change the injection liner on the GC, the detector gases should be shut off.

The column must be connected to the detector before igniting the flame.

The trailer should be kept well ventilated when using the GC.

Reference to the Materials Safety Data Sheets (MSDS) will be made for information on toxicity, flammability, and other hazard data.

## **3.3. Standard operating procedure for the sampling, collection, extraction and analysis of Alcohol Tracers utilized at Purdue University**

The following described Standard Operating Procedures (SOP) are currently utilized by the Environmental Engineering area of the School of Civil Engineering at Purdue University, West Lafayette, Indiana.

This SOP was updated March 3, 2004 by I.C. Poyer, School of Civil Engineering, Purdue University, West Lafayette, Indiana. It is a modification of SOP-UF-Hill-95-07-0010-v.2, prepared by D.P. Dai, H.K. Kim, and P.S.C. Rao, Soil and Water Science Department, University of Florida.

#### 3.3.1. Scope and application

1. This Standard Operating Procedure (SOP) describes the extraction and analytical procedures of alcohol tracers from a sorbent (Silver-impregnated Activated Carbon) packed into borehole flux meters. Some of the alcohols have been used as partitioning tracers in both laboratory and field studies to quantify the amount and distribution of DNAPLs in source zones. Here, these alcohols are used as "resident" tracers that are pre-loaded on to the sorbent packed into the flux

meter sock; loss of tracers via desorption and advective/diffusive/dispersive transport resulting from groundwater flow under natural hydraulic gradients is measured to estimate cumulative groundwater and contaminant fluxes.

2. The alcohol tracers used in the Purdue University field studies are methanol, ethanol, iso-propanol, t-butanol, and 2,4-dimethyl-3-pentanol.

3. The established analytical method to determine and quantify alcohol concentrations in extracted samples is direct injection of 1  $\mu$ L of sample into a Shimadzu GC17A gas chromatograph (GC) equipped with a flame-ionization detector (FID). This method provides reliable and reproducible quantitation of alcohols at concentrations greater than or equal to 1  $\mu$ g/mL, which is the reportable minimum detection limit (MDL). The linear standard calibration range for the FID response is from the reported MDL up to a concentration of approximately 1500  $\mu$ g/mL per analyte of interest.

# 3.3.2. Purpose

The purpose of this SOP is to: (1) insure reliable and reproducible results, and (2) track possible sources of error in the extraction of alcohols from a sorbent and the subsequent analysis by GC-FID analytical methodology.

# 3.3.3. Procedures

1. Sample Containers, Collection, Transportation and Storage

Sample Containers: Sorbent samples should be collected in 40mL VOA vials (Fisher Scientific Catalog # 05-719-106) sealed with Teflon-lined septa caps. Vials should contain 20mL of extraction solvent (*iso*-butanol), prepared previously in the laboratory. All vials and caps are non-reusable.

Sample Collection: Sorbent aliquots collected over 1-foot increments from an exfiltrated Flux meter should be transferred to a mixing bowl and homogenized with a metal spatula. Approximately 10 to 20 grams of mixed sorbent should be placed into the 40 mL VOA vials containing extracting solvent.

Transportation and Storage: Field samples should be stored, on site, in coolers containing "blue ice" then shipped via overnight air express (e.g., FedEx) to the Purdue University laboratory. Samples should be stored in a cold storage room or refrigerator at 4° C until extraction and GC analysis.

2. Laboratory Supplies and Materials

Volumetric class 'A' pipettes and volumetric class 'A' flasks for preparation of calibration standards and sample dilutions.

Disposable Pasteur glass pipettes (Fisher Catalog # 13-678-6A) for sub-sampling.

GC vials (2 mL) with Teflon-faced caps (Fisher Catalog # 03-375-16A) for GC analysis.

3. Reagents

Deionized water prepared by filtration of potable water through a Barnstead Ultrapure Deionization Unit. This water should be referred to as 'reagent water'.

Certified ACS grade pure alcohols purchased from one or more of the following vendors; Fisher Scientific, VWR and/or Sigma-Aldrich and used as received.

4. Calibration and Stock Standard Solutions

Individual alcohol stock standard solutions should be prepared in reagent water using volumetric glassware and stored in 20 mL glass vials with Teflon-lined caps. Stock solutions should be kept in a refrigerator at 4° C. Fresh stock standards should be prepared every six months and follow protocols outlined in the Federal Register, Rules and Regulations, Thursday, November 29, 1979, Part III, Appendix C, Section 5.10, "Standard Stock Solutions". The single modification from the cited procedure is the use of reagent water instead of methanol as the solvent.

Mixed calibration standards should be prepared by diluting stock standards in reagent water using volumetric glassware. A minimum of five standards should be prepared and should bracket the expected concentration range.

5. Quality Control (QC) Blank Spike/Matrix Spike

A blank spike should be prepared by the addition of 1 mL of calibration standard to 1mL of extraction solvent. A matrix spike should be prepared by the addition of 1mL of calibration standard to 1 mL of extracted sample. Spike recoveries should be calculated using the difference between the two measured concentrations and the known spike concentration.

6. Analytical Instrumentation

A Shimadzu GC17A Gas Chromatograph equipped with an AOC17 Autosampler, a temperatureprogrammable oven, heated auto-injector and detector zones, a 30 meter or greater capillary separations column, nitrogen carrier gas, standard compressed air and hydrogen flame gases and controlled by a PC-based data acquisition/analysis software system.

7. GC Procedure

Column dimensionsJ&W DB-624 Column, 75m X 0.53um X 3umInjection port temperature180CFID detector temperature220CColumn Temperature ProgramIsothermal at 60C for 3 min; ramp to 120C at 5 C/min, hold1min; ramp at 20C/min to 200C, hold 1 min.Nitrogen 99.995% purityFlame gasesAir, 99.995% purity; Hydrogen, 99.995% purity

#### 8. Quality Control of GC System

GC injector septa should be changed every 100 to 150 injections, or sooner if instrument performance deteriorates.

Injection port glass liner should be cleaned or changed after 100 to 150 injections or sooner if instrument performance deteriorates.

A method blank should be analyzed at the beginning of each sample set and after every 25 samples to monitor instrument background.

A complete set of calibration standards (minimum 5) should be analyzed at the beginning of each day with a mid-range continuing calibration standard analyzed after every 25 samples.

A matrix spike and a blank spike, and up to 5 sample duplicates should be analyzed with each daily sample set.

#### 10. Extraction of Alcohol Tracers from Sorbent Matrix

The collected sorbent samples should be rotated for a period not to exceed 24 hours on a Glas-Col Rotator, centrifuged for 5 minutes at 2000 rpm (Jouan, Inc., centrifuge), and sub-sampled into a 2 mL vial for GC analysis. Extraction vials will be stored at 4° C.

#### 11. Sample Analysis

Individual alcohol identification should be based on absolute retention times compared to calibration standards.

Alcohol concentrations should be calculated on chromatographic peak area response converted to units of concentration in mg/L based on standard calibration curves.

#### 12. Interferences

Contamination by carry-over may occur when high-level and low-level samples are sequentially analyzed. Subsequent dilution and reanalysis should be completed on samples identified as

outside the standard concentration bracket. Samples analyzed immediately following a 'high-concentration sample' should be reanalyzed.

In an attempt to minimize carryover, samples suspected of being in a higher concentration range should be isolated and bracketed by the analysis of reagent water samples.

# 13. Safety

Reference to the Materials Safety Data Sheets (MSDS) should be made for information on toxicity, flammability, and other hazard data.

# **3.4.** Standard operating procedure for analysis of Target Analytes in groundwater samples (February 20, 1996)

#### 3.4.1. Scope and application

1. This SOP describes the analytical procedures utilized by the Department of Environmental Engineering Sciences, University of Florida, for analysis of target analytes in groundwater samples from both lab and field studies. This analysis provides characterization of existing site and lab column aqueous contamination both before and following flushing technology applications.

2. This SOP was written by M.D. Annable, Department of Environmental Engineering Sciences, University of Florida, Gainesville, FL. It is a modification of SOP-UF-Hill-95-07-0012-v.2, prepared by D.P. Dai and P.S.C. Rao, Soil and Water Science Department, University of Florida.

3. The selected constituents are benzene, toluene, o-xylene, 1,1,1-trichloroethane, 1,3,5,-trimethylbenzene, 1,2-dichlorobenzene, decane, and naphthalene.

4. The method involves gas chromatography (GC) analysis for target analyte concentrations in aqueous samples. Headspace analysis with a flame-ionization detector (FID) is used to quantify the analyte concentrations in the sample. The method has been found to provide reliable and reproducible quantitation of the above constituents for concentrations > 5 ug/L. This value may be considered the minimum detection level (MDL).

5. Samples selected for GC-FID analysis may be chosen on the basis of preliminary screening which will provide approximate concentration ranges and appropriate sample injection times, and standard concentrations, etc.

#### 3.4.2. Purpose

The purpose of this SOP is to insure reliable and reproducible analytical results for soluble NAPL constituents in aqueous samples for laboratory-based GC-FID analyses, and to permit tracing sources of error in analytical results.

#### 3.4.3. Procedures

1. Sample Containers, Collection, Transportation and Storage

<u>Sample Containers</u>: Field samples will be collected in 20-mL glass sample vials (Fisher Catalog # 03-340-121) with teflon-faced rubber backed caps. Glass vials and caps are not reused.

<u>Sample Collection</u>: Each field sample vial will be completely filled with liquid, such that no gas headspace exists, and capped. The vials will not be opened until the time for analysis.

<u>Transportation and Storage</u>: Field samples will be stored in coolers containing "blue ice", and later stored in refrigerators in a trailer located on the site. Samples will be sent to UF labs packed in coolers and shipped via overnight air express (e.g., FedEx). The samples will be stored in the cold storage room or refrigerator at 4C, until GC analysis. After sub-sampling, the samples are returned to cold storage.

For lab studies, samples will be collected directly in 20 mL Headspace vials whenever possible and stored in a refrigerator if analysis is expected to take more than a day.

2. Sub-sampling and Dilution

Field samples will be sub-sampled placing 10-ml into 20-ml headspace vials containing 2 g of sodium chloride for automated GC analysis. Pipets will be used to transfer samples from 20-mL sample vials to the 20-mL GC headspace vials.

3. Apparatus and Materials

Glassware: Glass pipets are required for sub-sampling.

GC headspace vials (20-mL) with Teflon-faced caps are required for GC analysis.

Volumetric class A pipets and volumetric class A flasks are required for preparations of the calibration standards

<u>Gas Chromatograph System</u>: An analytical GC system with a temperature-programmable oven, headspace sample injection system, and either an integrator or a PC-based data acquisition/analysis software system are required. Also required are other accessories, including analytical columns and the gases required for GC-FID operation.

A Perkin Elmer Autosystems with an HS40 Auto-headspace sampler and a FID will be used for analysis of field and laboratory samples. The Perkin Elmer system will be linked to an IBM-compatible PC loaded with Turbochrom (version 4.01) software.

A J&W Scientific DB-624 capillary column (50m X 0.53mm,  $3\Box$ m film thickness) will be used. Zero-grade air and high purity hydrogen will be used for the FID. Ultra-high purity nitrogen or helium will be used for carrier gas.

4. Reagents

<u>Deionized</u>, <u>Double-Distilled Water</u>: Deionized, double distilled water is prepared by double distillation of deionized water in a quartz still. This water will be referred to as reagent water.

#### 5. Standard Solutions

<u>Stock Standard Solution</u>: Analytical standards will be prepared from reagent chemicals by the laboratory. Stock standards will each contain a single analyte dissolved in methanol and stored in 20 mL glass vials (Fisher Catalog # 03-393-D) with teflon-lined caps. These stock solutions will be kept in a refrigerator at 4 C. Fresh stock standards will be prepared every six months. The procedure for making stock standard solutions is essentially that given in the Federal Register, Rules and Regulations, Thursday, November 29, 1979, Part III, Appendix C, Section 5.10, "Standard Stock Solutions".

<u>Calibration Standards</u>: Calibration standards will be prepared by diluting the stock standards in water. Each calibration standard will contain each of the eight analytes listed above. Five concentrations will be prepared that cover the approximate concentration range from 0 to 20 mg/L.

6. QC blank Spike/Matrix Spike

Two 1 mL aliquots of the sample to be spiked will be transferred to clean vials. To one vial, 1 mL of reagent water will be added. To the second vial, 1 mL of a calibration standard will be added. The spike recovery will be calculated using the difference between the two measured concentrations and the known spike concentration.

7. Quality Control

A method blank will be included in every 50 samples

A complete set of calibration standards (5) will be run at the beginning of each day and after every fiftieth sample.

One standard and a blank will be included in every 25 samples.

A sample spike and a blank spike will be included in every 50 samples.

8. Instrumental Procedures

Gas Chromatography: For J&W DB-624 Column:

Headspace sample temperature90CInjection needle temperature100CTransfer line Temperature110CFID detector temperature225CCarrier gas pressure8psi

Temp Program: Isothermal at 50C for 0 min; Ramp to 200C at 5 C/min; hold for 10 min.

9. Sample Preparation

<u>Sub-sampling</u>: Field samples will be transferred from the 20 mL sample vials to the 20 mL GC headspace vials and capped with open-top, teflon-lined septa caps.

<u>Dilution</u>: Samples will be diluted if chromatographic peak areas for any of the analytes exceed those of the highest calibration standard. One mL of sample will be added to an appropriate amount of reagent water to make the dilution.

#### 10. Sample Analysis

<u>Analysis</u>: Sample headspace vials (20 mL) will be loaded onto the Perking Elmer HS40 autosampler. Samples will be pressurized for 1 min followed by a 0.1 minute injection time and a withdrawal time of 0.5 minute.

<u>Analyte Identification</u>: Analyte identification will be based on absolute retention times. The analytes of interest should elute at their characteristic retention times within  $\pm 0.1$  minute for the automated GC system.

<u>Analyte Quantitation</u>: When an analyte has been identified, the concentration will be based on the peak area, which is converted to concentration using a standard calibration curve.

#### 11. Interferences

Contamination by carry-over can occur whenever high-level and low-level samples are sequentially analyzed. To reduce carry over, the injector needle should purged with carrier gas between samples.

Potential carry-over will be checked by running a highly concentrated sample, but one still within the standard concentration range, followed by a blank. A negligible reading for the blank will insure that carry-over has been minimized.

#### 12. Safety

The main safety issue concerning the use of the GC relates to the compressed gases. The FID gases (hydrogen and air) form explosive mixtures. It is important to keep this in mind at all times, and be aware of the hazard potential in the event of an undetected hydrogen leak. All gas connections will be properly leak tested at installation.

High-pressure compressed-gas cylinders will be secured to a firm mounting point, whether they are located internally or externally.

When it is necessary to change the injection liner on the GC, the detector gases should be shut off.

The column must be connected to the detector before igniting the flame.

Reference to the Materials Safety Data Sheets (MSDS) will be made for information on toxicity, flammability, and other hazard data.

# **3.5.** Standard operating procedure for the sampling, collection, extraction and analysis of Perchlorate from sorbents packed in borehole flux meters

# 3.5.1. Scope and application

1. This Standard Operating Procedure (SOP) describes the extraction and analytical procedures of perchlorate from sorbent (Silver-impregnated Activated Carbon) packed into the borehole flux meters. The mass of perchlorate accumulated by sorption on the sorbent from the groundwater passing through the flux meter is used to estimate the cumulative contaminant flux.

2. The established analytical method to determine and quantify perchlorate concentrations in extracted samples is direct injection of 1  $\mu$ L of sample into a Dionex DX600 Ion Chromatograph equipped with an Electrochemical Detector (ED). This method provides reliable and reproducible quantitation of perchlorate at concentrations greater than or equal to 2  $\mu$ g/L, which is the reportable minimum detection limit (MDL). The linear standard calibration range for the ED response is from the reported MDL up to a concentration of approximately 100 mg/L for the analyte of interest.

## 3.5.2. Purpose

The purpose of this SOP is to: (1) insure reliable and reproducible results, and (2) track possible sources of error in the extraction of perchlorate from a sorbent and the subsequent analysis by IC-ED analytical methodology.

## 3.5.3. Procedures

1. Sample Containers, Collection, Transportation and Storage

Sample Containers: Field samples will be collected in 250mL wide-mouth jars, sealed with Teflon-lined septa caps.

Sample Collection: Sorbent aliquots collected over 1 foot increments from an exfiltrated Flux meter, will be transferred to a mixing bowl and homogenized with a metal spatula. Approximately 100 grams of mixed sorbent will be placed into the wide-mouth jar. Excess sorbent will be collected in a plastic-lined container for proper hazardous waste disposal.

Transportation and Storage: Sorbent samples will be stored, on site, in coolers containing "blue ice" then shipped via overnight air express (e.g., FedEx) to the Purdue University laboratory. Samples will be stored in a cold storage room or refrigerator at 4°C until extraction and IC-ED analysis.

2. Laboratory Supplies and Materials

Volumetric class 'A' pipets and volumetric class 'A' flasks for preparations of calibration standards and sample dilutions.

Disposable Pasteur glass pipets (Fisher Catalog # 13-678-6A) for sub-sampling.

IC vials (2mL) with Teflon-faced caps (Fisher Catalog # 03-375-16A) for IC analysis.

3. Reagents

Deionized water prepared by filtration of potable water through a Barnstead Ultrapure Deionization Unit. This water will be referred to as reagent water.

Certified ACS grade granular ammonium perchlorate purchased from Sigma-Aldrich.

4. Calibration and Stock Standard Solutions

A stock standard solution will be prepared in reagent water using volumetric glassware and stored in 20mL glass vials with Teflon-lined caps. The stock solution will be refrigerated at 4°C. Two concentration ranges will be prepared. The higher concentration range will be 100mg/L to 1mg/L. The low concentration range will be 2ug/L to 100ug/L. A minimum of five standards per range will be prepared.

5. Quality Control (QC) Blank Spike/Matrix Spike

A blank spike will be prepared by the addition of 1 mL of calibration standard to 1mL of reagent water. A matrix spike will be prepared by the addition of 1mL of calibration standard to 1 mL of sample. Spike recoveries will be calculated using the difference between the two measured concentrations and the known spike concentration.

6. Analytical Instrumentation

A Dionex DX600 Ion Chromatograph (IC) Autosystem equipped with an ED50 Electrochemical Detector, a GP50 Gradient Pump, a GD40 Eluent Generator, an AS50 Thermal Compartment, and an AS50 Autosampler will be used for analysis of all perchlorate samples. The Dionex IC system is linked to an IBM-compatible PC loaded with Peaknet (version 6.00) software for acquisition, analysis interpretation and quantitation.

A Dionex IonPac AS11 column and guard column will be used and the analyte perchlorate eluted with 35mM potassium hydroxide solution.

7. IC Parameters and Analytical Conditions

Analytical & Guard Column	Dionex IonPac AS11, 4mm
Column temperature	30C
Suppressor Current	104 mV
Eluent Concentration	35mM potassium hydroxide
Column flow rate	1.2 mL/min
Inightion loop volume 50vl (high our	agentration range), 050, 1 (low concentration range)

Injection loop volume 50ul (high concentration range); 950ul (low concentration range)

## 8. Quality Control of IC System

Nanopure water is used to provide ion-free solvent for the Eluent Generator and eliminate high background signal

A method blank will be analyzed at the beginning of each sample set and after every 25 samples to monitor instrument background.

A complete set of calibration standards (minimum 5) will be analyzed at the beginning of each day with a mid-range continuing calibration standard analyzed after every 25 samples.

A matrix spike and a blank spike, and up to 5 sample duplicates will be analyzed with each daily sample set.

9. Extraction of Perchlorate from Sorbent Matrix

Perchlorate extraction from the sorbent will be completed utilizing a Dionex ASE300 Accelerated Solvent Extractor, with hot reagent water as the solvent. Glass fiber filters and Ottowa 40 mesh sand will be used to filter and as a filler respectively, in the extraction cell.

#### 10. Sample Analysis

Perchlorate identification will be based on the absolute retention time compared to calibration standards.

Perchlorate concentrations will be calculated on a chromatographic peak area response converted to units of concentration in ug/L or mg/L based on the standard calibration range of analysis.

## 11. Interferences

Contamination by carry-over may occur when high-level and low-level samples are sequentially analyzed. Subsequent dilution and reanalysis will be completed on samples identified as outside the standard concentration bracket. Samples analyzed immediately following a 'high-concentration sample' will be reanalyzed.

In an attempt to minimize carryover, samples suspected of being in a higher concentration range will be isolated and bracketed by the analysis of reagent water samples.

12. Safety

Reference to the Materials Safety Data Sheets (MSDS) will be made for information on toxicity, flammability, and other hazard data.

# 4.0. Quality Assurance Project Plan

# 4.1. Purpose and Scope of the Plan

This Quality Assurance plan focuses on field installation, sampling and processing of data from the Flux Meters.

# 4.2. Quality Assurance Responsibilities

The responsibility for QA were shared by Kirk Hatfield and Mike Annable at the University of Florida. During field activities one of the PI's will be present to oversee QA procedures. Other personnel present during field sampling activities will include graduate students or post-doctoral researchers from the University of Florida, Purdue University, and the University of Waterloo.

# 4.3. Data Quality Parameters

This section discusses measures to be taken to ensure the representativeness, completeness, comparability, accuracy, and precision of the data.

# Accuracy

Accuracy is defined as the closeness of the results to the true value.

The percent recoveries of surrogates, QC check standards, and matrix-spiked analytes are used to evaluate the accuracy of an analysis. The percent recovery represented by X can be calculated using the following equations:

1

For surrogates and QC check standards:

$$X = \frac{SSR}{SA} \times 100$$

For matrix spikes:

$$X = \frac{SSR - SS}{SA} \ge 100$$

where:

SSR = Spiked sample result SS = Sample result SA = Spike added from spiking mix

The mean percent recovery (X) is defined by:

$$\overline{X} = \frac{\sum_{i=1}^{N} X_i}{N}$$
 2

where:

X<sub>i</sub> = The percent recovery value of a spike replicate N = Number of spikes

## **Precision**

Precision is a measure of the mutual agreement among individual measurements of the same parameters under prescribed similar conditions.

The analytical precision is determined using results from duplicate or replicate analyses of samples and from matrix spike results for a given matrix. The Relative Percent Difference (RPD) is used to evaluate the precision of duplicate analyses. Relative Percent Difference is defined in the following equation:

$$\% RPD = \frac{2(X1 - X2)}{\overline{x}} \times 100$$
 3

X1 = First duplicate value X2 = Second duplicate value

When replicate analyses are performed, precision is measured in terms of the Standard Deviation (SD) which is defined in the following equation:

$$S = \sum_{i=1}^{N} \left[ \frac{(X_i - \overline{X})^2}{N - I} \right]^{\%}$$

where:

X<sub>i</sub> = The recovery value of a spike replicate

X = Arithmetic average of the replicate values

N = Number of spikes

# **Completeness**

Completeness is defined as the percent of parameters falling within acceptance criteria and the results subsequently reported. A goal of 95 percent completeness has been set for all samples.

The general requirement of this quality assurance program is to analyze a sufficient number of standards, replicates, blanks, and spike samples to evaluate results adequately against numerical QA objectives.

# 4.4. Calibration Procedures, Quality Control Checks, and Corrective Action

The focus of the following section is to describe initial and continuing calibration procedures for analytical instrumentation, duplicate and control testing and data reduction, validation, and reporting.

# **Supplies and Quality Control Materials**

All supplies (i.e., glassware, chemicals, reagents) used will be of the best possible quality to ensure proper instrument calibration and avoid contamination. All reagents used are prepared from Analytical Reagent Grade (AR) chemicals or higher purity grades, unless such purity is not available. The preparation of all reagents will be documented, including source, mass, and dilutions. Each reagent will be clearly labeled with the composition, concentration, date prepared, initials of preparer, expiration date, and special storage requirements, if any.

# **Reagents**

Reagent solutions are stored in appropriate glass, plastic, or metal containers. Reagents are stored under conditions designed to maintain their integrity (refrigerated, dark, etc.). Shelf life is listed on the label and the reagent is discarded after it has expired. Dry reagents such as sodium sulfate, silica gel, alumina, and glass wool are either muffled at 400°C or extracted with solvent before use for organic chemical analyses. Water used in the laboratory is glass distilled or deionized, and periodically checked for purity. In addition, water used in the organics area is carbon-filtered or purchased as HPLC grade. All organic solvents used are either glass-distilled or pesticide grade. Solvents and reagent solutions are checked for contamination by employing reagent blanks, before use in any analysis.

# **Quality Control Reference Materials**

All Quality Control Reference Materials are acquired only from authorized vendors or sources commonly used by U.S. EPA Regional Laboratories.

# **Standards Traceability**

When standard reference materials arrive at the laboratory, they are registered in a bound log book, "Standards Notebook for Neat Materials and Primary Solutions." An example of a logging sequence is used to illustrate this process.

(1-S-XXX-12-4) (label and log sequence)

where:

1	=	Notebook log number
S	=	Standard Notebook"Neat and Primary Standards"
XXX		<ul> <li>Receiving analyst's initials</li> </ul>
12	=	Notebook page
4	=	Entry number on notebook page

All working standards prepared at the site lab are logged in the "Standards Notebook for Intermediate and Working Standards." A similar labeling convention has been adopted for classifying these working standard materials. An example is given below.

1-W-XXX-6-5 (label and log)

Where:

- 1 = Number of notebook
- W = Standards notebook "Intermediate and Working" Standard
- XXX = Analyst's initial
- 6 = Page Number
- 5 = Page entry number in sequence

#### **Instrument Calibration**

Every instrument used to analyze samples must pass the calibration criteria established in the appropriate SOP. Initial calibration criteria for instrument linearity, sensitivity, resolution, and deactivation must be met before samples can be analyzed. Sustained performance is monitored periodically during sample analyses by the use of continuing calibration check standards.

## **GC Section**

#### **Initial Calibration**

The linear calibration range of the instrument must be determined before the analysis of any samples. Gas chromatographic conditions used for sample analyses are used during calibration.

The calibration is performed in accordance with the SOP derived from the methods used. For most GC analyses, a 5-level calibration is run. The concentrations of the standards must bracket the linear range of the instrument. Calibration using fewer than 5-levels is done only when specifically allowed by the method.

#### **Relative Retention Times and Relative Response Factors**

Instrument calibration and sample analysis must be performed using appropriate internal standards to establish relative retention times (RRT) and relative response factors (RRF) where required. Internal standards appearing in a chromatogram will establish primary search windows for those target compounds nearby in the chromatogram. RRT are calculated using this equation:

$$RRT = \frac{RT^{target}}{RT^{is}}$$
 5

The RRF may be calculated as follows:

Absolute Response Factor = RF = AreaAmount

Note: <u>Amount</u> in this equation refers to the mass (e.g. ug) of compound mixed into the solution injected.

Each calibration standard is analyzed and the RRF is calculated for each analyte according to the following equation:

$$RRF = \frac{A_s \times C_{is}}{A_{is} \times C_s}$$

$$A_s = \text{Area of analyte}$$

$$A_{is} = \text{Area of internal standard}$$

 $C_{is}$  = Concentration of internal standard

 $C_s$  = Concentration of analyte

Note: Certain data processors may calculate the RRF differently.

The standard deviation (SD) and the % coefficient of variation (CV) of RRFs for the compounds are calculated using the following equations:

$$S = \sum_{i=1}^{N} \left[ \frac{\left( RRF_{i} - RRF_{m} \right)^{2}}{N - 1} \right]^{\%}$$
 7

Where:

= Number of RRFs

and

Ν

$$\%CV = \frac{S \times 100}{RRF_m}$$

# **Coefficient of Variation**

The %CV of each compound must be less than 30 percent. This criterion must be achieved for the calibration to be valid.

If the %CV is less than 20 percent, the RRF of the compound can be assumed to be invariant, and the average RRF can be used for calculations.

If the %CV is between 20 percent and 30 percent, calculations must be made from the calibration curve. Both the slope and the intercept of the curve must be used to perform calculations.

# **Initial Calibration Verification**

The calibration curve must be validated further by analyzing a QC check sample. The QC check sample must be obtained from EPA, another vendor, or it must be from another lot number. The QC check sample verifies the validity of the concentrations of the standards used to obtain the initial calibration.

All analytes in the QC check standard must be recovered within 80 to 100 percent. If any analyte exceeds this criterion, then a new calibration curve must be established. All sample results for a target analyte can be reported only from valid initial calibrations.

## **Continuing Calibration**

The working calibration curve or RRF for each analyte must be verified daily by the analysis of a continuing calibration standard. The ongoing daily continuing calibration must be compared to the initial calibration curve to verify that the operation of the measurement system is in control.

The continuing calibration check must be performed during each day of analysis to verify the continuing calibration of the instrument. A day is defined as 24 hours from the start run time of the last valid continuing calibration. Generally, a continuing calibration check sample is injected every 10 samples.

Verification of continuing calibration is performed by the analysis of a midpoint standard containing all of the analytes of interest. Verification of continuing calibration of the measurement system is done by calculating the percent difference (%D) of the continuing calibration RRF from the mean RRF from the initial calibration curve using the following equation:

$$\%D = \frac{(RRF_m - RRF) \times 100}{RRF_m} \qquad 9$$

where:

$RRF_m =$	The mean relative response factor from the initial calibration curve
RRF =	The relative response factor from the continuing calibration standard

The %D must meet the acceptance criteria established in the appropriate SOP. If these criteria are exceeded, a new calibration curve must be established.

## **Other Calibrations**

Weekly calibrations are performed for equipment such as balances, thermometers, ovens, incubators, and dissolved oxygen (D.O.) meters that are required in analytical methods, but which are not recorded in a dedicated QA instrument log.

#### **Balances**

Balances are checked with Class S weights on a daily basis. Before a weighing session, the analyst is required to perform at least one calibration check in the range of the material to be weighed. This value is also recorded on the specific balance control chart and must be within the control limit. The criteria for calibration checks are given in Table 4.1.

#### Table 4-1.

# **CRITERIA FOR BALANCE CALIBRATION CHECKS**

Analytical Balances		
Class S Weight	Warning Level	Control Level
(grams)	(grams)	(grams)
0.0100	0.0098-0.0102	0.0097-0.0103
0.1000	0.098-0.102	0.097-0.103
1.000	0.995-1.005	0.990-1.010
10.000	9.995-10.005	9.990-10.010
50.00	49.98-50.02	49.95-50.05
Top Loading Balance	ces	
1.00	0.95-1.05	0.90-1.10
10.0	9.9-10.1	9.8-10.2
50.0	49.7-50.3	49.5-50.5

#### Incubators, ovens, and waterbaths

Temperatures are checked daily with an NBS grade thermometer and necessary adjustments made as required. All temperature readings are recorded and posted on the appropriate equipment.

# **DO meters**

DO meter is calculated daily using a modified Winkler technique. The Winkler solution is titrated against 0.025N sodium thiosulfate.

# **Conductivity bridges**

Conductivity meter is standardized daily against a solution of KCl to obtain a new cell constant.

# pH meters

The pH meter is standardized daily using buffers at pH of 4, 7, and 10.

# Refrigerators

Refrigerators are maintained at 4°C, with control levels ranging from 1°C to 10°C. A temperature reading is taken each workday morning immediately after unlocking the refrigerator. The temperature reading is recorded and entered on the control chart posted on the door of the refrigerator. If a trend is apparent or if the temperature is outside the acceptable range, the Lab Manager is notified so that corrective action can be initiated if required.

## Freezers

Freezers are maintained at -10°C, with control levels ranging from 0°C to -35°C. A temperature reading is taken each workday morning immediately after unlocking the freezer. The temperature reading is recorded and entered on the control chart posted on the door of the freezer. If a trend is apparent, or if the temperature is outside the acceptable range, the Lab Manager is notified so that corrective action can be initiated if required.

## **Calibration Standards**

All calibration standards, including internal standards used in LMG, are obtained from chemical suppliers with certificates of high purity and concentration.

# Traceability

All standards are traceable to the National Institue of Standards and Testing (NITS) Standard Reference Materials (SRM) or to the U.S. EPA Reference Standards.

## **Working Standards**

The commercial standards are used as stock standards. Working standards are made from the stock standards at appropriate concentrations to cover the linear range of the calibration curve. The working standards are used for initial calibration curves, continuing calibration checks, and preparation of analyte spiking solutions as appropriate for a particular analysis. All stock and working solutions are uniquely identified, dated, labeled, and initialed.

## **Standards Logbook**

All stock solutions are given a unique code number and are entered into a bound "Primary Standards" logbook. The name of the compound and other pertinent information, including concentration, date of receipt, and analyst's name, are also entered.

Working standards are given a unique code number that allows them to be traced to a specific stock solution. The working standard is entered in a "Working Standards" logbook with analyst's name, date and method of preparation, and other pertinent information.

## **CORRECTIVE ACTIONS**

## Laboratory Imposed

Corrective actions will be initiated if the quality control criteria indicate an analysis is out of control.

- Check calculations for accuracy
- Check instrumentation to ensure it is operating properly. Recalibrate if necessary.
- Remake standards and reagents and reanalyze samples.
- Re-prep and re-analyze samples.

The analyst is responsible for initiating corrective actions for analytical problems encountered during analysis of samples. Most problems which occur and are corrected during the analytical run will be explained in the run log or analytical bench sheet for that run. A corrective action report (CAR) may be necessary for some problems encountered, such as complete system failure, chronic calibration failure, or severe matrix interferences.

During data review, the reviewer may initiate corrective actions based on problems or questions arising from the review. A CAR will be initiated.

The Laboratory Manager may initiate corrective actions if a problem is noticed during a QC review of data, a system audit, or a performance audit. A CAR will be initiated.

CARs are signed and dated by Project Manager, and by the Laboratory Manager. CARs will be filed in appropriate department files and in the Lab Manger's files.

# **Agency Imposed**

Any actions deemed necessary by regulatory agencies, such as EPA, will be taken. These actions are most likely to arise from a systems or performance audit, or from data review conducted by the agency.

# **Corrective Action Reports**

The field laboratory will have a Corrective Action System that ensures the proper documentation and dispositions of conditions requiring corrective action. The system will also ensure that the proper corrective action is implemented to prevent recurrence of the condition.

## **Situations Requiring Corrective Action Reports**

The Corrective Action System applies to all situations that affect data quality. These situations include, but are not limited to, quality control criteria being exceeded, statistically out-of-control events, deviations from normally expected results, suspect data, deviations from the standard operating procedure, and special sample handling requirements. Corrective actions may also be initiated as a result of other QA activities, such as performance audits, systems audits, laboratory/interfield comparison studies, and QA project-related requirements of certifying agencies such as EPA.

## **Corrective Action Procedures**

The procedure requires documenting the condition requiring corrective action on a Corrective Action Report and implementing corrective action based on the results of the investigation performed to determine the cause of the condition (Tables 4-2 and 4-3).

When a condition requiring corrective action arises, the Corrective Action Report is initiated. The initiator describes the condition requiring corrective action. An investigation, if necessary, is conducted to determine the cause of the condition. A corrective action is recommended based on the results of the investigation. The Corrective Action Report is reviewed by the Project Manager and the Field Site Manager who either approve the recommended corrective action or indicate a different corrective action. The originator has the responsibility of following up to be sure that the corrective action is implemented. Implementation of the corrective action is documented by the Corrective Action Report being signed and dated by the person who implemented the corrective action.

Table 4-2. Corrective	Actions	
QC Activity	Acceptance Criteria	Recommended Corrective Action
Initial instrument blank	Instrument response <mdl response<="" td=""><td>Prepare another blank, if same response, determine cause of contamination: reagents, environment, instrument equipment failure, etc.</td></mdl>	Prepare another blank, if same response, determine cause of contamination: reagents, environment, instrument equipment failure, etc.
Initial calibration standards	Coefficient of variation >0.99995 or standard concentration value <u>+</u> 10% of expected value	Reanalyze standards. If still unacceptable, then remake standards
QC Check Standard	$\pm$ 10% of expected value	Reanalyze standard. if still unacceptable, then remake standards, or use new primary standards if necessary
Continuing calibration Standards	$\pm$ of expected value	Reanalyze standard. If still unacceptable, then recalibrate and rerun samples from the last cc stnd. Check
Method blank	<mdl< td=""><td>Reanalyze blank. If still positive, determine source of contamination. If necessary, reprocess (i.e., digest or extract) sample set</td></mdl<>	Reanalyze blank. If still positive, determine source of contamination. If necessary, reprocess (i.e., digest or extract) sample set
Initial calibration Standards (GC/MS)	RRF <30%	Reanalyze standards. If still unacceptable, prepare new standards.
Surrogate recovery (GC/MS Semivolatiles)	0 or 1 outside CLP criteria	Re-extract and/or re-analyze
Surrogate recovery (GC/MS volatiles)	0 outside criteria	Re-analyze

Table 4-3. Corrective Ac	tion Report Criteria for Control Charts
Criteria	Corrective Action
A point outside $\pm 3$ standard deviations	Attempt to determine the source of the problem. Verbally report the deviation and results of preliminary investigation to the Field Site Manager, who will decide jointly what action to take. After implementing corrective action, complete the Corrective Action Report and submit it to the Project Manager and the Field Site Manager for approval.
Three consecutive points accuracy outside $\pm$	Conduct investigation. Check accuracy of data input, calculations, instrument, standards, etc., to locate the source
standard deviation	of the problem. Document results in a Corrective Action Report. Have the report approved by the supervisor. No results can be reported until the Corrective Action Report has been approved. Send a copy of the Corrective Action Report and a copy of the QC chart to the Field Site Manager.
Obvious outlier.	Conduct investigation. Check accuracy of data input, calculations, dilutions, instrument, standard, etc present initial findings to the Field Site Manager. They will jointly decide what actions need to be taken. Document the results in a Corrective Action Report and have it approved by the Field Site Manager. No results can be reported until the Corrective Action Report is approved. Send a copy of the Corrective Action report and a copy of the control chart to the Field Site Manager.
Obvious shift in the mean.	Conduct investigation. Check calculations, data entry, standards, instrument, calibrations, etc. Document results in a Corrective Action Report. Have the Corrective Action Report approved by the Field Site Manager. No results can be reported until the report is approved. Send a copy of the Corrective Action Report and a copy of the QC chart to the Field Site Manager.

# **4.5. Demonstration Procedures**

Initiating the flux meter experiments will involve limited field effort. All of the components of the device can be prepared prior to field activities. In the field, the primary activity will be assembly of the flux meters which can be completed with two people in a mater of minutes. Extraction and sub-sampling also required fairly minimal time and personnel. Only the controlled flow flume experiments will require establishing steady flow from one end of the flume using peristaltic pumps. These pumps will be calibrated in the field using simple time and volume measurements. Periodic flow measurements will be made to determine total average flow.

Samples collected at the Borden site will be sent to the University of Florida for analysis. In the laboratory, instrument maintenance will include the following.

# **Maintenance Schedule**

Preventive maintenance, such as lubrication, source cleaning, and detector cleaning, is performed according to the procedures delineated in the manufacturer's instrument manuals.

The frequency of preventive maintenance varies with different instruments. Routine maintenance performed includes cleaning and/or replacement of various instrument components. In general, the frequency recommended by the manufacturer is followed. In addition to the regular schedule, maintenance is performed as needed. Precision and accuracy data are examined for trends and excursions beyond control limits to determine evidence of instrument malfunction. Maintenance is performed when an instrument begins to degrade as evidenced by the degradation of peak resolution, shift in calibration curves, decreased ion sensitivity, or failure to meet one or another of the quality control criteria. Table 4-4 lists routine equipment maintenance procedures and frequency.

Instrument maintenance logbooks are maintained in the laboratory at all times. The logbook contains a complete history of past maintenance, both routine and nonroutine. The nature of work performed, the date, and the signature of the person who performed the work are recorded in the logbook. Preventive maintenance is scheduled according to each manufacturer's recommendation. Instrument downtime is minimized by keeping adequate supplies of all expendable items on hand. Expendable items are those with an expected lifetime of less than one year. Routine instrument preventive maintenance is handled by the instrument operator. Repair maintenance is performed by a full-time electronics technician, or by the manufacturer's service personnel.

Table 4-4. PREV	Table 4-4. PREVENTIVE MAINTENANCE							
Instrument	Activity	Frequency						
Gas Chromatograph	Change septum	As needed						
	Check carrier gas	Daily						
	Change carrier gas	As needed						
	Change in-line filters	As needed						
	Perform ECD wipe test	As license requires						
	Clean ECO	Return to vendor as needed						
	Check system for leaks	As needed						
	Clean/replace injection point liner	As needed						
	Clean/replace jet tip	As needed						
	Service flame photomeric detector	As needed						
IR	Change desiccant	Every six months						
	Electronics maintenance	Every six months						
UV	Clean and align optics	Annually						
	Replace lamp	As needed						
	Calibrate	Weekly						
pH Meter	Calibrate	Daily						
	Check fluid in probe	Daily						
D.O. Meter	Clean and replace membrane and	Daily						
	HCl solution							
	Calibrate	Daily						
Balance	Calibrate	Daily						
	Maintenance	Annually						
Ovens	Temperature checks	Daily						
Refrigerators and	Temperature checks	Daily						
Freezers								
COD Heating	Check temperature with NBS	As needed						
Block	thermometer							
Conductivity Meter	Standardize with KCl	Daily						
	Check probe visually	Daily						

# 4.6. Calculation of Data Quality Indicators

The focus of this section is to present methods of calculating data quality that will be used for this project.

# **Control Samples**

The laboratory will employ control samples to assess the validity of the analytical results of the field samples. Determination of the validity of field sample results is based on the acceptance criteria being met by the control sample. The acceptance criteria for each type of control sample

are delineated in the appropriate SOP. These acceptance criteria are based on the laboratory's statistical process capabilities determined from historical data, and meet the EPA CLP acceptance criteria as a minimum. Often, in-house criteria are more stringent than required by CLP. The control samples are analyzed in the same manner as the field samples. They are interspersed with the field samples at frequencies that are specified by the appropriate SOP.

## **Method Blank Analyses**

A method blank is a "clean" sample (i.e., containing no analyte of concern), most often deionized water, to which all reagents are added and analytical procedures are performed. Method blanks are analyzed at a rate of one per sample lot or at least every 20 samples. The blank is analyzed in order to assess possible contamination from the laboratory or the procedure. If the analyte of interest is found in the blank at above reporting levels, inorganic analysis is suspended until the source of contamination is found and corrective action is taken. The Laboratory Manager is notified when blank results are unacceptably high, and may assist in the investigation.

## Surrogate Spike Analyses

For certain analyses such as those performed by GC/MS, each sample and blank is spiked with one or more surrogate compounds before preparatory operations such as purging or extraction. These surrogate standards are chosen for properties similar to sample analytes of interest, but are usually absent from the natural sample.

Surrogate spikes evaluate the efficiency of the analytical procedure in recovering the true amount of a known compound.

The results of surrogate standard determinations are compared with the true values spiked into the sample matrix prior to extraction and analysis, and the percent recoveries of the surrogate standards are determined. Recoveries should meet the upper and lower control limits as specified for each compound. If control limits are exceeded for surrogate standards, the following sequence of actions is taken:

a. The sample is re-injected.

b. Raw data and calculations are checked for errors.

c. Internal standards and surrogate spiking solutions are checked for degradation, contamination, or solvent evaporation.

d. Instrument performance is checked.

e. If a, b, and c fail to reveal the cause of the noncompliance surrogate recoveries, the sample is re-purged or re-extracted.

f. If all the measures listed above fail to correct the problem for laboratory blank surrogate analyses, the analytical system is considered out of control, and the instrument must be recalibrated and examined for mechanical faults.

g. If all the measures listed above fail to correct the problem for field sample surrogate analyses, the deficiency probably is due to sample interferences, and not due to any procedural or mechanical problems in the laboratory. The surrogate spike recovery data and the sample data from both extractions are reported and are flagged. The Laboratory Manager is notified with an exceptions report and the corrective actions taken.

# Matrix Spike/Matrix Spike Duplicate Analyses

To evaluate the effect of the sample matrix on the analytical methodology, two separate aliquot samples may be spiked with a standard mix of compounds appropriate to a given analysis. The matrix spike and the matrix spike duplicate (MS/MSD) are analyzed at a frequency of one per lot or one per 20 samples, whichever is more frequent. The percent recovery for each of the spiking compounds is calculated. The relative percent difference (RPD) between the MS/MSD is also calculated.

The observed percent recoveries (%R) and relative percent differences (RPD) between the MS/MSD are used to determine the accuracy and the precision of the analytical method for the sample matrix. If the percent recovery and RPD results exceed the control limits as specified for each spiking compound, the sample is not reanalyzed. Poor recovery in matrix spiked samples does not necessarily represent an analytical system out of control. It is possible that unavoidable interferences and matrix effects from the sample itself preclude efficient recoveries. The poor recovery is documented for the Project Manager.

## **Internal Standards Analysis**

Once an instrument has been calibrated, it is necessary to confirm periodically that the analytical system remains in calibration. The continuing calibration and precision of the organics analytical system are checked for each sample analysis by monitoring the instrument response to internal standards. When internal standard addition is not appropriate to a particular method, other means of accuracy checks, such as standard addition, are used. Results from internal standard analyses are compared to the mean calibrated value. Deviation from this mean beyond a predetermined magnitude, depending on the type of analysis, defines an out-of-control condition. The system must then be brought back into control by:

- Checking the quality of the internal standards and reanalyzing the sample
- Recalibrating the system
- Correcting the malfunctions causing the instrument to fall out of calibration

# **Duplicate Sample Analyses**

Duplicate analyses are performed for cations analyses and upon special request for selected other parameters to evaluate the reproducibility of the method. Results of the duplicate analyses are used to determine the RPD between replicate samples. For each parameter analyzed, at least one duplicate sample is run per group of 20 samples.

The precision value, RPD, is reviewed by the section supervisor and the division manager. If the precision value exceeds the control limit or the established protocol criteria for the given parameter, the sample set is reanalyzed for the parameter in question unless it is determined that heterogeneity of the sample has caused the high RPD.

# **QC Check Standard Analyses**

Analysis of QC check standards is used to verify the preparation process or the standard curve, and is performed with each group of samples. Results of these data are summarized, evaluated, and presented to the section supervisor and the division manager for review.

The results of the QC check standard analysis are compared with the true values, and the percent recovery of the check standard is calculated. If correction of a procedure or instrument repair is done, the check standard is reanalyzed to demonstrate that the corrective action has been successful.

At least twice a year, a QC check standard for each parameter group is analyzed as a doubleblind sample. Samples are prepared, submitted, and evaluated by the Laboratory Manager.

# **Other Quality Control Samples**

Under some sampling analysis, additional quality control samples may be required. These may include:

a. **Blank/Spike--**Analyte of interest or surrogate is spiked into blank water rather than into a sample. The blank/spike goes through the entire analytical procedure, and percent recovery is calculated with no likelihood of matrix effect. For many contracts, an externally provided LCS sample (EPA) serves as a blank/spike sample.

b. **Trip Blank--**A sample bottle filled with laboratory blank water travels with the sample kit to the sampling site, and is sent back to the laboratory packed in the same container as any volatile samples collected. Trip blank analyses check for possible volatile contamination during shipping or sampling.

c. **Field Blank--**A field blank can be a sample container filled with laboratory blank water and sent to the sampling site, or it may be filled at the site with purchased distilled water or

decontamination water. The field blank analysis checks for possible contamination by the sampling team.

d. **Equipment Rinsates--**After equipment has been cleaned in the field, many contracts require that the equipment be rinsed and the rinsate analyzed for the same parameters requested on the samples. The rinsate analysis proves the equipment has been cleaned properly and will not contaminate the next samples taken.

# **Control Charts**

The laboratory will use control charts to monitor for out-of-control conditions.

# **Control Charting Process**

The control chart program uses a series of Lotus (or equivalent) macros to perform data processing and control charting. These macros also perform statistical decisions on the acceptability of the data.

The control chart used is a variation of the Shewart control chart of averages. The chart plots individual quantitative results against the order of time measurement. The plotted values are compared with control limits determined by the variability about the mean of the standard "in control" process. The control chart estimates the process mean and the variability from a moving window of 50 to 200 samples, depending upon the analytical parameters involved. The mean is estimated from the arithmetic average of the samples in the current window. The variability is estimated as the sample SD of the sample values in the current window. The program calculates the 2 SD and the 3 SD limits and displays them on the chart. The t-statistic is used to estimate the 99.7 percent tolerance limits for the degrees of freedom in the current window. Values outside the t-statistic limits are unconditionally rejected from inclusion in the sample window and automatically documented in a Corrective Action Report (CAR). The CAR prompts the analyst to initiate investigation and corrective action.

When the maximum number of samples has accumulated in the current window, the summary statistics of the mean and SD are written to the long-term data base. The last 20 samples in the old window are then transferred to a new window for continued use in the charting process.

The long-term data base charts the mean 1 SD error bars.

## Instrument Detection Limits, Method Detection Limits, and Reporting Limits

#### **Instrument Detection Limits (IDL)**

Instrument Detection Limit (IDL) studies are performed for inorganic parameters when an instrument is installed, when major maintenance or repair work has been done, and routinely once per calendar quarter.

To determine IDL, seven consecutive measurements per day are made on a prepared standard solution (in reagent water) of an analyte at a concentration 3 to 5 times the instrument manufacturer's suggested IDL. Each measurement is performed as though it were a separate analytical sample. This procedure is repeated on three nonconsecutive days. The standard deviation is calculated for each set of seven replicates and the average of the standard deviations is obtained. This average is multiplied by 3 to give the instrument detection limit (IDL).

## Method Detection Limits (MDL)

The Method Detection Limit (MDL) is the minimum concentration of a substance that can be measured and reported with 99 percent confidence that the value is above zero. The sample must be carried through the entire method under ideal conditions. MDL is determined according to the method outlined in 40 CFR 136, Appendix B. MDLs are determined at least annually for all parameters. MDL studies are also conducted for new methods introduced in the lab, after major maintenance or modification to an instrument, and as part of the training of new analysts.

To determine MDL, seven replicate analyses are made of analytes spiked into blank water at 1 to 5 times the estimated method detection limit. The spiked samples must be carried through the entire analytical procedure, including any extraction, digestion, or distillation process, for MDL calculation. The SD of these replicates is calculated. Where:t = The student t value for a 99% confidence interval

 $MDL = t \ x \ S \tag{10}$ 

S = Standard deviation of the replicate analyses

## **Reporting Limits**

In most cases, final report forms list reporting limits rather than either IDL or MDL. Reporting limits are taken from EPA SW846 published limits or from historical data. Matrixes or analyte concentrations which require dilution will change the detection limits for that sample.

#### 4.7. Performance and System Audits

In this section information is provided on performance audits and onsite system audits.

#### **Performance Evaluation Samples**

Performance evaluation samples are analyzed throughout the project for all parameters, as a constant check on accuracy and precision for all analyses.

#### Audits

Internal audits of the laboratory are conducted in two phases. The first phase is conducted by the Laboratory Quality Assurance Coordinator during the fourth quarter of the year. This is usually a 2-day systems audit which covers all sections of the laboratory. An audit report is issued within 2 weeks of completion. The Field Site Manager has the responsibility for coordinating all responses to the audit finding and for following up on the required corrective action. A followup audit is made when deemed necessary by the by the Field Site Manager or the Laboratory Manager. A quality assurance review questionnaire is provided in the Appendix.

The second phase consists of quarterly audits performed by the Field Site Manager. These are half-day or day-long audits, and are concentrated on specific areas that are deemed problem areas by the Field Site Manager. An audit report is issued at the completion of the audit. Responses and followup corrective action to the audit findings are required, and are monitored by the Field Site Manager.

All audit reports are issued to management and circulated to all staff. Copies are filed with the Field Site Manager and the Laboratory Manager.

# 4.8. Quality Assurance Reports

The performance of the field laboratory as assessed by the quality monitoring systems in place is reported by the Field Site Manager to management quarterly and as needed. Copies of all quality reports are maintained in the Field Site Manager and Laboratory Manager files.

Quality assurance reports to management include, but are not limited to, the following:

- Results of performance and systems audits
- Status of corrective actions
- Periodic assessment of data accuracy, precision, and completeness
- Significant QA problems and recommended solutions

In addition to the quarterly reports, a final report summarizing items covered in the quarterly reports is provided by the Field Site Manager to the Project Manager.

## 4.9. Data Format

## Introduction

In order to provide analytical data which is technically sound and defensible, a system of data management will be implemented in the laboratory. All activities which pertain to a sample are documented.

All data generated during the demonstration, except those that are generated by automated data collection systems, will be recorded directly, promptly, and legibly in ink. All data entries will be dated on the day of entry and signed or initialed by the person entering the data. Any change in entries will not obscure the original entry, will indicate the reason for such change, and will be dated and signed or identified at the time of the change.

In automated data collection systems, the individual responsible for direct data input will be identified at the time of data input. Any change in automated data entries will not obscure the original entry. Updated entries will indicate the reason for the change, the date, and the person responsible for making the change.

## **Data Tracking in the Laboratory**

The Field Site Manager is responsible for developing a system for tracking and maintaining sample identity between the collection point, analysis and reporting. This process will be periodically reviewed by the Project Manager.

## **Analyses and Data Reduction**

The Field Site Manager is responsible for the reduction of raw data when such steps are required to produce the correct data format for reporting. Data reduction may be done manually or through one of a number of computer programs used in the laboratory.

#### **Chromatogram Identification**

In the GC section computer software is used to identify chromatograms. A system-supplied file name (a hexadecimal date-time) and a user-supplied file name (related to an entry in the injection log) identify each acquisition.

#### **Data Reduction Formulas**

Linear regression formulas are used in a computer software system to calculate samples values for many general inorganic parameters and metals analyses. These programs use the general formula for linear regression:

11

$$Y' = a + bx$$

where:

Y' = The predicted value of y for a selected value of x

- = The value of y when x = 0a = The slope of the straight line
- b х
  - = Any value of x selected

Sample values for GC/MS parameters are calculated by systems software using the general formula:

$$\frac{Area_{Target} \quad x \; Amount_{IS}}{Area_{IS} \; x \; Response \; Factor}$$
12

GC data is calculated using either an internal or an external standard. For internal standards:

$$Concentration = \left(\frac{A_x^{sample}}{A_x^{standard}}\right) \left(\frac{A_{IS}^{standard}}{A_{IS}^{sample}}\right) \left(amt_x^{standard}\right) \left(\frac{P}{T}\right) \left(\frac{amt_{IS}^{sample}}{Amt_{IS}^{standard}}\right)$$
13

where: P = 1/fraction of extract to which IS is added

For calculations using an external standard:

$$Concentration = \left(\frac{A_x^{sample}}{A_x^{standard}}\right) \left(C_x^{standard}\right) \left(\frac{V}{T}\right)$$
 14

where: C

concentration of x in standard
 V = volume of final extract
 T = total sample extracted

## 4.10. Data Storage and Archiving Procedures.

Data from GC's will be saved and archived in P&E Turbochrom format. All data will be backedup on ZIP disks. This data will be batch processed into an Excel .csv file that can be easily converted to an Excel Worksheet. These files will be backed-up and transferred to individuals responsible for calculating flux results. All data related to the project will be organized for rapid retrieval and transfer to other interested parties. **Appendix E: Fines Calculations** 

Well Log	Тор	Bottom	Description	Fraction Fines	Measured Thickness	Weighted Fraction	Notes
Alameda							
PP41IW01	0	.0 4.	5 gravel				
	4	.5	water table				top of treatment zone
	4	.5 6.	0 gravel	0.025	1.5	0.038	
	6	.0 10.	0 clay	0.925	4.0	3.700	
	10	.0 15.	0 silt with sand	0.800	5.0	4.000	
	15	.0 27.	0 sandy silt	0.600	12.0	7.200	
	27	.0					bottom of treatment zone
	27	.0 35.	0 sandy clay				
			subtotal		22.5	14.938	
			average fraction fines			0.664	
			number of layers			4.000	
			layers per 10 feet			1.778	
			treatment zone thicknes	S		22.500	

Тор	Bottom	Description	Fraction Fines	Measured Thickness	Weighted Fraction	Notes
0.	0 1.	0 gravel with sand fill				
1.	0 2.	0 sand				
2.	0 3.	0 clayey sand				
3.	0 5.	0 sand				
5.	0 6.	0 silty sand				
6.	0 7.	5 sand				
7.	5	water table				top of treatment zone
7.	5 11.	3 sand	0.025	5 3.8	0.095	
11.	3 15.	0 sandy clay	0.600	) 3.7	2.220	
15.	0 20.	0 clay	0.925	5.0	4.625	
20.	0 22.	8 sandy clay	0.600	) 2.8	1.680	
22.	8 27.	5 silty sand	0.325	6 4.7	1.528	
27.	5 28.	0 gravel	0.025	0.5	0.013	
28.	0					bottom of treatment zone
28.	0 29.	0 sandy clay marl				
		subtotal		20.5	10.160	1
		average fraction fines			0.496	
		number of layers			6.000	
		layers per 10 feet			2.927	
		treatment zone thickness			20.500	
	0. 1. 2. 3. 5. 6. 7. 7. 11. 15. 20. 22. 27. 28.	$\begin{array}{cccccccccccccccccccccccccccccccccccc$	0.01.0 gravel with sand fill1.02.0 sand2.03.0 clayey sand3.05.0 sand5.06.0 silty sand6.07.5 sand7.511.3 sand11.315.0 sandy clay15.020.0 clay20.022.8 sandy clay22.827.5 silty sand27.528.0 gravel28.029.0 sandy clay marlsubtotal average fraction fines number of layers layers per 10 feet	Top         Bottom         Description         Fines           0.0         1.0 gravel with sand fill         1.0         2.0 sand         2.0 sand         2.0 sand         3.0 clayey sand         3.0 5.0 sand         5.2 sand         5.0 sand         5.2 sand         5.0 sand         5.2 sand         5.0 sand	Top         Bottom         Description         Fines         Thickness           0.0         1.0 gravel with sand fill         1.0         2.0 sand         2.0         3.0 clayey sand         3.0         5.0 sand         5.0         6.0 silty sand         6.0         7.5 sand         7.5         7.5 sand         7.5         3.0         7.5         9.0         7.5         3.0         7.5         7.5         3.0         7.5         7.5         7.5         3.0         7.5	Top         Bottom         Description         Fines         Thickness         Fraction           0.0         1.0 gravel with sand fill         1.0         2.0 sand         2.0         3.0 clayey sand         3.0         5.0 sand         5.0         6.0 silty sand         6.0         7.5 sand         7.5         water table         7.5         11.3 sand         0.025         3.8         0.095         11.3         15.0 sandy clay         0.600         3.7         2.220         15.0         20.0 clay         0.925         5.0         4.625         20.0         2.8 sandy clay         0.600         2.8         1.680           22.8         27.5 silty sand         0.325         4.7         1.528         27.5         28.0 gravel         0.025         0.5         0.013           28.0         29.0 sandy clay marl         subtotal         20.5         10.160         average fraction fines         0.496           number of layers         6.000         layers per 10 feet         2.927         10.160

Well Log	Тор	Bottom	Description	Fraction Fines	Measured Thickness	Weighted Fraction	Notes
Dover							
MW211D	0.0	1.	0 silty sand				
	1.0	10.	0 sand				
	10.0		water table				top of treatment zone
	10.0	17.	0 sand	0.025	7.0	0.175	
	17.0	19.	5 clay with sand	0.800	2.5	2.000	
	19.5	41.	5 sand	0.025	22.0	0.550	
	41.5	42.	5 clay	0.925	1.0	0.925	
	42.5	44.	0 clay with sand	0.800	1.5	1.200	
	46.5						bottom of treatment zone
	46.5	48.	0 clay				
			subtotal		34.0	4.850	
			average fraction fines			0.143	
			number of layers			5.000	
			layers per 10 feet			1.471	
			treatment zone thickness			36.500	

Well Log	Тор	Bottom	Description	Fraction Fines	Measured Thickness	Weighted Fraction	Notes
Moody							
FT07-DPS111-65	0.0		silty sand				
	2.7		sandy silt				
	5.0		sandy silt				
	10.0		sandy silt				
	15.0		sandy silt				
	15.5		water table	0.225	1.0	0.225	top of treatment zone
	15.5		silty sand	0.325			
	16.5 17.7		clay	0.925			
	20.0		sandy clay	0.600 0.325			
	20.0		silty sand	0.325			
	20.2		clay sandy silt	0.925			
	20.3		sandy clay	0.600			
	25.0		clay	0.925			
	25.4		sand	0.025			
	26.1		sandy clay	0.600			
	26.5		sandy silt	0.600			
	27.0		silty sand	0.325			
	27.2		clay	0.925			
	27.3		silty sand	0.325	0.1		
	27.4		clay	0.925	0.1	0.093	
	27.5	5 27.8	silty sand	0.325	0.3	0.098	
	27.8	3 27.9	clay	0.925	0.1	0.092	
	27.9	9 28.0	silty sand	0.325	0.1	0.033	
	28.0	)					bottom of treatment zone
	28.0	28.2	clay				
	30.0	30.5	sandy clay				
	30.5	5 30.9	clay				
	30.9	9 31.9	silty sand				
	31.9		sandy silt				
	35.0		clay				
	35.9		silty sand				
	40.0		silty sand				
	45.0		silty sand				
	50.0		silty sand				
	55.0		clay				
	60.0	63.4	clay				
			subtotal		9.9	5.740	1
			average fraction fines			0.580	
			number of layers			18.000	
			layers per 10 feet			18.182	
			treatment zone thickness			12.500	

Well Log	Тор	Bottom	Description	Fraction Fines	Measured Thickness	Weighted Fraction	Notes
Myrtle							
FT11-MW04	0.0	1.	0 clay				
	1.0		0 silty sand				
	2.0	4.	0 sandy silt				
	4.0	7.	5 clay				
	7.5		water table				top of shallow treatment zone
	7.5	8.	0 sand	0.025	0.5	0.013	
	14.0	15.	0 sand	0.025	1.0	0.025	
	15.0						bottom of shallow treatment zone
	15.0	16.	0 clay				
	16.0	18.	0 sandy clay				
	23.0	25.	0 clay				
	28.0	30.	0 clay				
	33.0	33.	5 clay				
	33.5	34.	6 silty sand				
	34.6	35.	0 sand with silt				
	38.0	40.	0 sand				
	43.0	44.	0 sand				
	44.0	45.	0 clay with sand				
			subtotal		1.5	0.038	
			average fraction fines			0.025	
			number of layers			2.000	
			layers per 10 feet			13.333	
			treatment zone thickness			7.500	

Well Log	Тор	Bottom	Description	Fraction Fines	Measured Thickness	Weighted Fraction	Notes
North Island General			silty sand	0.325			no log
Orlando							
OLD-17-04A/10C	0.0	0.5	5 gravel				
	0.5	1.	5 sand				
	1.5	6.0	) silty sand				
	6.0		water table				top of A/B treatment zone
	6.0		) silty sand	0.325	16.0	5.200	
	22.0						bottom of A/B treatment zone
	22.0		) silty sand	0.325			
	46.0		) sandy silt	0.600	9.0	5.400	
	55.0		) elevi	0.025	F 0		bottom of C treatment zone
	55.0	60.0	) clay	0.925	5.0		
	A/B treatm	nent zone	subtotal		16.0	5.200	
			average fraction fines			0.325	
			number of layers			1.000	
			layers per 10 feet			0.625	
			treatment zone thickness			16.000	
	C treatment zone		subtotal		33.0	13.200	
			average fraction fines			0.400	
			number of layers			2.000	
			layers per 10 feet			0.606	
			treatment zone thickness			33.000	

Well Log	Тор	Bottom	Description	Fraction Fines	Measured Thickness	Weighted Fraction	Notes
Point Mugu							
23W26B							
	0.0	)	asphalt				
	0.2	5.	1 sand				
	5.1		water table				top of Zone A
	5.1	. 9.	5 sand	0.025	5 4.4	4 0.109	)
	9.5	13.	0 silt	0.925	5 3.5	5 3.238	3
	13.0	) 15.	0 sand	0.025	5 2.0	0.050	)
	15.0	) 18.	0 silt	0.925	5 3.0	2.775	5
	18.0	23.	0 sand	0.025	5 5.0	0.125	5
	23.0	26.	0 sand	0.025	5 3.0	0.075	5
	26.0	28.	0 clay	0.925	5 2.0	) 1.850	)
	28.0	34.	5 sand	0.025	6.5	5 0.163	3
	34.5	35.	3 silt	0.925	5 0.8	3 0.740	)
	35.3	35.4	4 sand	0.025	5 0.1	1 0.003	}
	35.4	Ļ					bottom of Zone B
	A/B treatr	nent zone	subtotal		30.3	3 9.127	,
			average fraction fines			0.302	2
			number of layers			10.000	)
			layers per 10 feet			3.304	ļ.
			treatment zone thickness			30.270	)

Well Log	Тор Во	ttom Description	Fraction Fines	Measured Thickness	Weighted Fraction	Notes
Raritan						
MW304D	0.0	5.0 sand with silt				
	5.0	10.0 sand with gravel				
	10.0	water table				top of Upper Sand
	10.0	11.0 sand with gravel	0.025	1.0	0.025	
	11.0					bottom of Upper Sand
	11.0	21.5 clay	0.925	10.5	9.713	
	21.5					top of Lower Sand
	21.5	23.0 silt with sand	0.800	1.5	1.200	
	23.0	25.0 sand	0.025	2.0	0.050	
	25.0	30.0 sand with silt	0.100	5.0	0.500	
	30.0					bottom of Lower Sand
	30.0	clay	0.925			
	Upper Sand	subtotal		1.0	0.025	
		average fraction fine	25		0.025	
		number of layers			1.000	
		layers per 10 feet			10.000	
		treatment zone thick	kness		1.000	
	Lower Sand	subtotal		8.5	1.750	
		average fraction fine	es		0.206	
		number of layers			3.000	
		layers per 10 feet			3.529	
		treatment zone thick	kness		8.500	

Well Log	Тор	Bottom	Description	Fraction Fines	Measured Thickness	Weighted Fraction	Notes
Seal Beach							
AMW1		0.0 0.	2 aphalt				
		0.2 5.	0 silty sand				
		5.0 15.	0 clay				
		15.0	water table				top of treatment zone
		15.0 23.	0 clayey sand	0.325	8.0	2.600	
		23.0 28.	0 silty sand	0.325	5.0	1.625	
		28.0 28.	4 sandy clay	0.600	0.4	0.240	
		28.4 29.	5 silty sand	0.325	1.1	0.358	
		29.5 31.	0 sandy clay	0.600	1.5	0.900	
		31.0 32.	0 silty sand	0.325	1.0	0.325	
		32.0 33.	0 sandy clay	0.600	1.0	0.600	
		33.0 34.	0 silty sand	0.325	1.0	0.325	
		34.0 35.	1 sandy silt	0.600	) 1.1	0.660	
		35.1 37.	0 silty sand	0.325	1.9	0.618	
		37.0					bottom of treatment zone
			subtotal		22.0	8.250	
			average fraction fines			0.375	
			number of layers			10.000	
			layers per 10 feet			4.545	
			treatment zone thickness			22.000	

Well Log	Тор	Bottom	Description	Fraction Fines	Measured Thickness	Weighted Fraction	Notes
St Juliens							
IW1	0.0	1.0	) concrete				
	1.0	2.0	) gravel				
	2.0	5.0	) sandy silt				
	5.0	6.0	) silt				
	6.0		water table				top of treatment zone
	6.0	10.0	) silt with sand	0.800	) 4.0	3.200	
	10.0	17.0	) sand with silt	0.100	) 7.0	0.700	
	17.0	18.0	) sand	0.025	5 1.0	0.025	
	18.0						bottom of treatment zone
	18.0	20.0	) clay				
			subtotal		12.0	3.925	
			average fraction fines			0.327	
			number of layers			3.000	
			layers per 10 feet			2.500	
			treatment zone thickness			12.000	

Well Log	Тор	Bottom	Description	Fraction Fines	Measured Thickness	Weighted Fraction	Notes
Treasure Island							
EW-29	0.0	0.	0 asphalt				
	0.0	2.	0 sandy clay				
	2.0	5.	0 clay with sand				
	5.0	6.	0 clay				
	6.0	9.	0 sand with silt				
	9.0	10.	0 clay with sand				
	10.0	18.	0 sand with silt				
	18.0		water table				top of treatment zone
	18.0	30.	0 sand with silt	0.10	0 12.	0 1.200	
	30.0						bottom of treatment zone
			subtotal		12.	0 1.200	•
			average fraction fines			0.100	)
			number of layers			1.000	1
			layers per 10 feet			0.833	
			treatment zone thickness	5		12.000	1

Well Log	Тор	Bott	om	Description	Fraction Fines	Measured Thickness	Weighted Fraction	Notes
VAFB Site 15A								
15MW10		0.0	10.9	sand				
		10.9		water table				top of treatment zone
		10.9	26.5	sand	0.02	5 15.6	0.390	
		26.5	28.0	clay	0.92	5 1.5	1.388	
		28.0	39.0	sand	0.02	5 11.0	0.275	
		39.0						bottom of treatment zone
		39.0		bedrock				
				subtotal		28.1	2.053	
				average fraction fines			0.073	
				number of layers			3.000	
				layers per 10 feet			1.068	
				treatment zone thickness			28.100	

tment zone	
treatment zone	ne

Well Log	Тор	Bottom	Description	Fraction Fines	Measured Thickness	Weighted Fraction	Notes
VAFB Site 19							
19-MW-17A/B	0.0	0.5	5 asphalt				
	0.5	2.5	5 sand				
	2.5	5.0	) sand with silt				
	5.0	) 8.0	) clay				
	8.0	)	water table				top of Clay Zone
	8.0	10.0	) silt	0.925	2.0	1.850	
	10.0	) 13.5	5 clay	0.925	3.5	3.238	
	13.5	14.5	5 silt	0.925	1.0	0.925	
	14.5	5 15.0	) clay	0.925	0.5	0.463	
	15.0	) 17.0	) silty sand	0.325	2.0	0.650	
	17.0	) 19.5	5 clay	0.925	2.5	2.313	
	19.5	i					top of Sand Zone
	19.5	5 21.0	) silty sand	0.325	1.5	0.488	
	21.0	30.0	) sand	0.025	9.0	0.225	
	30.0	)					bottom of Sand Zone
	30.0	)	gravel				
	Clay Zone		subtotal		11.5	9.438	
			average fraction fines			0.821	
			number of layers			6.000	
			layers per 10 feet			5.217	
			treatment zone thickness			11.500	
	Sand Zone	2	subtotal		10.5	0.713	
			average fraction fines			0.068	
			number of layers			2.000	
			layers per 10 feet			1.905	
			treatment zone thickness			10.500	

Well Log	Тор	Bottom	Description	Fraction Fines	Measured Thickness	Weighted Fraction	Notes
VAFB Site 35							
35-MW-30		0.0 13.	0 sand				
		13.0 20.	0 sand with silt				
		20.0 26.	5 clayey sand				
		26.5	water table				top of treatment zone
		26.5 29.	0 clayey sand	0.325	5 2.5	5 0.813	
		29.0 29.	5 clayey gravel	0.325	5 0.5	5 0.163	
		29.5 30.	5 clayey sand	0.325	5 1.0	0.325	i i i i i i i i i i i i i i i i i i i
		30.5 31.	5 clayey gravel	0.325	5 1.0	0.325	i i i i i i i i i i i i i i i i i i i
		31.5 33.	0 clayey sand	0.325	5 1.5	5 0.488	1
		33.0					bottom of treatment zone
		33.0 36.	5 bedrock				
			subtotal		6.5	5 2.113	i
			average fraction fines			0.325	
			number of layers			5.000	)
			layers per 10 feet			7.692	
			treatment zone thickness			6.500	)

Appendix F: Representative Soil Boring Logs



# **Drilling Log**

## Monitoring Well **PP41IW01**

Project <u>Alameda CTO 10</u> Location <u>Alameda CTO 10</u> Location <u>Alameda CA</u> Surface Elev. <u>NG=11.33</u> Top of Casing <u>11.14 ft.</u> Screen: Dia <u>2 in.</u> Casing: Dia <u>2 in.</u> Fill Material <u>See Comme</u> Drill Co. <u>BC 2</u> Driller <u>Diego Rivera</u> Checked By <u>Gene Mulle</u>	<u>3 ft</u> Total Ho Water Le Length Length nts Log By nmeister	evel Initial <u>N4</u> <u>10 ft.</u> <u>24.5 ft.</u> Method <u>Hc</u> <u>Jim Teo</u> License	.4 ft. Rig Mow St	Wner         US Navy           Proj. No.         844918           North         470798.61 ft.         East         1481761.1 ft.           Static         4.4 ft.         Diameter         8 in.           Type/Size         Fiberglass/slotted 0.020 in.           Type         Fiberglass           g/Core         8" Hollow Stem           em Auger         Date         11/5/03           Driller License C57# 686255         Description	Page: 1 of 2 COMMENTS Water ~10 gallons; #2/12 Sand ~5 bags (100 lb bags); Bentonite ~5 bags (60 lb bags) NG = Natural Ground Static Water Level: 1/26/04
Completion	(ppm) Sample ID % Recovery	Blow Count Recovery Graphic Log	USCS Class.	(Color, Texture, Structure Soil Descriptions are Based on the I	
- 0 2			GW	<ul> <li>@ ~2-5 ft, Well graded gravel, dry, pebbles, rock very dense (95% gravel; 5% sand) (next to railro</li> <li>@ ~8-10 ft, Lean clay, yellowish brown (10YR 5/ pebbles and rocks, firm, coarse grained fraction i</li> </ul>	ad tie) 
- 10 - - 12 - - 12 - - 14 - - 14 - - 16 - - 18 - - 18 - - 20 - - 20 -			ML	<ul> <li><u>sand</u>; 10% silt; 40% clay)</li> <li><u>and</u>; 10% silt; 40% clay)</li> <li><u>and</u>; 10% silt; 40% clay)</li> <li><u>and</u>; 10% silt; 30% clay</li> <li><u>and</u>; 10% silt; 40% clay</li> <li><u>and</u>; 10% silt; 30% clay</li> </ul>	3/2), wet/saturated, with and gravel, liquid but l; 50% silt; 20% clay) et/saturated, cohesive,
Rev: 2/23/00 ALAMEDA CTO 107.GP1 IT_CORP.GD1 - 18				@ ~20-25 ft, Sandy silt, as above, color change 5/3)	to light olive brown (2.5Y
			CL	@ ~30 ft, Sandy lean clay, olive (2.5Y 5/4), wet/s <i>Continued Next Page</i>	



## **Drilling Log**

Monitoring Well

**PP41IW01** Page: 2 of 2

							Page: 2 of 2
	Alameda CTO Alameda, CA					_ 0	wher <u>US Navy</u> Proj. No. 844918
Location							Proj. No. <u>844918</u>
Depth (ft.)	Well Completion	DIA (mdd)	<u>Sample ID</u> % Recovery	Blow Count Recovery	Graphic Log	USCS Class.	Description
<u>s</u> e	Com	d d	<u>Sam</u> % Re	Blow Rec	Gra	JSCS	(Color, Texture, Structure) Soil Descriptions are Based on the USCS.
						_	Continued
— 30 — -       -							frequent medium grained black sand, soft to firm, slightly plastic, slightly sticky (50% sand; 25% silt; 25% clay)
- 32 -						CL	
- 34 -							@ ~35 ft, Sandy lean clay, as above
- 36 -							TD = 35.4 ft
 							Survey Datum: Horizontal: NAD 27, Vertical: NAVD 88 Note: Soil descriptions were generated from soil grab sample specimens collected from the veins of hollow stem auger flytes. As a result, the depths
							of all descriptions and soil horizon boundaries are approximate.
- 40 - 							
— 42 — -       -							
- 44 -							
- 46 -							
- 48 -							
 - 50 -							
 - 52 -							
 - 54							
— 56 — -       -							
— 58 — -       -							
- 60 -							
- 62 -							
 - 64 -							
- 56 - - 58 - - 58 - - 60 - - 62 - - 64 -  - 64 -  - 68 -  - 70 -							
- 68 - 							
- 70 -							

AIS-TN&A Joint	AIS-TN&A JV 317 E Main St Ventura, CA 93001 Telephone: (805) 585-2110 Fax: (805) 585-2111	WELL NUM	IBER 23W26B PAGE 1 OF 1
CLIENT US Dept of the Nav	y	PROJECT NAME IRP Site 24	
PROJECT NUMBER 200904		PROJECT LOCATION _NBVC Point Mugu, CA	
DATE STARTED 12/17/12	COMPLETED12/17/12	GROUND ELEVATION _10.13 ft HOLE SIZ	ZE _12" OD
		GROUND WATER LEVELS:	
	Stem Auger	- AT TIME OF DRILLING <u>4.60 ft</u> / Elev 5.53 ft	
LOGGED BY M Wanek	CHECKED BY J Harting, PE		
	mpler; AU-Hand Auger	AFTER DRILLING 5.13 ft / Elev 5.00 ft	
C DEPTH (ft) SAMPLE TYPE NUMBER RECOVERY %	LAW CGRAPHIC LOG LOG	FERIAL DESCRIPTION	WELL DIAGRAM Casing Top Elev: 9.35 (ft) Casing Type: 4" Sch 40 PVC
0	.0 XXXX 2" ASPHALT		
	Brown (7.5YR, 5/4) fine SAN	D with trace silt. Moist. Medium dense.	Bentonite-Ceme
	~0.2' each. Grades to fine-m each sequence.	edium sand with trace shell hash at bottom of	Seal 4"-Diameter
10 MC 93 0	ML inclusions approx. 1.5" diama ML - Becomes very dark gray (5 - Becomes dark yellowish-bri secondary mineralization	Y, 3/1) with no sand and no organic inclusions. own (10YR, 4/4) with strong brown (7.5YR, 4/6)	Sch 40 PVC Blank Casing
15 MC 87 0	.0 SW hash. Wet. Medium dense.	raded (fine to coarse) SAND with trace shell	
	ML Becomes sandy - Becomes dark yellowish bro	SILT. Wet. Medium dense. Finely laminated own (10YR, 3/4)	
	.2 .2 .2		✓ Bentonite Seal, Hydrated Bentonite
25 MC 73 2	Dark gray (10YR, 4/1) fine S		Chips +#3 Sand Filter Pack
	CL Gray (10YR, 4/3) CLAY. Mo Gray (10YR, 5/1) fine SAND.		
	.2 SP - Becomes dark gray (10YR,		0.020-inch Slotted 4"-Diameter Sch 40 PVC Screen
35 MC 100 0	SP Very dark gray (5Y, 3/1) SIL		#3 Sand Filter Pack
GENERAL BH /		/2) fine SAND. Wet. Medium dense m of borehole at 35.5 feet.	



## Drilling Log

## Monitoring Well **PP41IW01**

				Page: 1 of 2
Project Alameda CTO 107		_ 0	wner US Navy	COMMENTS
Location <u>Alameda, CA</u>			Proj. No. <u>844918</u>	Water ~10 gallons; #2/12 Sand ~5 bags (100 lb bags); Bentonite ~5 bags (60 lb
Surface Elev. NG=11.33 ft	Total Hole Depth35	.4 ft.	North <u>470798.61 ft.</u> East <u>1481761.1 ft.</u>	bags)
Top of Casing11.14 ft	Water Level Initial	1	Static <u>4.4 ft.</u> Diameter <u>8 in.</u>	NG = Natural Ground
Screen: Dia <u>2 in.</u>	0		Type/Size Fiberglass/slotted 0.020 in.	Static Water Level: 1/26/04
Casing: Dia <u>2 in.</u>	Length24.5 ft.		Type	
Fill Material See Comments		Ri	g/Core _ 8" Hollow Stem	
Drill Co. <u>BC 2</u>	Method	xllow St	em Auger	
Driller <u>Diego Rivera</u>	Log By <u>Jim Teo</u>		Date <u>11/5/03</u> Permit # <u>W03-0978</u>	
Checked By Gene Mullenmei	ster License	e No.	Driller License C57# 686255	
Depth (ft.) (ft.) Completion PID (ppm)	Sample ID % Recovery Blow Count Recovery Graphic Log	USCS Class.	Description (Color, Texture, Structure Soil Descriptions are Based on the	,
		GW	@ ~2-5 ft, Well graded gravel, dry, pebbles, rock very dense (95% gravel; 5% sand) (next to railro	s and cobbles, dense to ad tie)
 - 8  - 10		CL	@ ~8-10 ft, Lean clay, yellowish brown (10YR 5/ pebbles and rocks, firm, coarse grained fraction i sand; 10% silt; 40% clay)	s dense (40% gravel; 10%
12			<ul> <li></li></ul>	and gravel, liquid but ; 50% silt; 20% clay) t/saturated, cohesive,
Metro 18		ML	massive with few silty clay aggregates interspers sticky (20% sand; 50% silt; 30% clay) @ ~20-25 ft, Sandy silt, as above, color change	
LI COMMERCIAL Rev. 2/23/00 ALAMEDA CTO 107.662 IT 26		CL	@ ~30 ft, Sandy lean clay, olive (2.5Y 5/4), wet/s	
			Continued Next Page	



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COMMERCIAL

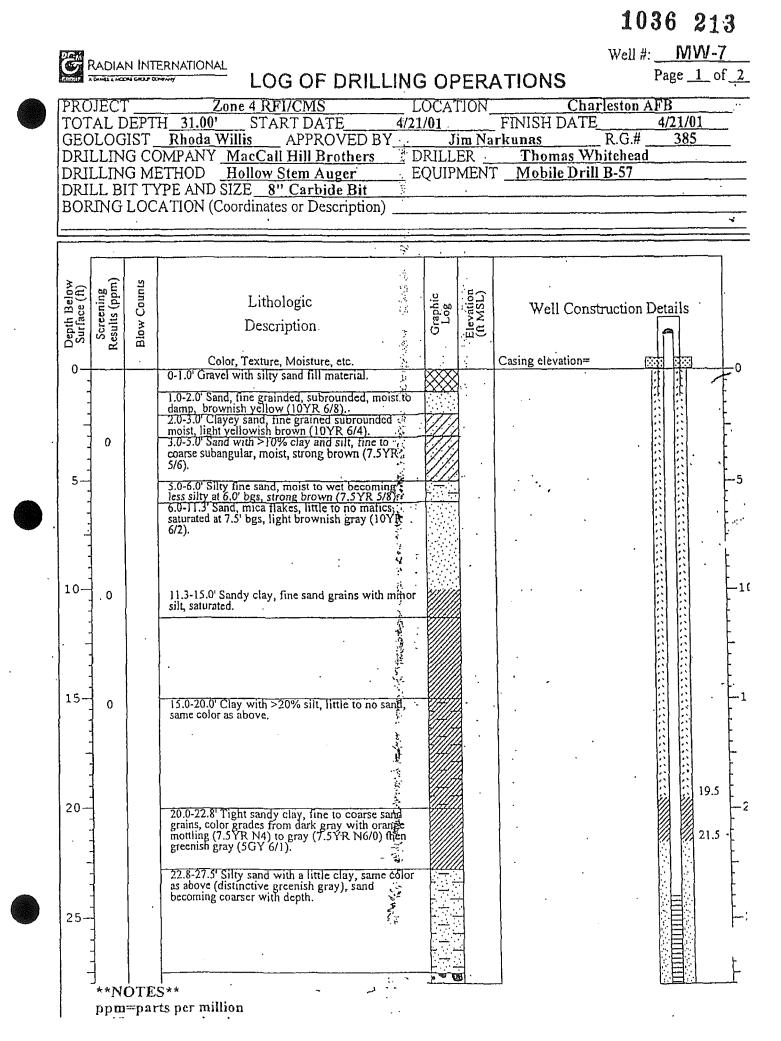
F

## **Drilling Log**

## Monitoring Well

### **PP41IW01** Page: 2 of 2

Project Alameda CTO 107 US Navy Owner Location Alameda, CA 844918 Proj. No. Sample ID % Recovery Blow Count Recovery Well Completion USCS Class. Description Graphic Log (ft.) (mqq) (Color, Texture, Structure) Soil Descriptions are Based on the USCS. Continued 30 frequent medium grained black sand, soft to firm, slightly plastic, slightly sticky (50% sand; 25% silt; 25% clay) 32 CL 34 @~35 ft, Sandy lean clay, as above TD = 35.4 ft 36 Survey Datum: Horizontal: NAD 27, Vertical: NAVD 88 Note: Soil descriptions were generated from soil grab sample specimens 38 collected from the veins of hollow stem auger flytes. As a result, the depths of all descriptions and soil horizon boundaries are approximate. 40 42 44 46 48 50 52 54 10/5/04 56 GDT CORP. 58 107.GPJ 60 CTO 62 ALAMEDA 64 2/23/00 66 Rev: 68 70





RADIAN INTERNATIONAL

# I OG OF DRILLING OPERATIONS

**1036 214** Well #: <u>MW-7</u> Page <u>2 of 2</u>

	JECI	·	Zone 4 RFI/CMS	_ L(	Charleston AFB	<u>Charleston AFB</u>			
ر ۱						·····	۵۰۰ می این م ۱۹۹۰ می این می	`	
Depth Below Surface (ft)	Screening Results (ppm)	Blow Counts	Lithologic Description		Graphic Log	Elevation (A MSL)	Well Construction Details		
4			Color, Texture, Moisture, etc.		more				
- - 30			Color, Texture, Moisture, etc. 27.5-28.0' Gravel layer. 28.0-29.0' Cooper Marl. Bottom of the hole at 29.0' bgs.	/			29.0 29.3 31.0	-	
		•						<u></u>	
- - 35 -				• .			:	ممراسع	
		-						عملمه	
40-								ببيليي	
- 								عرم منالحي	
45-								عبياب	
								ـ ـ ـ ـ ـ ـ ـ	
50-								لمريح والمراجع	
55-								للمراجع المراجع	
   <sup>60_</sup>									

	QC.	r. 9.1	998 10:59	iam I	DAMES8	MOORE	Ν	10.215	P.4
	<b>-</b>			• ••	••••				
						R			471
						C C JOH	NSON & MALHOTRA, P.C.	÷	
S	)IL DOR	ING LOG					WELL INSTALLATION A	n) comp	LETION DATA
_		DOVER 6			: T-1		Job No. : 5152-00 Date Brilled		9 <b>1</b> 20/88
70	stal De	: #1 21 pth : 48 ize & Ty		Scree	n Lengt	00' SH OF : h : 10' ng, : KL	SIT Driller : MATHES Method : HS o Ground Elev. : 21.92' Top of Casin Org. Vap. Instr. : HMU		: 25.00' 03/21/ <del>9</del> 0
			;Blows per é" i 140 Ibs.			i PPH	l Description	l Soi 1 ICI ess	
-     	0		; ; ; ; 4,18,16,12	}	       75 %	, 	Br silt, sand & pebbles; org mat & toal; dry	sv .	i Very dark I Greyish I Brown
1	2			1			Srey F to and sand with gradationally more		10YR 3/2 
;		   	: 4,8,12,14 }	t 1	  100 Z 	;	clay at depth (trace), trace pebbles; dry	sp I	IDIive brown   2.5Y 4/4 
	4	     	   	   	     83 I	i bkg	F Orange and grey mottled med to crs mand with I trace clay and vc mand grading to crs mand		   Light gray   2.5Y 7/2
í 1 1	6	   <del>* - + *</del> 1	 		} ]	) 	l with some vc sand; dry	 	i    Dark yellow
j	A	i     !	10,7,9,10	ř 1 1	; ; 83 7 ;	.2 	; ;   Greyisk-brange and to crs round to subang   uniform sand with trace clay; moist	i       5P	I Brown I Brown I 10YR 4/6
	v	; ; ; ;	1 6,7,8,7	- - - - - -	03 X	: : : .6 :	I MIGINA WIRT FLUTE FYRY! RAIDF 		l Light grey 1 2.5Y 7/2 1
	10	; { !	; ; ; ; ; ; ; ; ; ; ; ; ; ; ; ; ; ; ;	1 1 1	;	1.6	Tan f to med sand coarsening with depth to med to crs sand with f sand lenses, trace	     SN	: ; ;
í   	12	i   ======= 	 	i   	i  ,	i }	to little clay; wet	5 {	 
	14	 	i i 1,3,3,7 i	2 [ [ 	   878 X 	   _2 	l     Eran and be ever round be coherend coord	;     	l Light   Brownish   Grey   2.5Y 6/2
		; ; ;	1 4,5,6,12	• • • •	   na 	   4 	l Grey and to crs round to subround sand coarsening to crs to vc sand with occ . I med sand and trace clay lenses; wet	9P   	·
1111	16	] <del>************************************</del>	;     2,2,12,12 	i     	     58 7 	623 8 26   .   . 		;   	I Light grey I loyn 7/2
	18	} <del></del>	4,2,4,6		 59 %	     12	Very poorly sorted clay, sand, and pubbles. Clay is in green chunks, sand and pubbles are saroon; wet	(   SN 	Brown   7.5YR 7/2
* 1 1	20	   69-000 mm m.g	- ]			CONTINUED	ON NEXT PAGE	1	1
i 									

NO.219 P.4 Contraction of the second

OCT. 9.1998 10:59AM DAMES&MOORE

474 5

## C C JOHNSON & MALHOTRA, P.C.

Hell No. : M 2110 100

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(CP)

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Site ; 7-1

Job No. : 5152-00

Date Drilled : 04/19620/88

1-1	140 lbs	¦Interval	IRecov.	I PPN	1 Description	(Soil (Class	
4 3 1	;     2,6,5,9 	1	1	1 30	: : : Brown uniform round to subround crs sand		l Yellow l IQYR 7/B
; ; ; ;	1 3,6,7,7		i 50 z	30	; cparsening to crs to vc sand with trace ; pebbles; wat ; ;		I Yellowish I Drown I 10YR 5/9
1	1 5,8,6,5		; ; 58 %	1 1 20 1	/ 	• [ ] ?	istrong bro 1 7.5YR 5/9
9 1 2 3 9	2,5,7,11	;     	63 X			F 7 7	
	4,4,5,7		63 X	10		SP	1 1 1 1
	3,4,6,9		63 I	15	l I I Grange fine to wed sand with trace clay; wet		   Yellow   10YR 7/6 
	3, 3, 4, 8		71 7	15			Istrong broi 7.5YR 5/8
	1,4,7,9	• •	•	60			Brownish   Vellow   10YR 7/6
	2,0,11,14		; 58 Z :	30			   Yellow   £0YR 7/6
I SB	5,8,14,12			1	Drange mud to crs sand with trace clay and publics becoming crs sand with depth; wet i	•	
1	1,7,15,12	*   # &	50 2 1				Strong brav 7.5YR 5/8
	2,4,5,6	1	I	1	Grange silty clay with grey-brn lasy spist	CL (	
	L) : Sample	L) Sample       140       1bs.         0	b) Sample 140 lbs. Interval 2,6,5,9 2,6,5,9 3,6,7,7 3,6,7,7 5,8,6,5 2,5,7,11 4,4,5,7 4,4,5,7 3,3,4,6,9 3,3,4,6,9  3,3,3,4,8  58 50 5,9,11,14  58 1,7,15,12  1,7,15,12	b) Sample 140 1b5. Interval Recov. 2,6,5,9 66 7 2,6,5,9 66 7 3,6,7,7 50 7 5,8,6,5 58 7 2,5,7,11 63 7 4,4,5,7 63 7 4,4,5,7 63 7 3,4,6,9 63 7 3,3,4,8 71 71 7 3,3,4,8 71 65 7 1,4,7,9 66 7 2,8,11,14 58 7 58 7 1,7,15,12 58 7	$ \begin{array}{c ccccccccccccccccccccccccccccccccccc$	D: Sample 140 lbs.       Interval Recov.: PPN i       Description         D:	D:Banue       140       105.       Inferval Recov.       PPH       Description       IClass         D       2,6,5,9       65 I       30       Brown unifors round to subround crs sand       Coarsening to crs to vc sand with trace         1,6,7,7       50 I       30       Brown and orange and to crs sand with trace       pabbles; wat         1,6,7,7       50 I       58 I       20       Incom and orange and to crs sand with trace         2,5,7,11       63 I       10       Incom and orange and to crs sand with trace       Incom and orange and to crs sand with trace         1,4,4,5,7       63 I       10       Incom and orange and to crs sand with trace clay; wet       Incom and orange and to crs sand with trace clay; wet         3,4,6,9       63 I       10       Incom and orange and to crs sand with trace clay; wet       Incom and orange and to crs sand with trace clay; wet         3,3,4,8       71 I       15       Incom and to crs sand with trace clay; wet         1,4,7,9       06 I       60       Incom and to crs sand with trace clay and pebbles; becoming crs sand with trace clay and pebbles; becoming crs sand with depth; wet         1,7,15,12       58 I       20       Incom and to crs sand with depth; wet         1,7,15,12       50 I       15.4       Incom and to crs sand with depth; wet         1,7,15,16       71

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### C C JOHNSON & MALHOTRA, P.C.

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Hell No. : HV 2110

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Site : T-1 Job No. : 5152-00 Date Drilled : 04/19620/08

(fee	Ð	Sample	140	lbs.	Interval	Reco	v, i		•	Spil  Class	l Color I
	₽~~ [		0%hu\$99		h	111'10' 10' 10'	!		프로프트 ···································	i	•
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	1		1		I	1	1			ł	1
	1		Shelby	tube	1	ł	I		Not recorded	1	1
	1		1		I I	I	1			:	ł
44	6 1		ł		1 1	1	1			: CL	:
	1	SB		8 # <del></del>	[ <del>8344444</del> 7	4000	***	**		ł	ſ
	- 1	211D 46			: 1	ł	;		Charcoal grey silty clay with light grey	ł	l Very dark
	ł		1 3,3,	5,7	:	92	7 ;		v fine sand lenses; moist	1	l Gray
4	8 1		1	•	1	1	- 1	-	-	1	10YR 3/1

## BOTTOM OF BORE HOLE



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E		-		440000.000 <del>05</del> 10										AD ATTACANA														
HTF	RM	D	RII	LI	NG	) L	00	3				RICI													E NUM		11-4	. <del>.</del> .
1. COMPA													IG SU	BCON		CTOF	۹								07-1 EET	0951	SHEE	
URS													umbi													OF 4	8	
3. PROJEC	CT															DCAT	×							.L	<u>.</u>			·····
Moody	AFB	Data	Gar	os Inv	estig	ation												B. V	ald	osta.	Geo	rgia						
5. NAME C	F DRIL	LER	******												6. M	ANUF	ACTU	RER'S	DESI	GNATI	Geo DN OF	ORILL					1979)	
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7. SIZES A AND SAME	ND TY	PES O	É DRIL MENT	LING											8, H		OCA1	TION		1								
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12. OVERE		N THIC	KNES	S											15. I	JEPTI I	H GRO	DUNDV	VATE	RENCO	DUNTE	RED						
13. DEPTH				ск										-	16 (	)FPTI	JA	> WATER		FLAPS	ED TIN				10		-	
Ν	IA .															N	AU								10			
14. TOTAL	DEPTH		IOLE												17, (		rwa JA	TERLE	EVEL N	IEASU	REMEN	NTS (S	PECIF	Y)				
18. GEOTE			MPLE	s	1	DIS	TURB			Τ	UN	IDIST	URBE		Т			LNUM	BER O	F COR	E BOX	ES						
	)						NA	-			C			•				NA	-									
20, SAMPL				-	<u> </u>	VO				VETA				THER				0	THER	(SPEC		0	THER	ISPEC	(IFY)	21, TO RECO	OTAL (	ORE
22. DISPOS	6				10		<u>NC</u>			N/	G WE		(2	)	Pł	1		(2)	<u> 11</u>	<u>C</u>		6	B	<u> (</u> , (,	P		Ŵ	<b>₩</b>
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PROJECT		doras doras da										wydaleszaid									Н	OLE	l WWWWWWWWW					
Moody A	FB D	)ata (	Gaps	Inve	stiga	tion															F	רס-	200	111 -	.07			
			Contraction of the last				(internet and in the second second second second second second second second second second second second second					-	anno conserva		-					Citerature	1.	<u>10 /-</u>	<u>C40</u>	111	<u>(v)</u>		-	

ENG FORM 5056-R, AUG 94

(Proponent: CECW-EG)

PROJEC	TRW DRILLING LOG (CONTINUAT					FT07-DPS111-65
Mood	dy AFB Data Gaps Investigation	Inspector	n Maler			SHEET SHEE
ELEV. a.	b. C.	FIELD SCREENING RESULTS d.	GEOTECH SAMPLE OR CORE BOX NO. e.	ANALYTICAL SAMPLE NO. I.	BLOW COUNTS g.	REMARKS h.
	S. Ity SAND (M) - Louse, must, branising my	Field Sciency			<u>u</u> .	Time: 0823
	fine grined send, true organies (rosts) (top)	D' Breating ZAM				Recoury (R): 38"
		(B7) = ND				Pocket Pentronder
						(N)=
		ES= 0.0 pp				
ľ		14.7				Hydroeuban odar (Stinig)
	Becomes with clay & 10450	168				
	3 - Scoly SILT (Mc) Key all and and D	373				
	3 - Serby SILT (Mc) Vey silt just, grey, frequent Serb, true clay \$15%, lonply tic	545 285				Ne 6500 Perchaductor and
		502				2.9' to 3.71
ı	- - 4 -				1	10.45
	5 - Becames light brank					
	- Becans black and gray	FST 50 804	_			Time = 0031
	- Becanes black and gray			111-6+(ETD	2-025611-0	R= 54"/60"
0	G	821 105		2) P/0 M TOL		Hyorocarbon octor
	-Becares with more cky (225%)	513				(strong)
	- Sordy and Clayey SILT (me) Vary soft, net,	124				N= <500
7	7 - greyand tan, bu photic, Ane growed sends	୫ଢ୩				N= (3 013
	-	92				,
	*	1 77				
	8 - Becknes Saft Becares marst threat	192				N= 600
	Bocares marst the wet	0.0				
9	7 - Becares morst, medium shift	0,0				N= 1200
	Becares morst, medium shift Becares shill, with dark reddish and gruy color	7.2				1= 2 500
		0.0		-		
10		BZIND				
oject ody A	AFB Data Gaps Investigation ORM 5056-R, AUG 94				DLE NO.	111.65

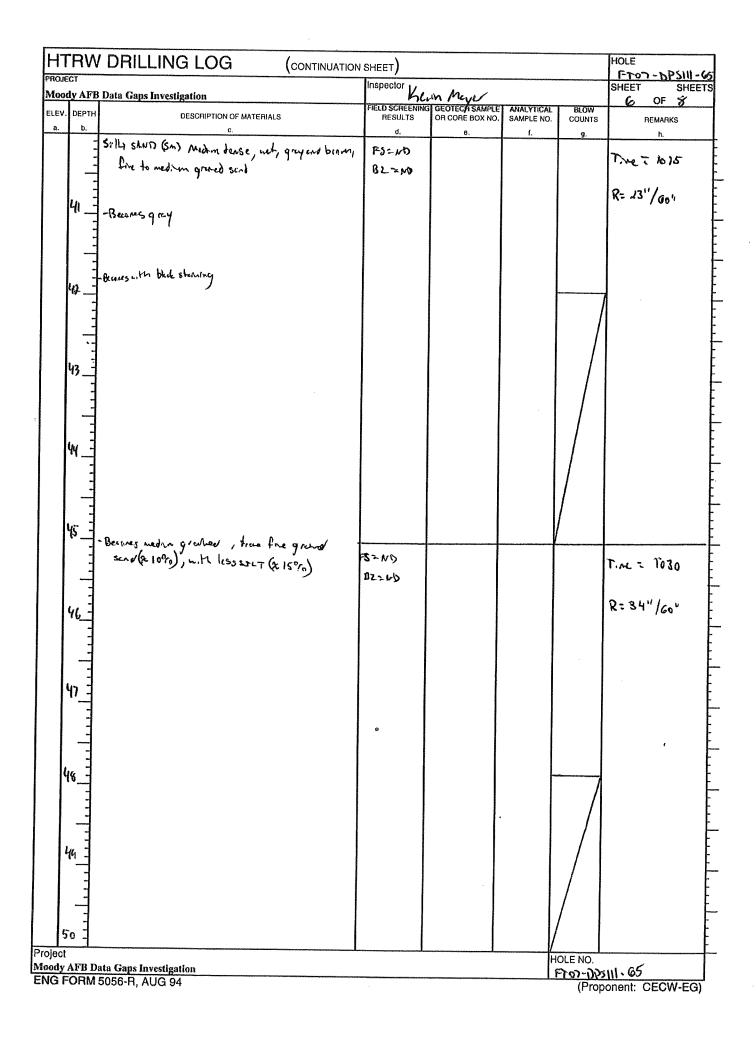
ROJE		DRILLING LOG (CONTINUATION	Inspector				1707-0PS111-65
100	dy AFI	3 Data Gaps Investigation	her	Mue			SHEET SHEETS
LEV. a.	DEPTH b.		FIELD SCREENING RESULTS	GEOTECH SAMPLE OR CORE BOX NO	SAMPLE NO.	COUNTS	REMARKS
	:	Sordy SPLT(MC) Saft, uch, gry, fre grand sid,	d.	е.	1.	g.	<u>h.</u>
	-	lonplished trace chy (2013 %)	- + 783+				Time: A840 N=700
		Scroy and clayery SILT (mc) Very stiff; must	45				1 ·
	11 -		82				R= 594/604
	-	grey, low pistre, frequed ands, concourse	49				
	-	readsh buring,	51				Strong hypernicerban
	-		ור				odar
	12 -		70				
			142				
	-	-Becanes grey and tan -Becanes grey	295				
	••		29				
	13	-Becanes gicy	22				
	· –		26				
	-		49				
	_]	i de suderte to mo	14,6				
		Beauses why SILT peckets relation ton, non pluste	3S.(,				
	μ_]	H prise	42				
	-	Hy Belcanes it in SILT pochets @ 14.45'	0.75				
	-1		BS=ND			·,	
	15-1	-Becomes must to wet					
	]		FS: 78 pp		Fton-Opsili		P antia
Z E			215 70		-17		Time = 2849
	E	Silly SAND (SM) Loose, marst heret, gray, fre			Ftor-OP3611 - 17		
	63	graved send, trace cluy (212%)	14.5		PH		R= 56"/60"
	-1		6.8		Tec		
	-	Becomes with a ley (2 15%)	324		Bifferry		
	1	CLATI(CH) Stiff, noist, grup and greenshyry, nighly plastic, true silt 210%, true fre gruned scio (210%)	109		Capacity		N: 3300
	17]	nighly plaster, true silt 210%, true fre	1.4		,		··· 3300
	4	grand schol (2 1096)	0.0	ľ			-
	1		BZ=ND				
	1						' F
h	8_T	Silly and Servy CLAY' (CL) SOPI, NOT +, g'cy,	5.8				
	-	Pile graded scrod, low plustic Balets with more scrod 240°re	413				N=600
I	+	Steers with more send 240°rs	27.3				Ē
	7	-Becares relien stalf	રુર.૧				F
	4 P	J	11.4			ľ	v= 1200
		Becanes gray ar reddien brown	79.2				
			37.4				ŀ
		Devenes with Iron stringing (early reddich)	0,0				F
1	20 ]		9				F
ect					k	HOLE NO.	
dy.	AFB D	ata Gaps Investigation 5056-R, AUG 94				FTO7-DP	JU. 65

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IO.IE		V DRILLING LOG (CONTINUATION	Inspector				HOLE FTOT-DASIII-63
000	dy AF	B Data Gaps Investigation	Kei	GEOTECH SAMPLE			SHEET SHE
.EV. a.	DEPTI b.	H DESCRIPTION OF MATERIALS	FIELD SCREENING RESULTS d.	OR CORE BOX NO.	ANALYTICAL SAMPLE NO.	BLOW COUNTS	REMARKS
		- Silly SAND (Sm) Dense, morst, gicy brown, fix to redem - gract; true clig(4 63m)	FS= 10 pp	e,	1	9.	T. we = 0855
	-	Sity CLAH (CL) - Still, unistroyour ad greening at, medun Phistic, free frequents coo (+ 52%) willing Because with phicker (CH)	69 ↓ .15				N: 3100
·	21 _	Some, SILT (mc) - Staff, most, date readism and brown, maplisher fine, SILT (mc) - Staff, most, date readism and brown, maplisher fire to readism gritmed sind	34 0,0				R= 34"/000"
		- and CENT (SC) - STILT, MOST, AIU, ELORAN, MIDDA GUL	Huraka, In- ple	LETO'T-OPSILI			0056=4
	ລາ	Surty Stat (me) - Some as 20.4 40 21.0 - with parts 20.4	BZZWD	-22 Gentech			Changes based on
	- -	Silly SAND (Sm) SERE 25 20.80' 13 20.35' Pryced brann	0				geotulusical
		- Sandy stor (m) - come is 201 to 21, 2's childred, rod bin	£.				results for PT07
	J3_						
		-					
ľ	<u>-</u> 4ג :						
	25_2					/	
			FS=ND		{		N= 3100 Tine t 09/1
		SAND(SP) Dense, net, fine to medium grand, Whitish gray	02=N0				R= 38"/6"
	26	3,					- /••
		Surgy CLAT'(Sc) - Stiff, most , gray, greenich yry, and tan, medium puetic					N=2800
	~ ' T	Sondy SFLT (Mu)-straf, musty reddilin brown,					
	-	Sitty SANDSard CLAY(CH) (SM-CH) Stifffly, notin of duse sind, most, gicenting iny clay, white and					
		in send, the to reduce gund send, highly platicely					N=9900 '
17	<u>-</u> %	f- (CH) layers					
					-		
:	21						
	Ţ					/	
ect						DLE NO.	
dy .	AFB	Data Gaps Investigation / 5056-R, AUG 94					5111.65

and the construction

11	ΗW	/ DRILLING LOG (CONTINUATION	SHEET)				HOLE
NEC			Inspector			·····	1=70-1-15111-6.5 SHEET SHEETS
		3 Data Gaps Investigation	FIELD SCREEKING	Len Myer SI GEOTECH EAMPLE	ANALYTICAL	BLOW	5 OF 8
V. C	DEPTH b.	DESCHI HON OF WATENALS	RESULTS	OR CORE BOX NO. e.	SAMPLE NO.	COUNTS	REMARKS h.
	-	Sendy CLAY(Sc) - Medium of Ff, mast, greensing by totan, disk red, for to making graned, muttled, banded, languise	FS=10				Time = 0938
	_	CLAY (CH) Stall, marsh, tanish brown, highly plastic, two	BZ=ND				N= 1400
		sift (22%) it race fire graned sind (R2%)					R=27"/60"
3	<u>-</u> א -	Silly Shrid (Sh) - median dense, moist, tennite gray, Are to median grand soid, true chy (a to 1)					N= 3000
		- Becores restrict brown, the more eley (21200), notfledard lowinches					
	9 <u>-</u>	Sordy SILT (ML) Medium Still, moist, redding brown,					
	-	fine grand sind, low placer, stratified lyers					ør: 1400
	•	,				1	-
3	3_						
	1					/	F
	1						-
	-						E
3	비그						-
	-						F
	-1				•		Ē
	, ]					/	ŀ
35		PLANK, N			/		
	1	Dieder	FS: MD				Time = 0957
	-	, 	BZ IND				N= 2800 R= 24"/60"
3	6 7						
	]	Silly SAND (SA) Dense, moret, tunneth gray, fire to medin grized, with clay (R 1500) Becomes gray, with no clay					
	1	grind, with cluy(2, 150%)					-
	1	sucres gray, with no clay					· · ·
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40	1				/		-
t	L	ata Gaps Investigation			<u> </u>	OLE NO.	



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	B Data Gaps Investigation	FIELD SCREENING	in Muyer GEOTECH SAMPLE	ANALYTICAL	BLOW	T OF 8
EV. DEPTH	6.	HEBULIS d	OR CORE BOX NO. 6.	SAMPLE NO.	COUNTS	REMARKS
	Silty SAND (Em) - Medan dense, net, byut reddroh brown, fine to redungrand	BZ2ND PS=ND	0.	<b>I</b> .	9.	Time = 1104
		10-100				
51_	-Becomes dirk reddish branen, with more sILT & 45%, true cluy (* 10%)					R= 181/6011
52				•		
53						
54_						
-						
75						
	they a strong stor fact - Saft, maist, greenist grey as read in Drank, low to medum plusher, mothed	FSIND				The = 1118
	pricercy mothly	BSIND				N= 24 / 60" "
56 -						- / 60"
						-
57_						· · · · · · · · · · · · · · · · · · ·
		🤊 :			/	, ,
58_						
59-						-  -  -  -  -  -
	1				/	
ect				————/н	OLE NO.	
dy AFB I	Data Gaps Investigation 4 5056-R, AUG 94			H	IOLE NO.	<u>DPSIII~65</u> onent: CECW

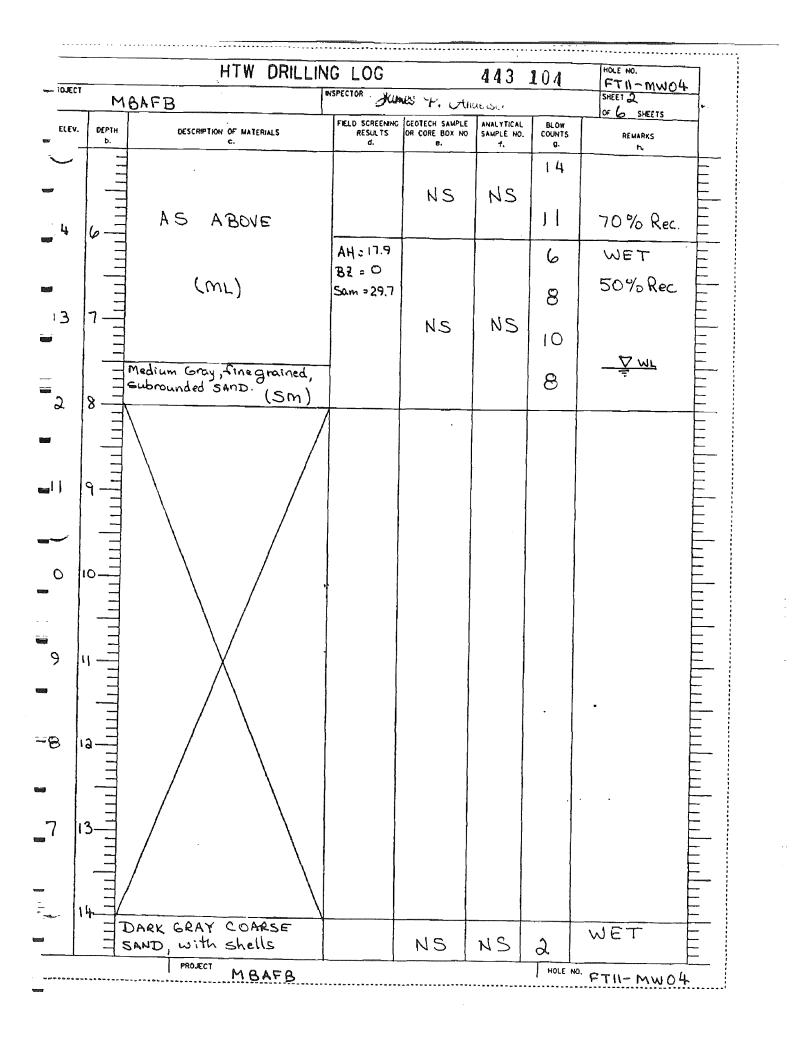
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oody	AFB	Data Gaps Investigation	Inspector Ke	un Mayer			SHEET SHEETS
EV. DE	EPTH b.	DESCRIPTION OF MATERIALS c,	FIELD SCREENING RESULTS d.	GEOTEØH SAMPLE OR CORE BOX NO. 0.	ANALYTICAL SAMPLE NO, I.	BLOW COUNTS	REMARKS
		Stilly CLAY (CL) suff, moust, reddom light taning reaction gruy	FS=ND BZ=ND			9.	Time = 114B
61	-	- Becomes medium station, mediumplashic					R= 39"/Go"
	-	-Becanes motiled , with sudir (CH) fitchey nodulos					
62	-						
G3	-	Bisci CLAY(CH) Stull, mr.St, gicensingary, highly plastic, alter reddith iron staking (mattled) (speakled)					M: 3300
						/	
64							
GS							
							Botton of brily @ GS, 0'
¢4.							
,							
47.							
- 68							ب الم الم الم
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49_							
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iy AF	B D	ata Gaps Investigation 5056-R, AUG 94			H	IOLE NO.	און - גל ponent: CECW-EG)

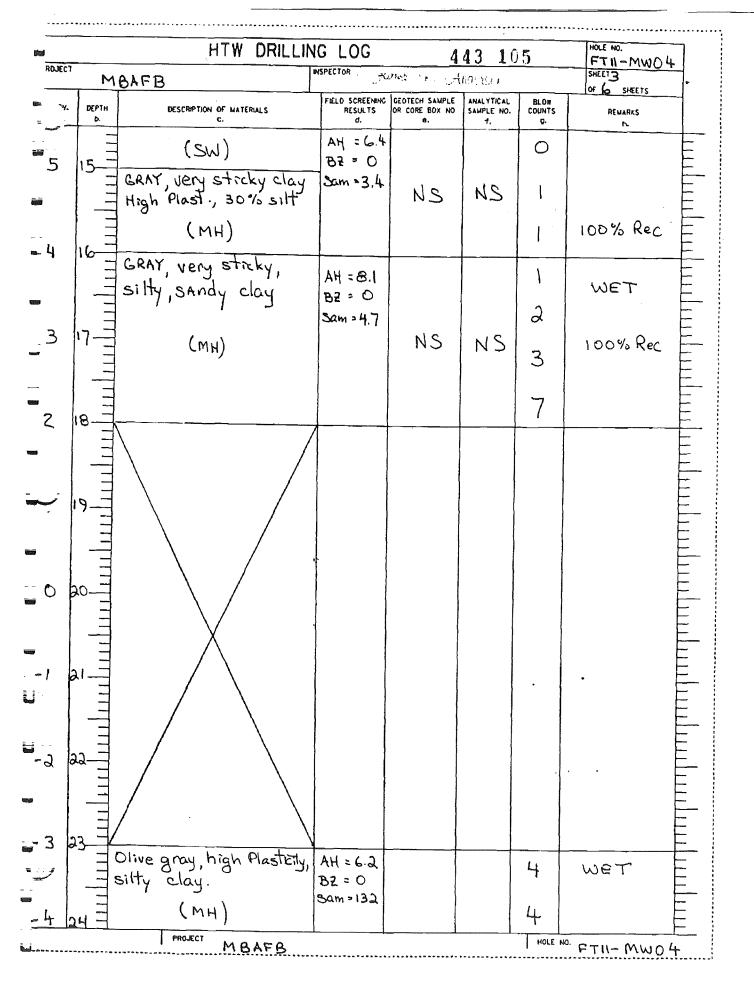
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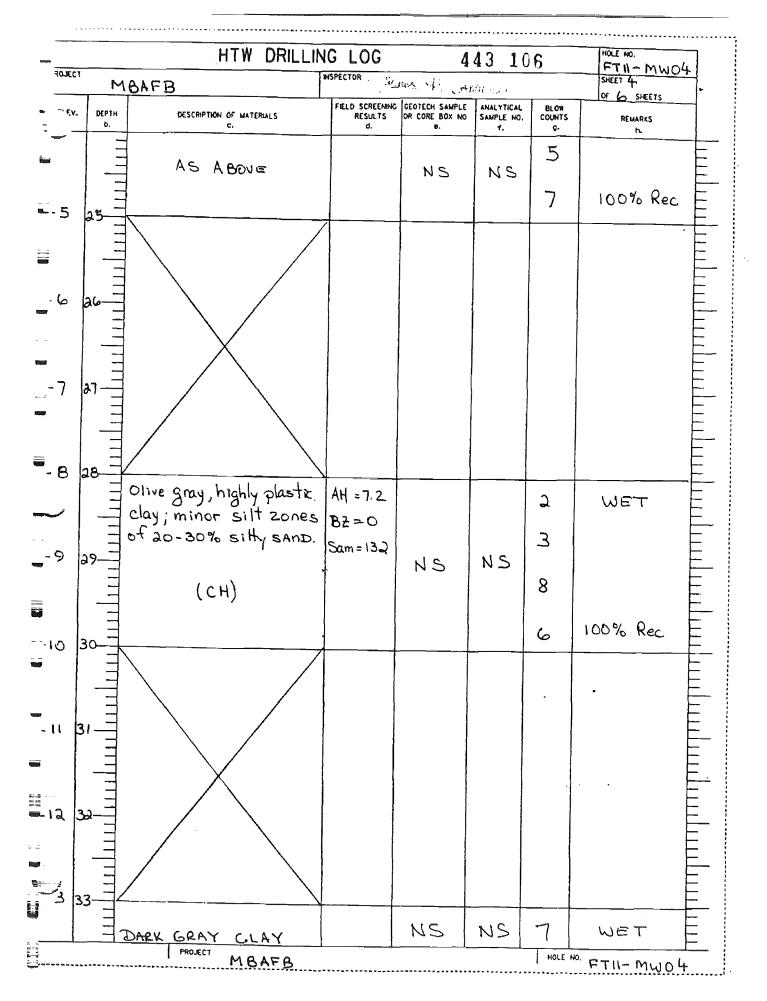
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alian a				HTW (	DRILLI	NG LO	OG					HOLE	NO. 11-MW04	┝
	COMPAN		bisco		2.	DRILLING SI		RACTOR				SHEET		
3.	PROJEC	т		NIG FORCE	Bice	. '	4. LOCA		1			<u></u>	2 302213	-
	ME C			AIR FORCE	ON2E		5. MANI	FACTURER'S D	ESIGN	ATION OF DRILL				-
7.	SIZES AN	UE NO TYPES O IMPLING EOL	FFREID	S.S. Split	Snoon	« I		LOCATION	•	<u>B-5</u>	(Na	55090		-
							15 9. SURI	S' Nort	<u>h</u>	of Mu	<u>105{</u> 52	1636005	.25	
							1	9.97			1			_
					···· •			E STARTED $12/17$	/9-	ع	1. DATE COM	1945		
	OVERSU	RDEN THICK	(NESS NA			1	5. DEP	TH GROUNDWAT	ER EN	NCOUNTERED	•	•		
13.	. DEPTH	DRILLED IN	TO ROCK				6. DEP	TH TO WATER	AND	ELAPSED TIME	AFTER DRILLIN	G COMPLET	ED	
. 14.	TOTAL	DEPTH DF	HOLE				17. OTH	ER WATER LEV	EL M	EASUREMENTS I	SPECIFY			_
18.	GEOTEC	HNICAL SAN	IPLES	DISTURBED	אט	DISTURBED	ľ	9. TOTAL NUME	ER O	F CORE BOXES	K. F.A.			_
20	D. SAMPL	ES FOR CH	EMICAL ANALYSIS	VDC	METAL	.5	OTHER	(SPECIFY)	01	HER (SPECIFY)	OTHER (S		21. TOTAL COR	
·•		NI	4										RECOVERY	
22	. DISPOS	SITION OF H	αε	BACKFILLED	MOHITORIN	G WELL	OTHER	(SPECIFY)	23.	SIGNATURE OF	INSPECTOR			1
L					X					Ad. E	7 705	6 <i>.</i> .		
•	ELEV. a.	ОЕРТН Б.	{	RIPTION OF MATERIALS C.		FIELD SCRE RESUM PID	LTS	GEOTECH SAN		ANALYTICAL SAMPLE NO. 4.	BLOW COUNTS Q.	F	IEMARKS	
 	-		Medium silty c	to dark bri lay.	own	AH = 21 BZ = 0					2	M	OIST	
				(SC)		3am=2	2.7	NC		115	7			
	19		Tan to b fine gra Fe streak	rown, round ined silly s	led, and.			NS		NS	10			
	18	د ا ا ا ا		(SM)	-						8	803	% Rec	
				rown to oro		AH = 15					7	m	oist	
				sandy silt.		82= C					©			<u> </u>
			20%	lay		Sam=3	9.B				පි			E
	71	3-11	•	(ML)				NS		NS	9	, .		
											12	50	% Rec	
	طا	4	Gray to Silly C	rust to or	ange	AH = 10 B2= (					7		OIST	
	5		•	(ML)		Sam=1		NS		NS	12			
			PR	MBA	FB	<b>.</b>			4	<b>k</b>	HOLE NO		- MW04	

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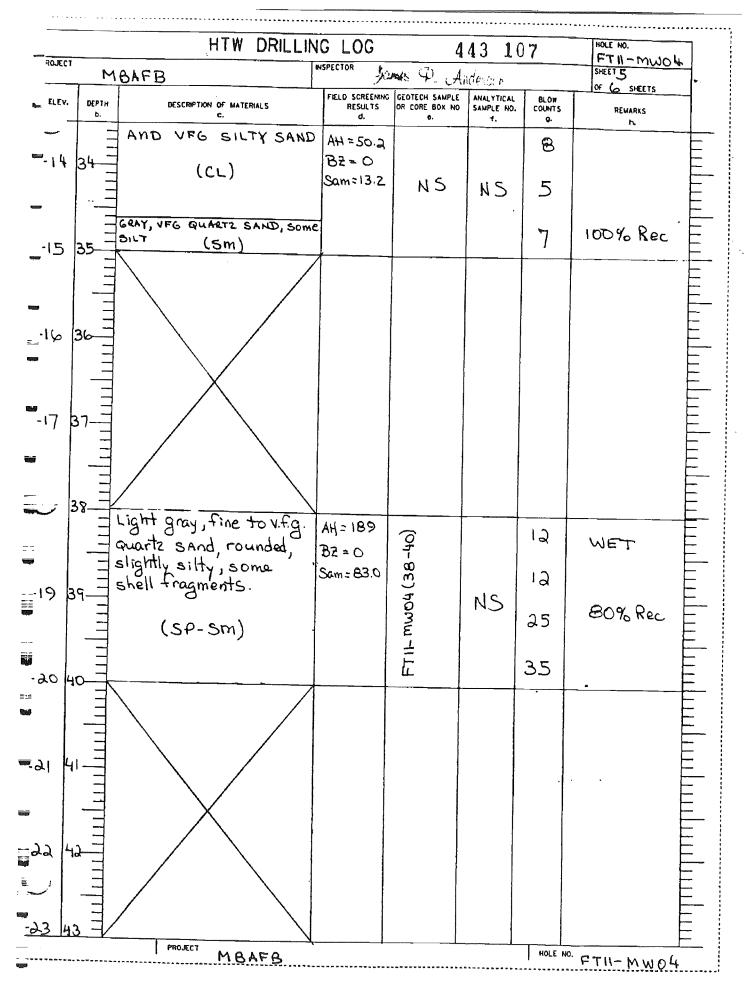


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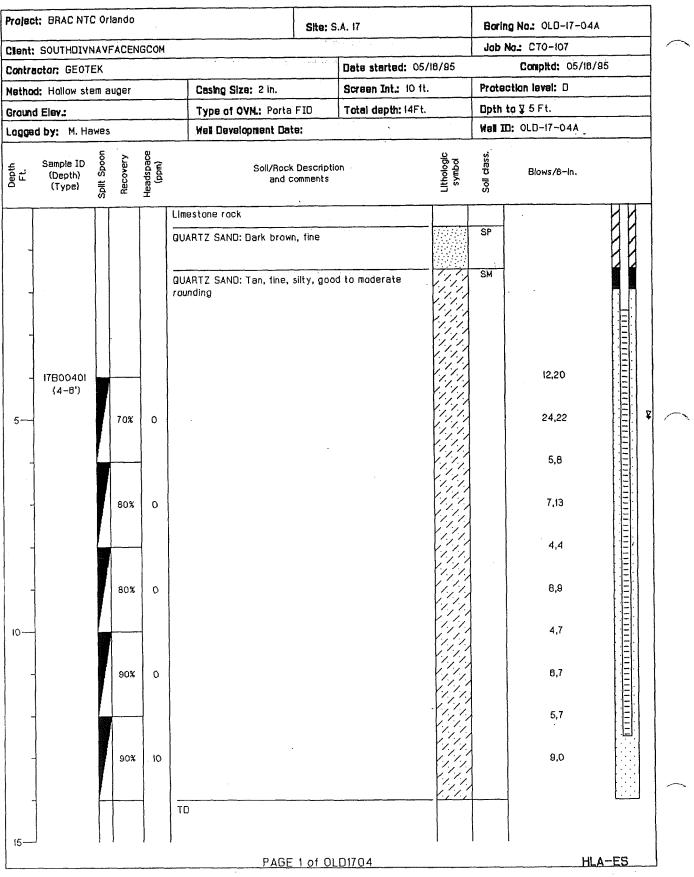


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		HTW DRIL	LING LOG			108	HOLE NO. FTII-MW04	7
ROJECT	Μ	BAFB	INSPECTOR	Y. All	F15. 1		SHEET 6 OF 6 SHEETS	
F I EV.	DEPTH D.	DESCRIPTION OF MATERIALS C.	FIELD SCREEMING GEO	OTECH SAMPLE CORE BOX NO e.	ANALYTICAL SAMPLE NO. 4.	BLOW COUNTS Q.	REMARKS	
ÿ		Light gray, f.g., guart; Sand, rounded, minor	shell $AH = 50.4$ BZ = 0		<b></b>	11	wet	
રન	44	frag. and mica. (Sm)	Sam - 98.6	NS	NS	ID .		
		OLIVE gray, Clay wi lenses of silty sand. Bottom High Plastic clay	th			1(		
25	45	(ОН	Total Depth =	- 45'		ol	80% Rec	
		AH = Auger Hold						
		AH = Auger Hold BZ = Breathing Z Sam = Split Spoor	one					
		Sam = Split Spool	s Sample					
~								
						-	•	
1								
******		PROJECT MBAFB				HOLE NO	FTII-MW04	and the first of the same

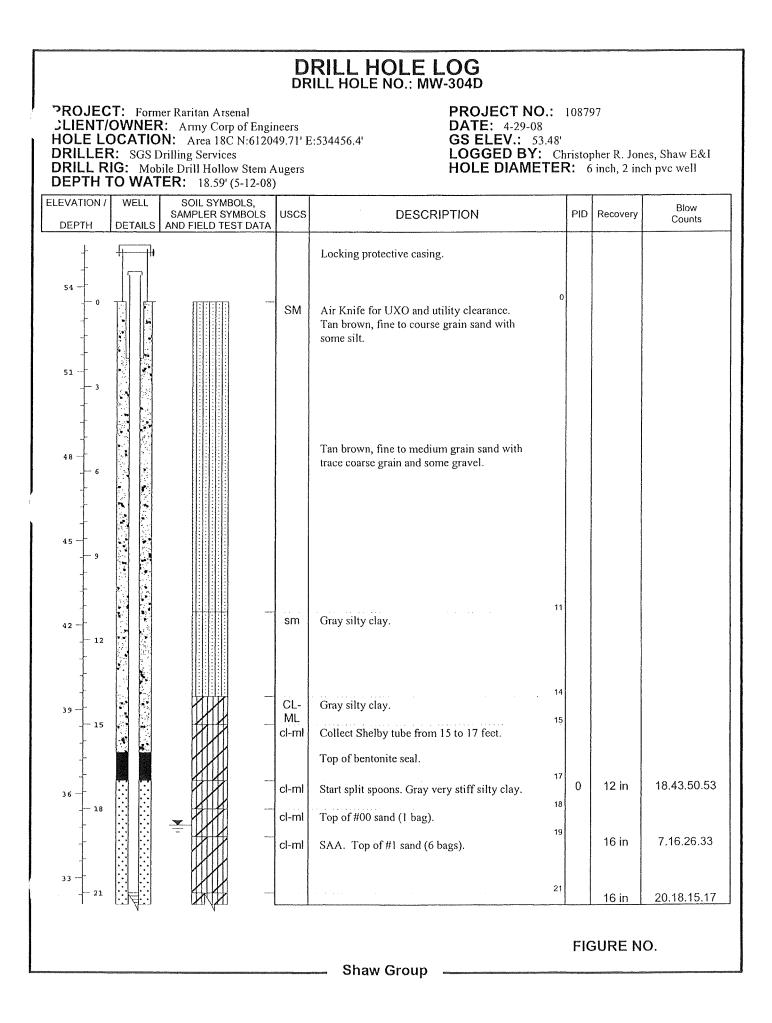
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Clent: SOUTHDIVNAVFACENGCOM						Site: S.A. 17			Job Na.: CTO 107		
			ALEN			Da	e sterted: 02/0	04/97		Compitd: 02/04/	/97
Contractor: Custom Nethod: SPT Casing Size: 2 in.						Screen Int: 5 ft.			Protection level: D		
Ground					Type of OVN.: FID		Total depth: BOFt.		Dpth to ¥ 8 Ft.		
	d by: JN				Well Development Date:				······	D: 0LD-17-10C	
				<i>a</i> )							
Depth Ft.	Sample ID (Depth) (Type)	Split Spoon	Recovery	Headspace (ppm)	Soli/Roc and	k Description comments		Lithologic symbol	Soll class.	Blows∕ò−in.	
					egin sample collection a .D-17-04 boring/well.	t base of exis	ing				
5											
-				0 -					SP	7,7,8,5	1
4			95%		ght gray silty fine sand	, wet.			Sr		Ц
-			95%	0						4,4,3,2	И
15			90%	0						3,2,2,10	H
· • •			10%								7
-				0					:	2,2,8,4	Ц
-			60%							2,2,5,5	H
20		Y	100%							2,2,0,0	7
				0						8,6,2,1	4
		1	95%								H
				- 0						1,1,1,2	7
25		1	70%	0 -						2,2,1,1	Ц
		V	100%		ight gray silty fine sand	d with clay, dr		: مند: النبا جو رجو ر	SP/SC		H
				0						5,8,7,11	7
-		1	90%								Z
зо—					ight gray silty line sam	 1, wet.		-	SP	- 4,4,8,7	4
		1	50%							0.488	Z
			50%	- 0	:					2,4,6,8	H
			,	- 0						5,4,5,10	P
35-		1	45%					1	1	1	Ы

Project: BRAC NTC Orlando Site: S.A. 17 Client: SOUTHDIVNAVFACENGCOM								Borin	ring No.: OLD-17-10C		
			ACEN	GCOM	· · · · · · · · · · · · · · · · · · ·	<u>a ja state da an</u>	<u></u>		Job	Na.: CTO 107	
Contractor: Custom							Date started: 0		Compitd: 02/04/97		
Nethod: SPT Casing Size: 2 in.						Screen Int.: 5 1		t.	Prote	atection level: □ th to ¥ 8 Ft.	
					Type of OVN.: FID						
r000e	d by: JN				Well Development Date	): 			Well II	J: 0LD-17-10C .	
Depth Ft.	Sample ID (Depth) (Type)	Split Spoon	Recovery	Headspace (ppm)	Soll/Rock and co Continued	mments		Lithologic symbol	Soll class.	Blows/6-in.	
			15% 0%	0					SP	ė,10,14,13	
-			0	0						5,4,2,4	
40			0%	0						5,4,5,5	
1		6	0%	0						8,8,7,4	
45—			0	0						NS	
-	ni, s., j	10	0%	0	Dark gray wandy clay, soft nodules atter 3".	ticlty, hard brown		SC	1,2,2,4		
-		100%	0			· · · · · · · · · · · · · · · · · · ·		1,0,0,1			
50		7	0	0						NS	
		10	0%							0,0,0,0	
55-		100	0%	-	een silty clay, stiff, low plasticity (H		(Hawthorne).	CL	3,2,2,1		
-		10	100%							8,10,13,13	
		10	0%	-	ireen silty clay.				CL/SC	2,3,4,8	
				T	٥						
- 95 -											
<u> </u>											
·o_]											



# DRILL HOLE NO.: MW-304D

PROJECT: Former Raritan Arsenal CLIENT/OWNER: Army Corp of Engineers HOLE LOCATION: Area 18C N:612049.71' E:534456.4' DRILLER: SGS Drilling Services DRILL RIG: Mobile Drill Hollow Stem Augers DEPTH TO WATER: 18.59' (5-12-08)

PROJECT NO.: 108797 DATE: 4-29-08 GS ELEV.: 53.48' LOGGED BY: Christopher R. Jones, Shaw E&I HOLE DIAMETER: 6 inch, 2 inch pvc well

	WELL ETAILS	SOIL SYMBOLS, SAMPLER SYMBOLS AND FIELD TEST DATA	uscs	DESCRIPTION	PID	Recovery	Blow Counts				
30			cl-ml SM SW SW	SAA. Top of well screen (10 slot).21.5Black-gray, silt with fine grain sand and22.75little clay.23Fine sand.23	4 0	5 in	14.10.11.9				
			sp	25 White-gray fine to medium grain sand with gray silt.	0	14 in	4.9.17.25				
27 27			SP sp	26.5 SAA with some trace gravel. 27 SAA. Black wood layer.	0	20 in	27.19.15.29				
24 - 30			sp ch	29 White-gray fine to medium grain sand with gray silt. 30 Start of Fire clay layer. Bottom of well.	0	15 in	15.27.18.36				
21 -				Start of The eray rayer. Bottom of wen.							
- 33											
18 36											
15											
12 - 42											
9											
	FIGURE NO Shaw Group										

PROJE PROJE LOCAT DRILLI SAMPI GROU TOP O LOGGI	ECT NUM ECT NAM TION ING METH LING MET ND SURF, F CASING ED BY	E <u>Naval V</u> Naval Weapo IOD <u>CME</u> IHOD <u>1.5'</u> ACE ELEVAT B ELEVATION	Irvine Phon Fax: ( <u>9-56254</u> Veapon ns Stat 75 Holl CA-Ma TON (F I (FT M nett	e, CA § e: (949) 4-622 is Stat ion-Se low St odified iT MS ISL)	02617 9) 752 725-39 5.001.1 ion-Se eal Bea em Au Split 9	ite 150 -5452 07 -K4.EQUIP BORING/WELL NUMBER AMM al Beach, Site 70 DATE DRILLED 9/5/07 CASING TYPE/DIAMETER 4"S ger SCREEN TYPE/SLOT 4"Sched Spoon GRAVEL PACK TYPE #2/16 Mc GROUT TYPE/QUANTITY Nea STATIC WATER LEVEL (FT BELO GROUND WATER ELEVATION (FT	/1 chedule ule 40 P nterey S t Cemen W TOC)	40 PVC VC 0.010-slot Slotted Screen and t Grout / Medium Bentonite Chp
PID (ppm)	BLOW COUNTS	RECOVERY (inches)	DEPTH (feet bgs)	U.S.C.S.	GRAPHIC LOG	LITHOLOGIC DESCRIPTION	CONTACT	WELL DIAGRAM
		-		SP		3" Asphalt cored (18" diameter) Hand augered to 5 feet bgs for utility clearance. <u> ~0.5' of road base (Silty Sand-Gravel mixture)</u> 1-4: SILTY SAND: Olive brown (2.5Y3/1); 60% fine to coarse, poorly graded, angular to subround sand; 30% micaceous silt; 10% clay, in "balls".	/ 0.2 / 0.7 /	-4" PVC slip cap Concrete Annular Seal Borehole Diameter = 11"
0.0/0.0	1,1, 3,4	20.4"/24" -	- 5			5-7: CLAY: Very dark gray (2.5Y3/1); 95% moderately plastic clay; 5% micaceous silt; wet.	5.0	
0.0/0.0	8,8, 7,13	24"/24" -	-	СН		7-8: Similar to above with plant and wood fragments. 8-9: Similar to above with Dark gray clay "balls"		15 feet of 4" Sch 40 PVC Blank Riser
0.0/0.0	1,1, 3,4	24"/24" -	- -10			9-9.6: Similar to above with layer of fine sand. 9.6-11.2: Similar to above with transition to Dark grayish brown Silty Clay with worm casings.	9.6	
0.0/0.0	6,9 3,5,7,	18"/24" - - 6"/24" -	-	CL ML		11.2-13: VERY SILTY CLAY: Light olive brown (2.5Y5/4); 60% moderately plastic clay; 30% micaceous silt; 10% fine to coarse, subangular to subround sand; <1% fine gravel; iron oxide mottling; wet; worm casings. 13-15: Similar to above; wet to saturated in center.		<ul> <li>+ Hydrated PureGold</li> <li>Medium</li> <li>Bentonite Chips</li> </ul>
0.0/0.0	7,8 3,5, 6,8	- 24"/24" -	- -15			15-16.5: CLAYEY SAND: Dark olive brown (2.5Y3/3) to dark greenish gray (GLEY4/5GY); 70% fine to medium,	15.0	
0.1/0.0	11,13, 17,17	24"/24" <sup>-</sup> -	-			subround sand; 20% stiff clay; 10% micaceous silt. 16.5-23: Similar to above with increasing fine sand and predominantly dark greenish gray in color. Similar to 15'-16.5'.		20 feet of 4"
0.0/0.0	3,5, 10,9	18"/24" <sup>-</sup> -	- -20	SP				SCH 40 PVC 0.010-slot Slotted Screen with Threaded
0.0/0.0	11,14, 11,13	20.4"/24" -	-					Couplings
0.4/0.0	7,10,5, 8,11	24"/24" - -				23-23.5: SILTY SAND: Olive brown (2.5Y4/3); 70% well graded, subangular to subround, fine sand; 30% micaceous silt; saturated.	23.0	
0.0/0.0	11,15, 17,22	24"/24" - _	-25—	SP		23.5-23.7: Dark brown (10YR3/3) sandy clay layer. 23.7-28: Similar to 23'-23.5' with some iron oxide mottling concentrated at bottom of section.		
0.0/0.0 0.4/0.0 0.0/0.0 0.0/0.0 0.0/0.0	NM	24"/24" -		SP		28-28.4: VERY SANDY CLAY: Light brownish gray	28.0	##2/16 Monterey Sand Filter Pack
0.0/0.0	5,10, 11,11	24"/24" - -	- - -30	SC SP SM		└ (2.5Y6/2); 55% plastic clay; 30% fine to coarse, └ subangular to subround sand; 10% micaceous silt; 5%	/ /_29.5 //	



111 Academy, Suite 150 Irvine, CA 92617 Phone: (949) 752-5452 Fax: (949) 725-3907

# **BORING/WELL CONSTRUCTION LOG**

PROJECT NUMBER PROJECT NAME

50999-56254-6225.001.TK4.EQUIP

BORING/WELL NUMBER \_\_\_\_\_AMW1\_\_\_ DATE DRILLED

		II		[	·1	Continued from Previous Page	1		
PID (ppm)	BLOW COUNTS	RECOVERY (inches)	DEPTH (feet bgs)	U.S.C.S.	GRAPHIC LOG	LITHOLOGIC DESCRIPTION	CONTACT DEPTH	WEL	L DIAGRAM
).0/0.0	11,12, 15,17	24"/24" - _		SP SC SP <del>SM</del> SP		<sup>11</sup> subangular, fine gravel; saturatedj 28.4-29.5: Similar to 23.7'-28' 29.5-31: Similar to 28'-28.4', but heavilty bioturbated 31-32: Similar to 23.7'-28' 32-33: Similar to 29.5'-31'.	31.0 32.0 33.0		♥#2/16 Monter Sand Filter Pack
0.0/0.0	4,7, 11,21	20.4"/24" _	_	SP		33-34: Similar to 31'-32'.	34.0		
).0/0.0	4,5, 7,11	24"/24" -	-35—_ -	SM SP ML SP SM		34-34.2: Similar to above with fine to medium, subangular to subround gravel. \ 34.2-35.1: VERYSANDY SILT: Olive (5Y4/3); 60% / \ micaceous silt; 35% well graded, subround, fine sand; / \< <u>10% clay; wet to saturated.</u> / 35.1-37: Similar to 33'-34' with iron oxide mottling.	34.2 35.1 37.0		<ul> <li>Threaded S0</li> <li>40 PVC Bott Cap Slough</li> </ul>
			- - 40 -			Total Depth of Boring = 36.5 feet bgs (with slough) Total Depth of Well = 35.1 feet bgs			
			-						
			45 - -						
			-						
			_						
			50 - -						
			-						
			-						
			55 - -						
			-						
			-						
			60 -						
			-						
			-						

# VISUAL CLASSIFICATION OF SOILS

BORING N	UMBER	IW1		PROJECT	NAME	SJCA Site 21							
PROJECT N		136517		COORDIN		N 3453248.		22836.2219		DAT			
ELEVATIO	N	9.85		GWL	Depth		Date/Time		STARTED		5/3/2014		
LOG BY		B. Roberts			Depth		Date/Time		COMPLET		5/3/2014		
DRILLING	METHOD	Direct Push							PAGE/P	AGES	1/1		
DEPTH (ft)	SAMPLE TYPE & NUMBER	RECOVERY (ft)	(mqq) DIq			USCS SYMBOL	R	EMARKS					
1				Concrete floor							Cored		
2				Gravel subbase m	naterial								
3	1	2.5	0	sandy SILT - brow	'n	ML/SM							
6				clayey SILT/silty C	CLAY - bro	own, stiff, dr	1		ML-CL				
7	2	3.0	0	SILT and SAND - t	prown, w	et			ML-SM				
9 10													
11 12 13 14 15	3	4.0	0	SAND - brownish	orange,	ittle silt, wet			SM/ML				
16 17				SAND - brown, lit	tle silt								
18	4	5.0	0	SAND - gray, wet					SP				
19 _20				CLAY - gray					CL				
NOTES: c = m = f =	very fine ntractor:	Parratt Wo Geoprobe 7 A. Chapel		End of boring = 2 ft = feet NA= not applic ppm = parts per r ack Rig	able				L I				

\*The free Adobe Reader may be used to view and complete this form. However, software must be purchased to complete, save, and reuse a saved form.

File Origi	hal with I	OWR				St	ate of Calif	iornia	r		DV	/R Use Only	- Do	Not Fill In		
- - 1					N	lell Co	mpleti	on Report								
		of $\frac{1}{\Gamma M}$				Refer	to Instruction	Pamphlet <sup>–</sup>		LL	Sta	e Well Num	ber/Si	le Number		
		nber <u>EW</u>					e008546	55				I N		W		
Date wor	к Began	10/21/2	008.	Date	VVOIK EI	nded <u>10/2</u>	2/2008									
Local Pel	mit Ager	псу										APN/TF	I RS/Oth	er		
	inder		· · · · · · · · · · · · · · · · · · ·					1	14 		1.87 11					
01	-	<ul> <li>Vertic</li> </ul>	Geolo	<b>gic Log</b> izontal	<u></u>	. Creek		1				Owner				
					Drilling I	e Specif	У	Name US Navy Caretaker Site Office								
Denth	from Su	rface	Auger	Des	cription			Mailing Address         Palm Avenue B-1, Suite 161           City         San Francisco         State         CA         Zip         94130								
Feet	to Fe	eet	Desc	ribe material,				City Sa	in Francis	co		State	<u>, CA</u>	<u></u>		
0	0	As	phalt							alaa Kaa	Well I	ocation				
0	2	Sa	indy clay - y	ellowish br	own			Address	Treasur	e Island	(forme	r naval ba	ise),	EW-29		
2	5	Sil	ty clay with	sand - bro	wn			City Sa	n Francis	со		Cour	ity Sa	an Francisco		
5	6	Cla	ay with silt -	brown										eg. Min. Sec.		
6	9	Po	orly graded	sand with	silt - oli	ve brown			Deg.	Min.	Sec.		D	leg. Min. Sec.		
9	10	Cla	ay with sand	and silt -	dark gre	enish gra	ау	Datum_		Decimal	Lat		Deci	mal Long		
10 -	30	Po	orly graded	sand with	silt - oli	ve brown		APN Bo	ok	Page			Parce	mal Long 키 on		
								Townsh	ip	Range	)i	handari -	Secti	on		
									Locat	ion Ske	tch			Activity		
								(Sketch	must be drawn	by hand aft North	er form is	printed.)	<b>O</b> N	Activity ew Well		
					·····				h Stree	K			ОM	odification/Repair Deepen		
			,					15-64	N 21.	T			C	Other		
									÷.,	1			O D	ESTROY escribe procedures and materials ider "GEOLOGIC LOG"		
								11 \		1			ur	escribe procedures and materials ider "GEOLOGIC LOG"		
-										$\mathbf{V}$				Planned Uses		
										1	1			ater Supply		
<b></b>	_									12	21'			Domestic Dublic		
								West				East		Irrigation 🔲 Industrial		
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<b> </b>			······		1997 1975									emediation		
							· · · · ·		O Sparging							
			· ••••			•.**	5-1 1-1	-11	est Well							
								1         South         O         Vapor Extraction           Illustrate or describe distance of well from roads, buildings, fences, etc. and attach a map. Use additional paper if necessary.         O         Vapor Extraction								
								rivers, etc. ar Please be ac	d attach a map. curate and com	Use additional plete,	paper if nec	essary.	<b>O</b> 0	ther		
								Water I	.evel and	Yield o	f Com	pleted We	all 🗌			
				1.5.5 1917 - 1917										t below surface)		
				<u>ila da.</u> Tang				L Donth to	Static							
			······································			<u> </u>								red		
Total De	epth of B	oring	30			Feet		11	ed Yield *			A) Test T		<u></u>		
Total De	epth of C	ompleted	Well 30			Feet		Test Le				rs) Total D				
L				131				L no	t be repres	entative	or a wel		17 11 11 1			
Danil	from	Borehole	en en en en	Casi		Woll.	Outoide	Coroce	Slot Size	Depth	<u>i from</u>	Annula	rMat	erial		
Depth Surf		Diameter	Туре	Mater	ial	Wall Thickness	Outside Diameter	Screen Type	Slot Size if Any		i from face	Fill		Description		
Feet to		(Inches)	15			(Inches)	(Inches)		(Inches)		o Feet			-		
	10	10	Blank	PVC Sch. 40		<u> </u>		-1-6-	0.000	0	4 7	Cement				
10	30	10	Screen	PVC Sch. 40	)			slots	0.020	4	-	Bentonite	-	2/12 cond		
di la								1	30	Filter Pack		2/12 sand				
												<u>}</u>				
		A 44 c - h		 	P		<b> </b>						a gradati	l Synapod Andrew Content of Angel		
	Poole-1	Attachn	ients		1 the m	ndereigner	l cortific th		Certificati			the host	of my	knowledge and bolief		
	Geologic	Log Istruction I	Diagram		Name	Shaw En	vironmen	tal Inc.	WIX 4	xplor	ation	" W	e/1's	knowledge and belief		
		ical Log(s	•			Person, I	Firm or Corpor	ration				•				
			/ al Analyses		4000	Port Chic	Address		Con	City	1		e	24520		
	Other <u>S</u>	ite Map	•		Signed				HRIS TA	<u>WM</u>	1/29	/09		3326		
Attach addi DWR 188 F		nation, if it ex	ists.		L		ensed Water V	Vell Contractor			Date Si	ned C-	57 Lic	ense Number		

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E IS NEEDED, USE NEXT CONSECUTIVELY NUMBERED F

	L	ABORA	TORY	TEST	DATA	2	S) (S			WELL NO. 15MW10
DEPTH IN FEET OVA READING	(ppm) PERMEABILITY	BENZENE	TOLUENE (ppm)	XMENE (ppm)	TRIAR DETROIEUM TRIAR DECKROONS (PPm)	WELL SUMMARY, BACKFILL	PENETRATION RESISTANCE (BLOWS/6 INCHES) 	nscs	PROFILE	$\begin{array}{c} \begin{array}{c} \text{COORDINATES} \underbrace{N & 2121181.12}_{\underline{E} & 5781272.40} \\ \hline \\ \hline \\ \hline \\ \hline \\ \hline \\ \hline \\ \hline \\ \hline \\ \hline \\ $
0							2 5 6 7 0024-11 5 10 14 22 5 6 7 7 0022-21 11 6 4	58° -17 CL 58°		DESCRIPTION Brownish yellow (10YR %) SAND, medium groin, dry, well sorted. Toroundwater measured at 10.92 ft. on 11/08/00 Brownish (7.5Y 4/4) SAND, domp, well sorted. Brownish (7.5Y 4/3) SAND, wet, medium groined, minor sit. Light olive brown (7.5Y %) SAND, wet, fine groined, well sorted. Light yellowish brown (2.5Y %) gravelly SAND, medium groined to 26.5 26. Dark greenish brown (2.5Y %) sally CLAY. 28.0 Light brownish groy (2.5Y %) SAND, medium groin, very well sorted-gravel at 32 ft. Growel cobbles on center plug. Light yellowish brown (2.5Y %) SAND, medium groin well sorted-some os at 30 ft. 39.0 Bedrock shole of 39 feet. TOTAL DEPTH = 39.5 FEET

PROJECT NO. 779480-3D210691 CLIENT: Site 15 Vandenberg AFB SEE LEGEND FOR LOGS AND TEST PITS FOR EXPLANATION OF SYMBOLS AND TERMS

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INTERNATIONAL TECHNOLOGY CORPORATION

Project Nerve Vondenberg AFB       IIIIIIIIIIIIIIIIIIIIIIIIIIIIIIIIIIII	<u>12-3-01</u> <u>12-3-01</u> CE EL. <u>47.40</u> ' very well-sorted
Method AUger	<u>12-3-01</u> <u>12-3-01</u> CE EL. <u>47.40</u> ' very well-sorted
Method AUger	<u>12-3-01</u> CE EL. <u>47.40'</u> very well-sorted
Method AUger	<u>12-3-01</u> CE EL. <u>47.40'</u> very well-sorted
Method AUger	<u>12-3-01</u> CE EL. <u>47.40'</u> very well-sorted
Method AUger	CE EL. <u>47.40</u> '
Method AUger	very well-sorted
Method AUger	10.07
No. H 4/0 Lovel	10.07
Static Hole open_45         Total Hele Depti_45         Borshole Dia_12*         Fill Methods Bentonite-Comment grout         Screen Dia_4* Sch.40 PVC/0.02* stot         Depth Hele Open_45         Depth Hele Open_46         Depth Hele Open_46         Depth Hele Open_46         15         File PackTopeh (12): Lonsics cond/ 13 = 20 = 10         15         File PackTopeh (2): Lonsics cond/ 13 = 20 = 10         15         File Open Hele Open_47         15         File Open_47         15         15         15         16         15         16         16         15         16         16         17         18         18         19         10         10	30.0" very well
Total Hole Dopti. 45       5       5         Borshole Da12"       5         Fill Material Bentonite-Content grout       10         Scroen Dia.4" Sch.40 PVC/0.02" alot       10         Dopti historial 6'-43'       10         Depti historial 0'-18'       15         Fill Material 0'-18'       15         Fill Provide the provided of the	<u>10.0</u> 7 very well
Boraholo Da       12"         Fill Methods Bentonite-Cement grout	10.0"
Borshole Dia_12"       -         Fill Methods Bentonite-Connell grout       -         Screen Dia_4" Sch.40 PVC/0.02" stot       -         Dopth Interval       6'-43'         Depth Interval       -         Depth Interval       -         Bent/Depth Interval       -         Filler       -         Dopth Interval       -         Seal/Dopth Tipe       -         Filler       -         25-       -         Seal/Dopth Tipe       -         NOTE:       -         -       -         -       -         -       -         -       -         -       -         -       -         -       -         -       -         -       -         -       -         -	<u>30.0</u> ' very well
Fill Metodel Bentonite-Cement grout       Image: Sch.40 PVC/0.02° stot         Screen Dia 4° Sch.40 PVC/0.02° stot       10°         Dopth Interval 8°-43′       Image: Sch.40 PVC         Dopth Interval 8°-43′       Image: Sch.40 PVC         Dopth Interval 8°-43′       Image: Sch.40 PVC         Dopth Interval 9°-43′       Image: Sch.40 PVC         Dopth Interval 0°-18′       Image: Sch.40 PVC         Dopth Interval 0°-18′       Image: Sch.40 PVC         Sch.40 PVC       Image: Sch.40 PVC         Dopth Interval 0°-18′       Image: Sch.40 PVC         Sch.40 PVC       Image: Sch.40 PVC         Dopth Interval 0°-18′       Image: Sch.40 PVC         Sch.40 PVC       Image: Sch.40 PVC         Image: Sch.40 PVC       Image: Sch.40 PVC         Sch.40 PVC       Image: Sch.40 PVC         Image: Sch.40 PVC       Image: Sch.40 PVC         Image: Sch.40 PVC       Image: Sch.40 PVC         Image: Sch.40 PVC       Image: Sch.40 PVC         Image: Sch.40 PVC       Image: Sch.40 PVC         Image: Sch.40 PVC       Image: Sch.40 PVC         Image: Sch.40 PVC       Image: Sch.40 PVC         Image: Sch.40 PVC       Image: Sch.40 PVC         Image: Sch.40 PVC       Image: Sch.40 PVC         Image: Sch.40 PVC	very well
Screen Dia 4" Sch. 40 PVC/0.02" stot Dopth Interval 8'-43' Cesting Dia. 4" Sch. 40 PVC Dopth Interval 0'-18' Tis- Filter Pack/Dopth [7/12 (oncitor sond/ Seried, dune sond. 15- Filter Pack/Dopth [7/12 (oncitor sond/ Seried, dune sond. 15- Seried, dune	very well
Dopth Interval 8-43'	very well
Dopth Interval 8-43	÷
Doph Herval 0-18'       15-         Filter Pack/Doph/{2/12 (uncticr sond/ 5.9-45)       15-         Seal/Dept/Type Biotorite chips/ 1.3-5.9       -         NDIE:       -20-         -25-       -         -30-	<u>.</u>
Dopth Herval 0-18'       15-         Filter Pack/Dopth/T/12 (concilor exold/ 5.9-45'       15-         Beal/Depth/Type Biotonite chips/ 1.3-5.9       -20-         NOTE:       -20-         -25-	<u>.</u>
Filer Pack/Dopth [2/12 Uncefor and/ 5.9 - 46 Beal/Depth/Type Benlmite chips/ 1.5 - 5.9 NDIE: 20- 20- 20- 20- 20- 20- 20- 20-	÷
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Seed/DeptivType Bentaulite chies/ 1.5-5.9 NOTE: NOTE: Seed/DeptivType Bentaulite chies/ -20- -20	• [
NOTE: 20 20 20 20 20 20 20 20 20 20	
Loose, yellowish brown (10)R 5/4) SAND, dry, medium grais sorted well rounded quartz grains.	20.0
sorted well rounded quartz grains.	
25- 	od, vory well
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50 	1
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-     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -	
-     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -       -     -     -     -	·
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Loose, very dark grayish brown (10YR 3/2) SAND, wat, me	1
Loose, very dark grayish brown (10YR 3/2) SAND, wat, me	
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Coose, very dork gravish brown (10YR 3/2) SAND, wet, me 40- 40- 50- 50- 50- 50- 50- 50- 50- 5	
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or     -40-     <	ium proined
	in the sand.
三 (Top of clay uncertain)	42.0
	L
h Hedium dense, Gray (SY 5/1) SHALE, thinly bedded, possibl	crushed.
COTAL DEPTH = 45.0 FEET	
45     45       -45     -45       -50       -50       -55       -60       -60       -60       -60       -60       -60       -60       -60       -60       -60	
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PROJECT NO. 779480 CLIENT: VANDENBERG AFB SEE LEGEND FOR LOGS AND TEST PITS FOR EXPLANATION OF SYMBOLS AND TERMS

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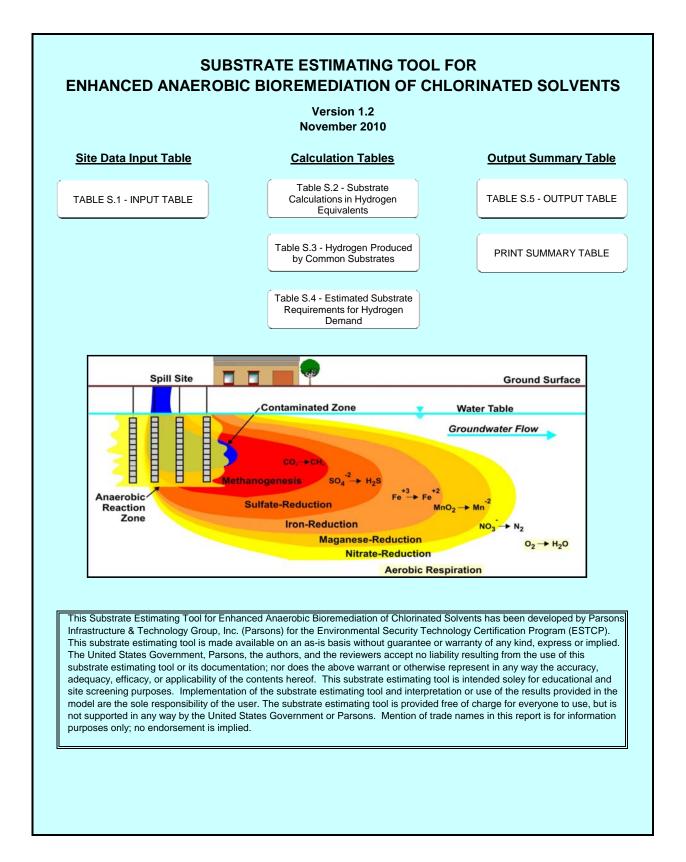


Borehole Diam. (in.): 10.0		Total	Dent	h (ft):	30.0			Boring ID:19-MW-17A/B         Well ID: 19-MW-17A/B           Project:         VAFB TO 106									
Northing (ft):		-	ng (ft		00.0			-		950986.06010707	Site: IRP	Sito 10			····-		
Drill Start Date: 03-28-200	)6			: 09:	:40				bigged By: GMB Reviewed By:								
Drill Finish Date: 03-28-200		1		e: 12:		*******		Drilling Contractor: S & G Drilling Field Instrumentation: PID									
Depth 1st H <sub>2</sub> O (ft):				e: 03-28-2006 09:57 Drill Rig Type/Method: Hollow Stem Auger											~		
Depth H <sub>2</sub> O After Drilling (ft):				e: N/A						Tim Blakeboro, Tim McNa							
Comments:										Well Comp. Date: 0		Completion	Tím	e:	13:5	0	
Samplers: continuous CA			it-spo	on						Soil Backfill Date: N		Backfill Tim			N/A		
	al								-						. % (	of S	oi
Well Completion	Sample Interval Retained	Sample Type	Recovery (%)	Blow Count/6"	(mqq) Olq	Water Level	ېDepth (feet)	Graphic Log	USCS Soil Classification	Descr	iption		Gravel	Coarse Sand	Med. Sand	Fine Sand	
							0-	-	SP	<u>Surface: asphalt</u> Poorly Graded Sand, red loose, moist, no odors	dish brown (5	SYR 4/3).			70	30	)
Hydrated bentonite chips(0'-6') Filterpack: #1C								-	SP-SM	Poorly Graded Sand with (5YR 4/3), loose, moist, r		brown			60	30	
Filterpack: #1C		- I ·	100 100 100	3 3 3	1.7		5-		CL-ML	Silty Clay, black (GLEY1 moist, no odors	2.5/N), mediu	ım stiff,					1
			100 100 100	4 1 1	5.1	¥			СН	Fat Clay, yellowish brown moist, plastic, no odors	(10YR 5/4),	very soft,					1
(6'-17')			100 100 100	2 2 1	4.8				MH MH	Elastic Silt, yellowish brow wet, plastic, no odors Elastic Silt, yellowish brow					<u> </u>	 	1
		1	100	2			10-			plastic, no odors	-	-					
			100 100 100	1 1 2	7.3		•		СН	Fat Clay, dark grayish bro soft, moist, fat CLay, stick	wn (10YR 4/: xy, no odors	2), very					1
2 Pipes, One screen: 0.010" factory slotted liner (7'-17')			100 100 100 100	4 1 1 2	6.2				сн	Fat Clay, brown (10YR 4/ odors	3), very soft, v	wet, no					1
liner (7'-17')		1	100	2 3 1	3.3		-	ÍÍÍ	МН	Elastic Silt, brown (10YR odors	4/3), very soft	t, wet, no					1
			50 100	1			15-		CH	Fat Clay, brown (10YR 4/	3), very soft, v	wet, no					1
		1	100 100 100	1 2 3	3.5		-		SM	\odors Silty Sand, brown (10YR - odors	1/3), loose, w	et, no				70	
2' Bentonite seal (17'-19')		1   1	00 00 00 00	1 2 3 4 1	3.1		-		СН	Fat Clay, yellowish brown stiff, moist, no odors	(10YR 5/4), r	nedium					1
Filterpack: #3 Monterey Sand (19'-30')		1	00 00 00	1 1 2	3.0		20-		SM	Silty Sand, yellowish brow loose, wet, no odors	n (10YR 5/4)	very				70	3
		1	00 00 0	6 7 1	4.8		-		SW	Well Graded Sand, light ye 6/4), medium dense, wet,			tr	65	25	10	
creen: 2", 0.020"—		1	33 00 00	4 9	2.4		-			Becomes finer grained at a				55	40	5	
factory slotted screen (20'-28')		1	0 00 0	8 2	3.4		25			Becomes dark grayish bro occasional large Gravel at	wn (10YR 4/2 24'	?), with		50	45	5	
		1		4 14 17 2	2.9		-			Becomes dense at 26'							
		3	33 00		3.2		-										
		1(	00		6.0		30-			Becomes very coarse grain Becomes more Gravely at					35 30	5 5	
1	1									Total depth 30'						- 1	

<b>Tetra Tech, Inc.</b> 4213 State Street, Suite 100 Santa Barbara, CA 93110-2847 VOICE (805) 681-3100 FAX (805) 681-3108	Project: Site 32 Cluster Groundwater Investigation Log of Boring No. 35-MW-30 Location: OU4, Site 32C, approximately 120 feet northwest of soil boring location									
VOICE (805) 681-3100 FAX (805) 681-3108	35-MIP-6.           Northing: 2130769.74         Easting: 5791152.77									
Date Started: 1/19/07 Completed: 1/19/07	2" Sch. 80 PVC Blank Casing: from 23.0 ft bgs to 0.0 ft bgs									
Logged By: M. Houlahan Checked By Def 16#6962	2" Sch. 80 PVC Screening Casing: from 33.0 ft bgs to 23.0 ft bgs									
Drilled By: Ramon Zepeda / BC <sup>2</sup> Environmental Corp.	Bentonite/Cement Grout: from 19.0 ft bgs to 1.0 ft bgs									
Drilling Equipment: CME 95 HSA	Bentonite Transitional Seal: from 21.0 ft bgs to 19.0 ft bgs									
Sampling Equipment: 2" Stainless Steel CA Split Spoon Sampler	Sand Filter Pack: from 36.5 ft bgs to 21.0 ft bgs									
GS Elev.: 311.68 ft above MSL TD: 36.5 ft bgs	TOC Elevation: 313.63 ft above MSL									
က်ကို MATERIAL DESCRIPTION	Material Material Type Blow PiD (ppm) PiD (ppm) PiD (ppm) Amb/Smp									
<ul> <li>5 Sand, pale brown (10YR 6/3), fine grained, subrounded, poorly graded, medium dense, trace silt, trace gravel [&lt;10 mm]. cemented, dry.</li> <li>10 As above, few clay.</li> </ul>	SP SP 10 12 14 100 0.0/ 0.0 0.0/ 0.0 0.0/ 0.0 1330 Begin drilling. 1330 Hegin drilling.									
<ul> <li>15 Sand with silt, brown (10YR 5/3), fine grained, subrounded, poorly graded, 10 percent silt, trace clay, dry.</li> <li>20</li> </ul>	$ \begin{array}{c ccccccccccccccccccccccccccccccccccc$									
	26.5 ft bgs $\sum$ Groundwater depth after well development on: <u>03/08/2007</u>									
rC Number: 99062-06	TETRA TECH, INC. Page 1 of 2									

6	ł	<b>Tetra Tech, Inc.</b> 4213 State Street, Suite 100 Santa Barbara, CA 93110-2847 VOICE (805) 681-3100 FAX (805) 681-3108	Log c Locati 35-MI	of Borin on: OU4 P-6.		) mately 12	0 feet n	orthwes	t of soil boring location
Depth (ft. bgs)		MATERIAL DESCRIPTION	Material Symbol	Type Tigu	Well Construction	Easting: 5 Blow Der 6"		PID (ppm) Amb/Smp	REMARKS
20		<u>Clayey sand</u> , dark grayish brown (10YR 4/2), fine grained, subrounded, poorly graded, 10 percent silt, slightly moist.		SC		10 12 14	100	0.0/	
									Note - screen slot size is 0.020".
25		As above.		$\bigtriangledown$		13 16 18	100	0.0/ 0.0	
		As above.		Ā		14 18 20	100	0.0/ 0.0	
		As above. 40 percent fragments of fractured shale. <u>Clayev gravel with sand, grayish brown (10YR 5/2), fine</u> gravel, subrounded, poorly graded, 25 percent clay, 15		GC			75		
30		percent sand, slightly moist. <u>Clayey sand</u> , grayish brown (10YR 5/2), fine grained, subrounded, poorly graded, 30 percent clay, wet.		sc ¥ GC		50	33	0.0/ 0.2	
		Clavey gravel with sand, grayish brown (10YR 5/2), fine gravel. subrounded, poorly graded, dense, 25 percent clay. 15 percent sand. slightly moist.		SC		22 50	67		
	_	<u>Shale</u> , very pale brown (10YR 7/3), dry, Sisquoc Shale.		Shale		33 50	67		
35		As above.			X				1510 End drilling.
40									
40									
45									
тс	Nun		6.5 ft I TETRA	· -	•	lepth afte	r well	develo	pment on: <u>03/08/2007</u> Page 2 of 2

Appendix G: Loading Calculations



	_			
Site Name:	Treatment 1			RETURN TO COVER PAGE
	NOTE: Unshaded		•	
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes
Width (Perpendicular to predominant groundwater flow direction)	60	1-10,000	feet	
Length (Parallel to predominant groundwater flow)	140	1-1,000	feet	
Saturated Thickness	22	1-100	feet	
Treatment Zone Cross Sectional Area	1320		ft <sup>2</sup>	
Treatment Zone Volume	184,800		ft <sup>3</sup>	
Treatment Zone Total Pore Volume (total volume x total porosity)	345,668		gallons	
Treatment Zone Effective Pore Volume (total volume x effective porosity)	276,535		gallons	
Design Period of Performance	1.4	.5 to 5	year	
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3
2. Treatment Zone Hydrogeologic Properties				
Total Porosity	25%	.05-50	percent	Default = 25%
Effective Porosity	20%	.05-50	percent	Default = 20%
Average Aquifer Hydraulic Conductivity	0.01	.01-1000	ft/day	
Average Hydraulic Gradient	0.05	0.0001-0.1	ft/ft	
Average Groundwater Seepage Velocity through the Treatment Zone	0.00		ft/day	
Average Groundwater Seepage Velocity through the Treatment Zone	0.9		ft/yr	
Average Groundwater Discharge through the Treatment Zone	1,802		gallons/year	
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors				
Oxygen	1.3	0.01 to 10	mg/L	Default = 5
Nitrate	0.22	0.1 to- 20	mg/L	Default = 1
Sulfate	5920	10 to 5,000	mg/L	Default = 50
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0
B. Solid-Phase Native Electron Acceptors				
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.022	mg/L	
Trichloroethene (TCE)	15.146	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.393	mg/L	
Vinyl Chloride (VC)	0.004	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)		mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

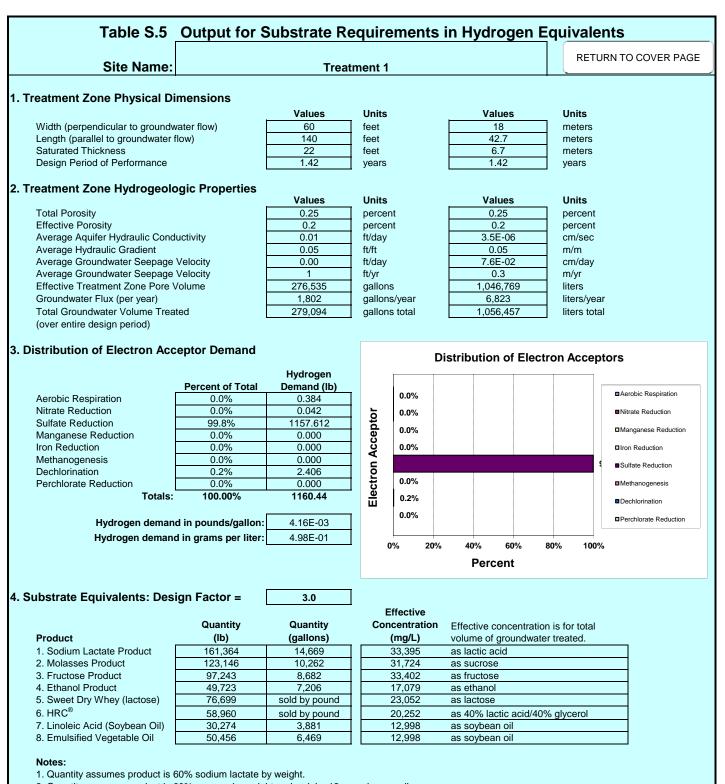
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ / quoodo ecconomical y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

#### **B. Aquifer Matrix**

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as $CaCO_3$	Default = 10%



2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.

3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.

4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.

5. Quantity assumes product is 70% lactose by weight.

6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

8. Quantity assumes commercial product is 60% soybean oil by weight.

# Treatment 1

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		14669.43659
2. Molasses Product		10262.16968
3. Fructose Product		8682.435413
4. Ethanol Product		7206.160231
5. Sweet Dry Whey (lactose)	76699.08034	
6. HRC®	58959.61468	ł
7. Linoleic Acid (Soybean Oil)		3881.262855
8. Emulsified Vegetable Oil		6468.771425
9. Lactoil Product		3105.010284
10. Lactic Acid Product		0
11. Hydrogen Gas	1160.44	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents					
Site Name:	Treatment 2			RETURN TO COVER PAGE	
	NOTE: Unshaded	boxes are use	r input.		
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes	
Width (Perpendicular to predominant groundwater flow direction)	180	1-10,000	feet		
Length (Parallel to predominant groundwater flow)	40	1-1,000	feet		
Saturated Thickness	22	1-100	feet		
Treatment Zone Cross Sectional Area	3960		ft <sup>2</sup>		
Treatment Zone Volume	158,400		ft <sup>3</sup>		
Treatment Zone Total Pore Volume (total volume x total porosity)	296,287		gallons		
Treatment Zone Effective Pore Volume (total volume x effective porosity)	237,030		gallons		
Design Period of Performance	1.4	.5 to 5	year		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
2. Treatment Zone Hydrogeologic Properties					
Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity	20%	.05-50	percent	Default = 20%	
Average Aquifer Hydraulic Conductivity	0.01	.01-1000	ft/day		
Average Hydraulic Gradient	0.05	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	0.00		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	0.9		ft/yr		
Average Groundwater Discharge through the Treatment Zone	5,407		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors					
Oxygen	0.8	0.01 to 10	mg/L	Default = 5	
Nitrate	0.10	0.1 to- 20	mg/L	Default = 1	
Sulfate	3643	10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0	
B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.077	mg/L	
Trichloroethene (TCE)	14.646	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.087	mg/L	
Vinyl Chloride (VC)	0.000	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)		mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

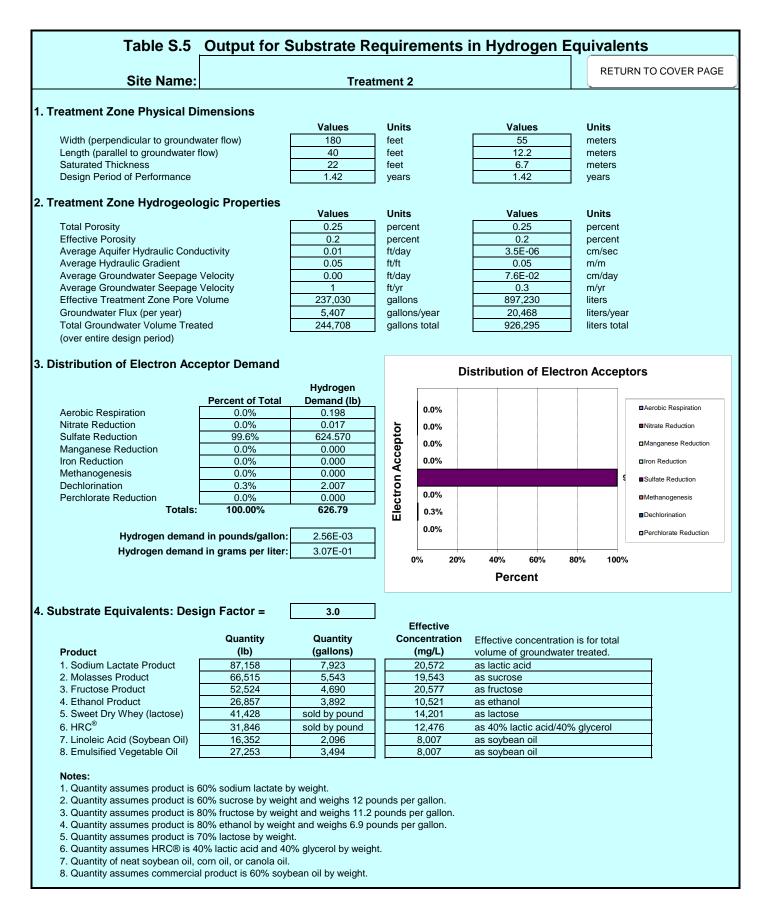
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



# Treatment 2

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		7923.419618
2. Molasses Product		5542.917488
3. Fructose Product		4689.653803
4. Ethanol Product		3892.271595
5. Sweet Dry Whey (lactose)	41427.56228	
6. HRC®	31845.92434	
7. Linoleic Acid (Soybean Oil)		2096.39096
8. Emulsified Vegetable Oil		3493.984933
9. Lactoil Product		1677.112768
10. Lactic Acid Product		0
11. Hydrogen Gas	626.79	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

	Table S.1 Input for Substrate Requirements in Hydrogen Equivalents				
Site Name:	Treatment 3			RETURN TO COVER PAGE	
	NOTE: Unshaded	boxes are use	r input.		
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes	
Width (Perpendicular to predominant groundwater flow direction)	200	1-10,000	feet		
Length (Parallel to predominant groundwater flow)	320	1-1,000	feet		
Saturated Thickness	22	1-100	feet		
Treatment Zone Cross Sectional Area	4400		ft <sup>2</sup>		
Treatment Zone Volume	1,408,000		ft <sup>3</sup>		
Treatment Zone Total Pore Volume (total volume x total porosity)	2,633,664		gallons		
Treatment Zone Effective Pore Volume (total volume x effective porosity)	2,106,931		gallons		
Design Period of Performance	3.2	.5 to 5	year		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
2. Treatment Zone Hydrogeologic Properties					
Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity	20%	.05-50	percent	Default = 20%	
Average Aquifer Hydraulic Conductivity	0.01	.01-1000	ft/day		
Average Hydraulic Gradient	0.05	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	0.00		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	0.9		ft/yr		
Average Groundwater Discharge through the Treatment Zone	6,008		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors					
Oxygen	0.0	0.01 to 10	mg/L	Default = 5	
Nitrate	1.46	0.1 to- 20	mg/L	Default = 1	
Sulfate	1975	10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0	
B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

# 4. Contaminant Electron Acceptors

0.063	mg/L	
9.150	mg/L	
1.649	mg/L	
0.019	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.002	mg/L	
	mg/L	
	mg/L	
	mg/L	
	9.150 1.649 0.019	9.150        mg/L         1.649        mg/L         0.019        mg/L          mg/L           mg/L           mg/L           mg/L           mg/L           mg/L          0.002        mg/L          mg/L          0.002        mg/L          mg/L           mg/L

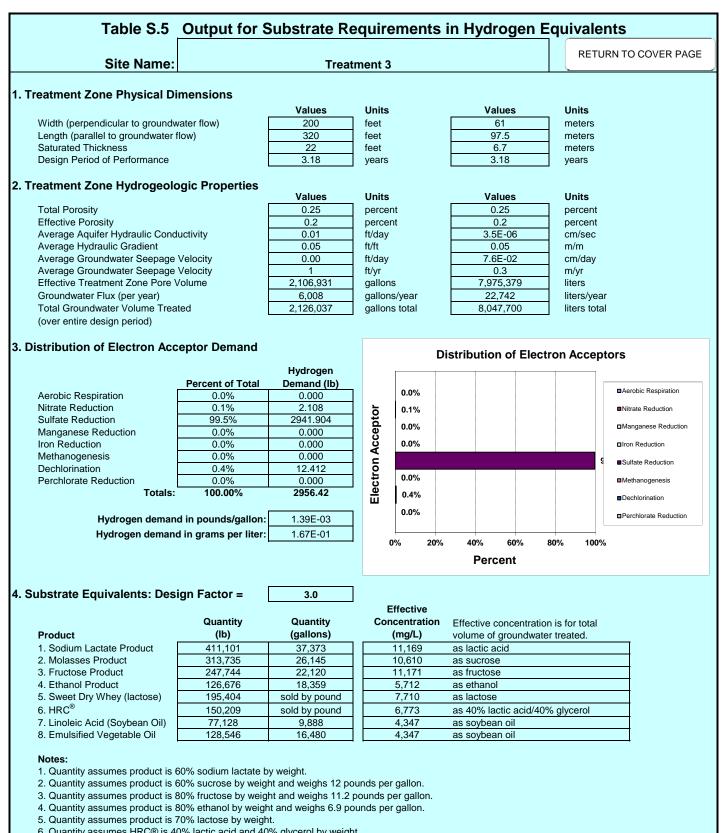
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 µs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

8. Quantity assumes commercial product is 60% soybean oil by weight.

### Treatment 3

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		37372.84436
2. Molasses Product		26144.59445
3. Fructose Product		22119.95706
4. Ethanol Product		18358.899
5. Sweet Dry Whey (lactose)	195403.7413	
6. HRC®	150209.4842	
7. Linoleic Acid (Soybean Oil)		9888.166579
8. Emulsified Vegetable Oil		16480.27763
9. Lactoil Product		7910.533263
10. Lactic Acid Product		0
11. Hydrogen Gas	2956.42	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:         Treatment 9         RETURN TO COVER PAGE           1. Treatment Zone Physical Dimensions         NOTE: Unshaded boxs are user input.         User Notes           Width (Perpendicular to predominant groundwater flow direction)         22         1-10.000         feet           Length (Parallel to predominant groundwater flow)         14         1-1,000         feet           Saturated Thickness         22.5         1-100         feet           Treatment Zone Cross Sectional Area         495          ft <sup>2</sup> Treatment Zone Volume         6,930          ft <sup>3</sup> Treatment Zone Volume (total volume x total proreity)         12,963          gallons           Treatment Zone Volume (total volume x telfective porosity)         10,370          gallons           Design Pactor (times the electron acceptor hydrogen demand)         3.0         2 to 20         unitiess         Default = 3           2. Treatment Zone Hydrogelogic Properties         Total Prorsity         25%         .05-50         percent         Default = 25%           Effective Porosity         20%         .05-50         percent         Default = 25%           Effective Porosity         20%         .05-50         percent         Default = 25%           Effecti	Table S.1 Input for Substrate Requirements in Hydrogen Equivalents					
1. Treatment Zone Physical Dimensions     Values     Range     Únits     User Notes       Width (Perpendicular to predominant groundwater flow direction)     22     1-10,000     feet       Length (Parallel to predominant groundwater flow)     14     1+1,000     feet       Saturated Thickness     22.5     1-100     feet       Treatment Zone Cross Sectional Area     445      ft <sup>2</sup> Treatment Zone Volume     6,930      ft <sup>2</sup> Treatment Zone Volume (total volume x total porosity)     12,963      gallons       Treatment Zone Effective Pore Volume (total volume x effective porosity)     10,370      gallons       Design Pactor (times the electron acceptor hydrogen demand)     3.0     2 to 20     unitless     Default = 3       2. Treatment Zone Hydrogeologic Properties      to 5.50     percent     Default = 25%       Effective Porosity     25%     .05-50     percent     Default = 20%       Average Aquiler Hydraulic Conductivity     4.8     .01-1000     ft/day       Average Groundwater Seepage Velocity through the Treatment Zone     0.0201-01     ft/ft       Average Groundwater Seepage Velocity through the Treatment Zone     32,443      ft/day       Average Groundwater Seepage Velocity through the Treatment Zone     0.05-60     0.01-10	Site Name:	Treatment 9			RETURN TO COVER PAGE	
1. Treatment Zone Physical Dimensions     Values     Range     Únits     User Notes       Width (Perpendicular to predominant groundwater flow direction)     22     1-10,000     feet       Length (Parallel to predominant groundwater flow)     14     1+1,000     feet       Saturated Thickness     22.5     1-100     feet       Treatment Zone Cross Sectional Area     445      ft <sup>2</sup> Treatment Zone Volume     6,930      ft <sup>2</sup> Treatment Zone Volume (total volume x total porosity)     12,963      gallons       Treatment Zone Effective Pore Volume (total volume x effective porosity)     10,370      gallons       Design Pactor (times the electron acceptor hydrogen demand)     3.0     2 to 20     unitless     Default = 3       2. Treatment Zone Hydrogeologic Properties      to 5.50     percent     Default = 25%       Effective Porosity     25%     .05-50     percent     Default = 20%       Average Aquiler Hydraulic Conductivity     4.8     .01-1000     ft/day       Average Groundwater Seepage Velocity through the Treatment Zone     0.0201-01     ft/ft       Average Groundwater Seepage Velocity through the Treatment Zone     32,443      ft/day       Average Groundwater Seepage Velocity through the Treatment Zone     0.05-60     0.01-10		NOTE: Unshaded	boxes are use	r input.		
Length (Parallel to predominant groundwater flow)         14         1-1,000         feet           Saturated Thickness         22.5         1.100         feet           Treatment Zone Cross Sectional Area         495          ft <sup>2</sup> Treatment Zone Volume         6,930          ft <sup>3</sup> Treatment Zone Total Pore Volume (total volume x total porosity)         12,963          gallons           Design Factor (times the electron acceptor hydrogen demand)         3.0         2 to 20         unitless         Default = 3           2. Treatment Zone Hydrogeologic Properties         Total Porosity         25%         .05-50         percent         Default = 3           2. Treatment Zone Hydrogeologic Properties         25%         .05-50         percent         Default = 25%           Effective Porosity         25%         .05-50         percent         Default = 25%           Effective Porosity         20%         .05-50         percent         Default = 25%           Effective Porosity         20%         .05-50         percent         Default = 25%           Effective Porosity         0.005         .0001-0.1         ft/ft         Average Aquifer Hydraulic Conductivity         4.8         .01-1000         ft/ft         Average Aquifer Hydraulic S				•	User Notes	
Saturated Thickness         22.5         1-100         feet           Treatment Zone Cross Sectional Area         495          ft <sup>2</sup> Treatment Zone Volume         6,930          ft <sup>3</sup> Treatment Zone Total Pore Volume (total volume x total porosity)         12,963          gallons           Treatment Zone Effective Pore Volume (total volume x effective porosity)         10,370          gallons           Design Period of Performance         1.3         .5 to 5         year           Design Pactor (times the electron acceptor hydrogen demand)         3.0         2 to 20         unitless <b>2. Treatment Zone Hydrogeologic Properties</b> Total Porosity         25%         .05-50         percent         Default = 25%           Effective Porosity         20%         .05-50         percent         Default = 20%           Average Aquifer Hydraulic Conductivity         4.8         .01-1000         ft/day           Average Groundwater Seepage Velocity through the Treatment Zone         0.12          ft/day           Average Groundwater Discharge through the Treatment Zone         32,443          gallons/year           Soil Buk Density         1.7         1.42.0         gm/cm <sup>3</sup> Default = 1.7 <td>Width (Perpendicular to predominant groundwater flow direction)</td> <td>22</td> <td>1-10,000</td> <td>feet</td> <td></td>	Width (Perpendicular to predominant groundwater flow direction)	22	1-10,000	feet		
Treatment Zone Cross Sectional Area       495        ft²         Treatment Zone Volume       6,930        ft³         Treatment Zone Total Pore Volume (total volume x total porosity)       12,963        gallons         Treatment Zone Effective Pore Volume (total volume x total porosity)       10,370        gallons         Design Pariod of Performance       1.3       .5 to 5       year         Design Factor (times the electron acceptor hydrogen demand)       3.0       2 to 20       unitless       Default = 3         2. Treatment Zone Hydrogeologic Properties       Total Porosity       25%       .05-50       percent       Default = 25%         Effective Porosity       20%       .05-50       percent       Default = 20%         Average Aquifer Hydraulic Conductivity       4.8       .01-1000       trday       Average Aquifer Hydraulic Conductivity       4.8       .01-1000       trday         Average Groundwater Seepage Velocity through the Treatment Zone       4.3.8        ft/ya         Average Groundwater Discharge through the Treatment Zone       4.3.8        ft/ya         Average Groundwater Discharge through the Treatment Zone       43.8        ft/ya         Soil Bulk Density       1.7       1.4-2.0       <	Length (Parallel to predominant groundwater flow)	14	1-1,000	feet		
Treatment Zone Volume       6,930        ft <sup>3</sup> Treatment Zone Total Pore Volume (total volume x total porosity)       12,963        gallons         Treatment Zone Effective Pore Volume (total volume x total porosity)       10,370        gallons         Design Period of Performance       1.3       .5 to 5       year         Design Factor (times the electron acceptor hydrogen demand)       3.0       2 to 20       unitless       Default = 3         2. Treatment Zone Hydrogeologic Properties       Total Porosity       25%       .05-50       percent       Default = 25%         Effective Porosity       20%       .05-50       percent       Default = 25%         Average Aquifer Hydraulic Conductivity       4.8       .01-1000       ft/day         Average Groundwater Seepage Velocity through the Treatment Zone       0.12        ft/day         Average Groundwater Seepage Velocity through the Treatment Zone       32,443        gallons/year         Soil Bulk Density       1.7       1.4-2.0       gm/cm <sup>3</sup> Default = 1.7         Soil Bulk Density       0.05%       0.01-10       percent       Default = 1.7         Soil Bulk Density       0.05%       0.01-10       percent       Default = 1.7         Soil Bul	Saturated Thickness	22.5	1-100			
Treatment Zone Total Pore Volume (total volume x total porosity)       12,963        gallons         Treatment Zone Effective Pore Volume (total volume x effective porosity)       10,370        gallons         Design Pariod of Performance       1.3       .5 to 5       year         Design Factor (times the electron acceptor hydrogen demand)       3.0       2 to 20       unitless       Default = 3         Contract Total Porosity         Effective Porosity       25%       .05-50       percent       Default = 25%         Effective Porosity       20%       .05-50       percent       Default = 25%         Average Aquifer Hydraulic Conductivity       4.8       .01-1000       ft/day         Average Groundwater Seepage Velocity through the Treatment Zone       0.12        ft/day         Average Groundwater Seepage Velocity through the Treatment Zone       32,443        gallons/year         Soil Bulk Density       1.7       1.4-2.0       gm/cm³       Default = 1.7         Soil Braction Organic Carbon (fcc)       0.05%       0.01-10       percent       Default = 1.7         Soil Burk Density       3.2       0.01 to 10       mg/L       Default = 5         Soil Braction Organic Carbon (fcc)       0.05%       0.01-10 <td< td=""><td>Treatment Zone Cross Sectional Area</td><td>495</td><td></td><td>ft<sup>2</sup></td><td></td></td<>	Treatment Zone Cross Sectional Area	495		ft <sup>2</sup>		
Treatment Zone Effective Pore Volume (total volume x effective porosity)       10.370        gallons         Design Period of Performance       1.3       5 to 5       year         Design Factor (times the electron acceptor hydrogen demand)       3.0       2 to 20       unitless       Default = 3         2. Treatment Zone Hydrogeologic Properties         Total Porosity       25%       .05-50       percent       Default = 25%         Effective Porosity       20%       .05-50       percent       Default = 20%         Average Aquifer Hydraulic Conductivity       4.8       .01-1000       ft/day         Average Groundwater Seepage Velocity through the Treatment Zone       0.12        ft/day         Average Groundwater Seepage Velocity through the Treatment Zone       32,443        gallons/year         Soil Bulk Density       1.7       1.4-2.0       gm/cm³       Default = 1.7         Soil Bulk Density       0.05%       0.01-10       percent       Default = 0.05%         3. Native Electron Acceptors       3.2       0.01 to 10       mg/L       Default = 5         Myrgen       3.2       0.01 to 10       mg/L       Default = 5         Sulfate       186       10 to 5,000       mg/L       Default = 50 <td>Treatment Zone Volume</td> <td>6,930</td> <td></td> <td>ft<sup>3</sup></td> <td></td>	Treatment Zone Volume	6,930		ft <sup>3</sup>		
Design Period of Performance       1.3       .5 to 5       year         Design Factor (times the electron acceptor hydrogen demand)       3.0       2 to 20       unitless       Default = 3         2. Treatment Zone Hydrogeologic Properties         Total Porosity       25%       .05-50       percent       Default = 25%         Effective Porosity       20%       .05-50       percent       Default = 20%         Average Aquifer Hydraulic Conductivity       4.8       .01-1000       ft/day         Average Aquifer Hydraulic Conductivity       4.8       .01-1000       ft/day         Average Groundwater Seepage Velocity through the Treatment Zone       0.12        ft/day         Average Groundwater Seepage Velocity through the Treatment Zone       43.8        ft/yr         Average Groundwater Discharge through the Treatment Zone       32.443        gallons/year         Soil Bulk Density       1.7       1.4-2.0       gm/cm³       Default = 1.7         Soil Fraction Organic Carbon (foc)       0.05%       0.01-10       percent       Default = 0.05%         3. Native Electron Acceptors       3.2       0.01 to 10       mg/L       Default = 5         Oxygen       3.2       0.01 to 10       mg/L       Default = 1 <tr< td=""><td>Treatment Zone Total Pore Volume (total volume x total porosity)</td><td>12,963</td><td></td><td>gallons</td><td></td></tr<>	Treatment Zone Total Pore Volume (total volume x total porosity)	12,963		gallons		
Design Factor (times the electron acceptor hydrogen demand)       3.0       2 to 20       unitless       Default = 3         2. Treatment Zone Hydrogeologic Properties         Total Porosity       25%       .05-50       percent       Default = 25%         Effective Porosity       20%       .05-50       percent       Default = 20%         Average Aquifer Hydraulic Conductivity       4.8       .01-1000       ft/day         Average Groundwater Seepage Velocity through the Treatment Zone       0.12        ft/day         Average Groundwater Seepage Velocity through the Treatment Zone       32,443        gallons/year         Soil Bulk Density       1.7       1.4-2.0       gm/cm³       Default = 1.7         Soil Fraction Organic Carbon (foc)       0.05%       0.01-10       percent       Default = 0.05%         3. Native Electron Acceptors       3.2       0.01 to 10       mg/L       Default = 5         Oxygen       3.2       0.01 to 10       mg/L       Default = 1         Sulfate       186       10 to 5,000       mg/L       Default = 1	Treatment Zone Effective Pore Volume (total volume x effective porosity)	10,370		gallons		
2. Treatment Zone Hydrogeologic Properties         Total Porosity       25%       .05-50       percent       Default = 25%         Effective Porosity       20%       .05-50       percent       Default = 20%         Average Aquifer Hydraulic Conductivity       4.8       .01-1000       ft/day         Average Hydraulic Gradient       0.005       0.0001-0.1       ft/ft         Average Groundwater Seepage Velocity through the Treatment Zone       0.12        ft/day         Average Groundwater Seepage Velocity through the Treatment Zone       32,443        gallons/year         Soil Bulk Density       1.7       1.4-2.0       gm/cm³       Default = 1.7         Soil Bulk Density       0.05%       0.01-10       percent       Default = 1.7         Soil Fraction Organic Carbon (foc)       0.05%       0.01-10       percent       Default = 0.05%         3. Native Electron Acceptors       A. Aqueous-Phase Native Electron Acceptors       O.000       0.1 to 20       mg/L       Default = 5         Nitrate       0.00       0.10 to 20       mg/L       Default = 1       50	Design Period of Performance	1.3	.5 to 5	year		
Total Porosity         25%         .05-50         percent         Default = 25%           Effective Porosity         20%         .05-50         percent         Default = 20%           Average Aquifer Hydraulic Conductivity         4.8         .01-1000         ft/day           Average Hydraulic Gradient         0.005         0.0001-0.1         ft/ft           Average Groundwater Seepage Velocity through the Treatment Zone         0.12          ft/day           Average Groundwater Seepage Velocity through the Treatment Zone         43.8          ft/yr           Average Groundwater Discharge through the Treatment Zone         32,443          gallons/year           Soil Bulk Density         1.7         1.4-2.0         gm/cm <sup>3</sup> Default = 1.7           Soil Fraction Organic Carbon (foc)         0.05%         0.01-10         percent         Default = 0.05%           3. Native Electron Acceptors	Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
Effective Porosity20%.05-50percentDefault = 20%Average Aquifer Hydraulic Conductivity4.8.01-1000ft/dayAverage Hydraulic Gradient0.0050.0001-0.1ft/ftAverage Groundwater Seepage Velocity through the Treatment Zone0.12ft/dayAverage Groundwater Seepage Velocity through the Treatment Zone43.8ft/yrAverage Groundwater Discharge through the Treatment Zone32,443gallons/yearSoil Bulk Density1.71.4-2.0gm/cm³Default = 1.7Soil Fraction Organic Carbon (foc)0.05%0.01-10percentDefault = 0.05%3. Native Electron Acceptors3.20.01 to 10mg/LDefault = 5Oxygen3.20.000.1 to -20mg/LDefault = 1Nitrate0.001.1 to 5,000mg/LDefault = 1						
Average Aquifer Hydraulic Conductivity         4.8         .01-1000         ft/day           Average Hydraulic Gradient         0.005         0.0001-0.1         ft/day           Average Groundwater Seepage Velocity through the Treatment Zone         0.12          ft/day           Average Groundwater Seepage Velocity through the Treatment Zone         0.12          ft/day           Average Groundwater Seepage Velocity through the Treatment Zone         43.8          ft/yr           Average Groundwater Discharge through the Treatment Zone         32,443          gallons/year           Soil Bulk Density         1.7         1.4-2.0         gm/cm³         Default = 1.7           Soil Fraction Organic Carbon (foc)         0.05%         0.01-10         percent         Default = 0.05%           3. Native Electron Acceptors				percent		
Average Hydraulic Gradient0.0050.0001-0.1ft/ftAverage Groundwater Seepage Velocity through the Treatment Zone0.12ft/dayAverage Groundwater Seepage Velocity through the Treatment Zone43.8ft/yrAverage Groundwater Discharge through the Treatment Zone32,443gallons/yearSoil Bulk Density1.71.4-2.0gm/cm³Default = 1.7Soil Fraction Organic Carbon (foc)0.05%0.01-10percentDefault = 0.05%Oxygen3.20.01 to 10mg/LDefault = 5Oxygen3.20.01 to 10mg/LDefault = 1Nitrate0.000.1 to -20mg/LDefault = 1Sulfate18610 to 5,000mg/LDefault = 50					Default = 20%	
Average Groundwater Seepage Velocity through the Treatment Zone       0.12        ft/day         Average Groundwater Seepage Velocity through the Treatment Zone       43.8        ft/yr         Average Groundwater Discharge through the Treatment Zone       32,443        gallons/year         Soil Bulk Density       1.7       1.4-2.0       gm/cm <sup>3</sup> Default = 1.7         Soil Fraction Organic Carbon (foc)       0.05%       0.01-10       percent       Default = 0.05%         3. Native Electron Acceptors		4.8	.01-1000	ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone       43.8        ft/yr         Average Groundwater Discharge through the Treatment Zone       32,443        gallons/year         Soil Bulk Density       1.7       1.4-2.0       gm/cm <sup>3</sup> Default = 1.7         Soil Fraction Organic Carbon (foc)       0.05%       0.01-10       percent       Default = 0.05%         Aqueous-Phase Native Electron Acceptors       Aqueous-Phase Native Electron Acceptors       Acceptors       Soil Tract       0.01 to 10       mg/L       Default = 5         Nitrate       0.00       0.1 to 20       mg/L       Default = 1       Default = 1         Sulfate       186       10 to 5,000       mg/L       Default = 50       Default = 50			0.0001-0.1			
Average Groundwater Discharge through the Treatment Zone       32,443        gallons/year         Soil Bulk Density       1.7       1.4-2.0       gm/cm <sup>3</sup> Default = 1.7         Soil Fraction Organic Carbon (foc)       0.05%       0.01-10       percent       Default = 0.05%         3. Native Electron Acceptors						
Soil Bulk Density       1.7       1.4-2.0       gm/cm <sup>3</sup> Default = 1.7         Soil Fraction Organic Carbon (foc)       0.05%       0.01-10       percent       Default = 0.05%         3. Native Electron Acceptors         A. Aqueous-Phase Native Electron Acceptors         Oxygen       3.2       0.01 to 10       mg/L       Default = 5         Nitrate       0.00       0.1 to - 20       mg/L       Default = 1         Sulfate       186       10 to 5,000       mg/L       Default = 50						
Soil Fraction Organic Carbon (foc)       0.05%       0.01-10       percent       Default = 0.05%         3. Native Electron Acceptors       A. Aqueous-Phase Native Electron Acceptors       0.01 to 10       mg/L       Default = 5         Oxygen       3.2       0.01 to 10       mg/L       Default = 5         Nitrate       0.00       0.1 to - 20       mg/L       Default = 1         Sulfate       186       10 to 5,000       mg/L       Default = 50		,				
3. Native Electron Acceptors         A. Aqueous-Phase Native Electron Acceptors         Oxygen       3.2       0.01 to 10       mg/L       Default = 5         Nitrate       0.00       0.1 to- 20       mg/L       Default = 1         Sulfate       186       10 to 5,000       mg/L       Default = 50				gm/cm³		
A. Aqueous-Phase Native Electron Acceptors           Oxygen         3.2         0.01 to 10         mg/L         Default = 5           Nitrate         0.00         0.1 to 20         mg/L         Default = 1           Sulfate         186         10 to 5,000         mg/L         Default = 50	Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
Nitrate         0.00         0.1 to- 20         mg/L         Default = 1           Sulfate         186         10 to 5,000         mg/L         Default = 50	·					
Sulfate         186         10 to 5,000 mg/L         Default = 50		-		0		
				0		
Carbon Dioxide (estimated as the amount of Methane produced)     0.0     0.1 to 20     mg/L     Default = 0	Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0	
B. Solid-Phase Native Electron Acceptors	B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)     0.1 to 20 mg/L     Default = 0	Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L		
Iron (III) (estimated as the amount of Fe (II) produced) 0.1 to 20 mg/L Default = 0	Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

### 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.001	mg/L	
Trichloroethene (TCE)	18.394	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	21.345	mg/L	
Vinyl Chloride (VC)	7.627	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)	0.001	mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

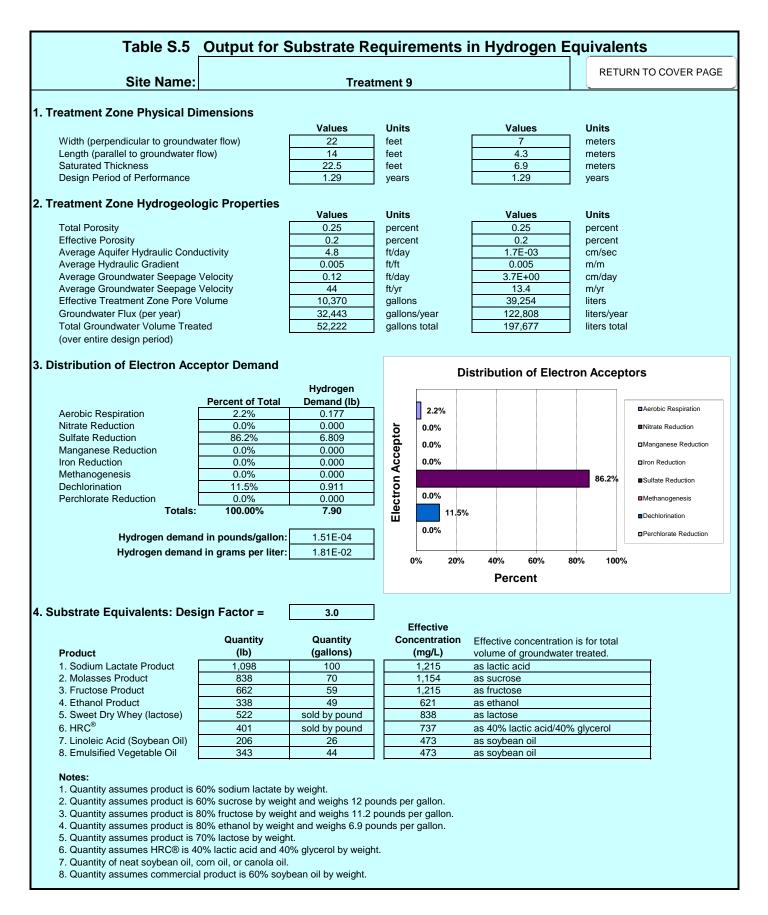
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



### Treatment 9

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		99.82835101
2. Molasses Product		69.83604797
3. Fructose Product		59.08565096
4. Ethanol Product		49.03931304
5. Sweet Dry Whey (lactose)	521.9520646	
6. HRC®	401.2315725	
7. Linoleic Acid (Soybean Oil)		26.41274382
8. Emulsified Vegetable Oil		44.02123971
9. Lactoil Product		21.13019506
10. Lactic Acid Product		0
11. Hydrogen Gas	7.90	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 10			RETURN TO COVER PAGE
	NOTE: Unshaded	boxes are use	r input.	
. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes
Width (Perpendicular to predominant groundwater flow direction)	22	1-10,000	feet	
Length (Parallel to predominant groundwater flow)	14	1-1,000	feet	
Saturated Thickness	22.5	1-100	feet	
Treatment Zone Cross Sectional Area	495		ft <sup>2</sup>	
Treatment Zone Volume	6,930		ft <sup>3</sup>	
Treatment Zone Total Pore Volume (total volume x total porosity)	12,963		gallons	
Treatment Zone Effective Pore Volume (total volume x effective porosity)	10,370		gallons	
Design Period of Performance	0.3	.5 to 5	year	
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3
Treatment Zone Hydrogeologic Properties	25%	.05-50	percent	Default = 25%
Effective Porosity	20%	.05-50	percent	Default = 20%
Average Aquifer Hydraulic Conductivity	4.8	.03-30	ft/day	
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft	
Average Groundwater Seepage Velocity through the Treatment Zone	0.12		ft/day	
Average Groundwater Seepage Velocity through the Treatment Zone	43.8		ft/yr	
Average Groundwater Discharge through the Treatment Zone	32,443		gallons/year	
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%
. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors	5.0	0.01 += 10		Default = 5
Oxygen Nitrate	0.00	0.01 to 10 0.1 to- 20	mg/L	Default = 5 Default = 1
Sulfate	273		mg/L	Default = 1 Default = 50
		10 to 5,000	mg/L	
Carbon Dioxide (estimated as the amount of Methane produced) B. Solid-Phase Native Electron Acceptors	0.0	0.1 to 20	mg/L	Default = 0
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0

# 4. Contaminant Electron Acceptors

0.000	mg/L	
0.024	mg/L	
0.404	mg/L	
0.204	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.000	mg/L	
0.000	mg/L	
	mg/L	
	mg/L	
	0.024 0.404 0.204	0.024          mg/L           0.404          mg/L           0.204          mg/L           0.204          mg/L            mg/L             mg/L             mg/L             mg/L            0.000          mg/L           0.000          mg/L           0.000          mg/L           0.000          mg/L

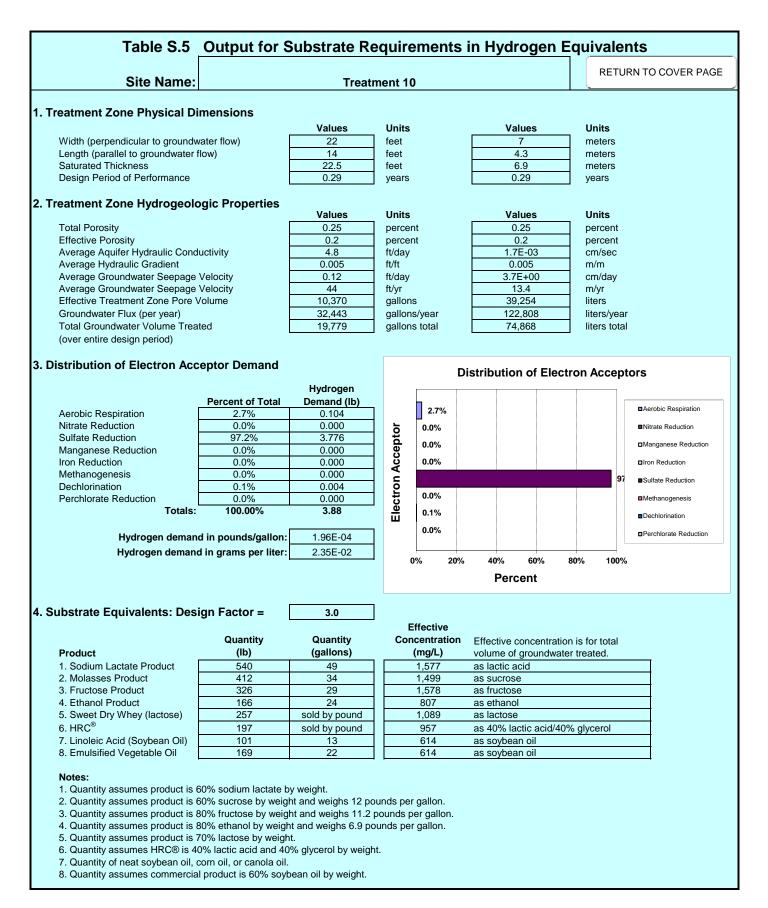
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



# Treatment 10

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		49.10475666
2. Molasses Product		34.35178591
3. Fructose Product		29.06375276
4. Ethanol Product		24.12204057
5. Sweet Dry Whey (lactose)	256.7439897	7
6. HRC®	197.3625581	
7. Linoleic Acid (Soybean Oil)		12.99221459
8. Emulsified Vegetable Oil		21.65369098
9. Lactoil Product		10.39377167
10. Lactic Acid Product		0
11. Hydrogen Gas	3.88	3

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents						
Site Name:	Treatment 11			RETURN TO COVER PAGE		
	NOTE: Unshaded	boxes are use	r input.			
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	170	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	20	1-1,000	feet			
Saturated Thickness	20.5	1-100	feet			
Treatment Zone Cross Sectional Area	3485		ft <sup>2</sup>			
Treatment Zone Volume	69,700		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	130,374		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	104,299		gallons			
Design Period of Performance	3.8	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
2. Treatment Zone Hydrogeologic Properties						
Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	11	.01-1000	ft/day			
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr			
Average Groundwater Discharge through the Treatment Zone	523,451		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors						
Oxygen	5.0	0.01 to 10	mg/L	Default = 5		
Nitrate	0.00	0.1 to- 20	mg/L	Default = 1		
Sulfate	3	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors						
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

# 4. Contaminant Electron Acceptors

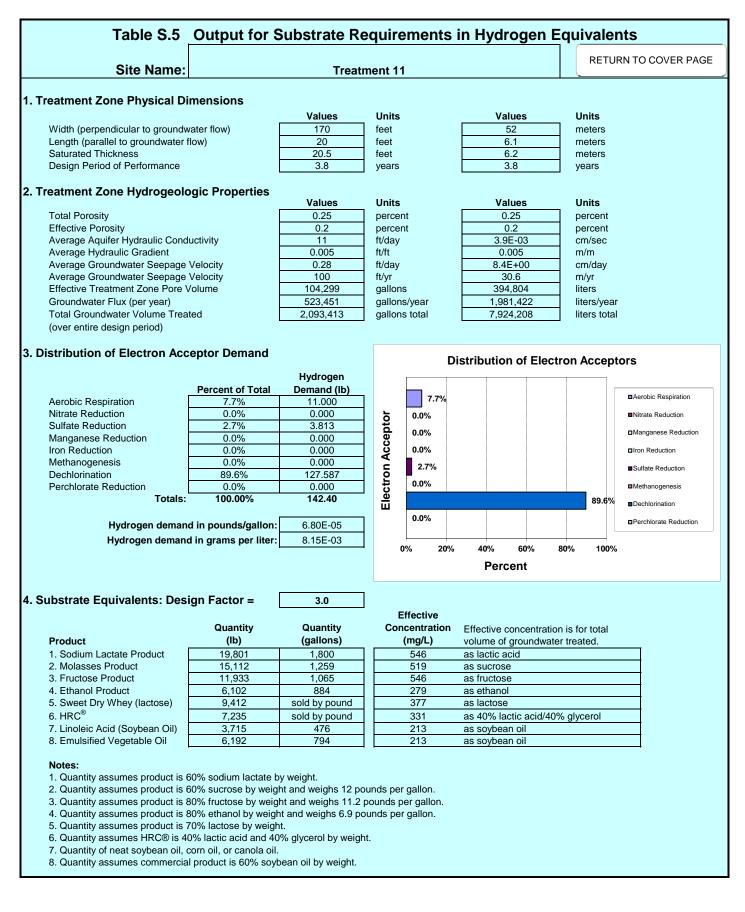
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

#### **B. Aquifer Matrix**

Total Iron	10000	200 to 20,000 mg/kg Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g
Neutralization Potential	10.0%	1.0 to 100 Percent as $CaCO_3$ Default = 10%



# Treatment 11

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		1800.120235
2. Molasses Product		1259.294397
3. Fructose Product		1065.441578
4. Ethanol Product		884.2844624
5. Sweet Dry Whey (lactose)	9411.920197	,
6. HRC®	7235.069649	)
7. Linoleic Acid (Soybean Oil)		476.2786737
8. Emulsified Vegetable Oil		793.7977895
9. Lactoil Product		381.0229389
10. Lactic Acid Product		0
11. Hydrogen Gas	142.40	)

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 12			RETURN TO COVER PAGE		
NOTE: Unshaded boxes are user input.						
. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	170	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	30	1-1,000	feet			
Saturated Thickness	24	1-100	feet			
Treatment Zone Cross Sectional Area	4080		ft <sup>2</sup>			
Treatment Zone Volume	122,400		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	228,949		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	183,159		gallons			
Design Period of Performance	0.9	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
Treatment Zone Hydrogeologic Properties     Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	11	.03-30	ft/day			
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr			
Average Groundwater Discharge through the Treatment Zone	612,821		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
Native Electron Acceptors     A. Aqueous-Phase Native Electron Acceptors	5.0	0.01 to 10		Default = 5		
Oxygen Nitrate	0.00	0.01 to 10	mg/L	Default = 5		
Sulfate	12	10 to 5,000	mg/L mg/L	Default = 1 Default = 50		
	0.0	0.1 to 20				
Carbon Dioxide (estimated as the amount of Methane produced) B. Solid-Phase Native Electron Acceptors	0.0	0.1 to 20	mg/L	Default = 0		
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

### 4. Contaminant Electron Acceptors

0.001	mg/L	
21.560	mg/L	
6.109	mg/L	
0.000	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.000	mg/L	
0.000	mg/L	
	mg/L	
	mg/L	
	21.560 6.109 0.000	21.560          mg/L           6.109          mg/L           0.000          mg/L            mg/L             mg/L             mg/L             mg/L             mg/L            0.000          mg/L           0.000          mg/L           0.000          mg/L           0.000          mg/L           0.000          mg/L

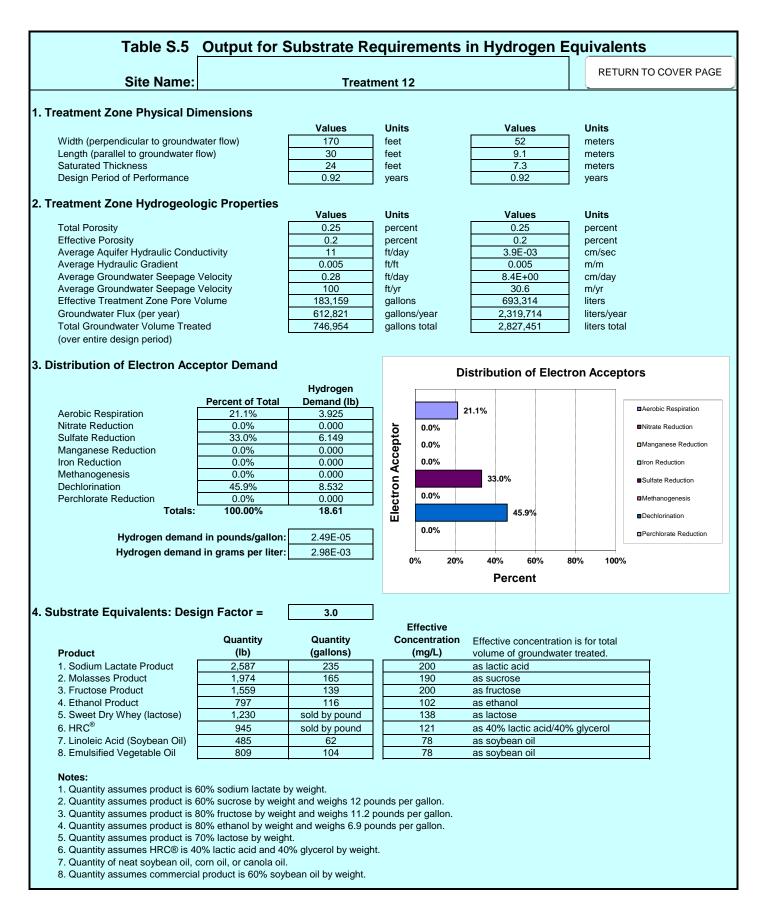
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



## Treatment 12

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		235.2040596
2. Molasses Product		164.5396506
3. Fructose Product		139.2108037
4. Ethanol Product		115.5407797
5. Sweet Dry Whey (lactose)	1229.763322	2
6. HRC®	945.3356061	
7. Linoleic Acid (Soybean Oil)		62.23066404
8. Emulsified Vegetable Oil		103.7177734
9. Lactoil Product		49.78453124
10. Lactic Acid Product		0
11. Hydrogen Gas	18.61	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 23			RETURN TO COVER PAGE	
Site Name.		• .			
NOTE: Unshaded boxes are user input.					
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes	
Width (Perpendicular to predominant groundwater flow direction)	50	1-10,000	feet		
Length (Parallel to predominant groundwater flow)	50	1-1,000	feet		
Saturated Thickness	24	1-100	feet		
Treatment Zone Cross Sectional Area	1200		ft <sup>2</sup>		
Treatment Zone Volume	60,000		ft <sup>3</sup>		
Treatment Zone Total Pore Volume (total volume x total porosity)	112,230		gallons		
Treatment Zone Effective Pore Volume (total volume x effective porosity)	89,784		gallons		
Design Period of Performance	1.8	.5 to 5	year		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
2. Treatment Zone Hydrogeologic Properties					
Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity	20%	.05-50	percent	Default = 20%	
Average Aquifer Hydraulic Conductivity	11	.01-1000	ft/day		
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr		
Average Groundwater Discharge through the Treatment Zone	180,241		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors	r				
Oxygen	5.0	0.01 to 10	mg/L	Default = 5	
Nitrate	2.40	0.1 to- 20	mg/L	Default = 1	
Sulfate	2	10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0	
B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

### 4. Contaminant Electron Acceptors

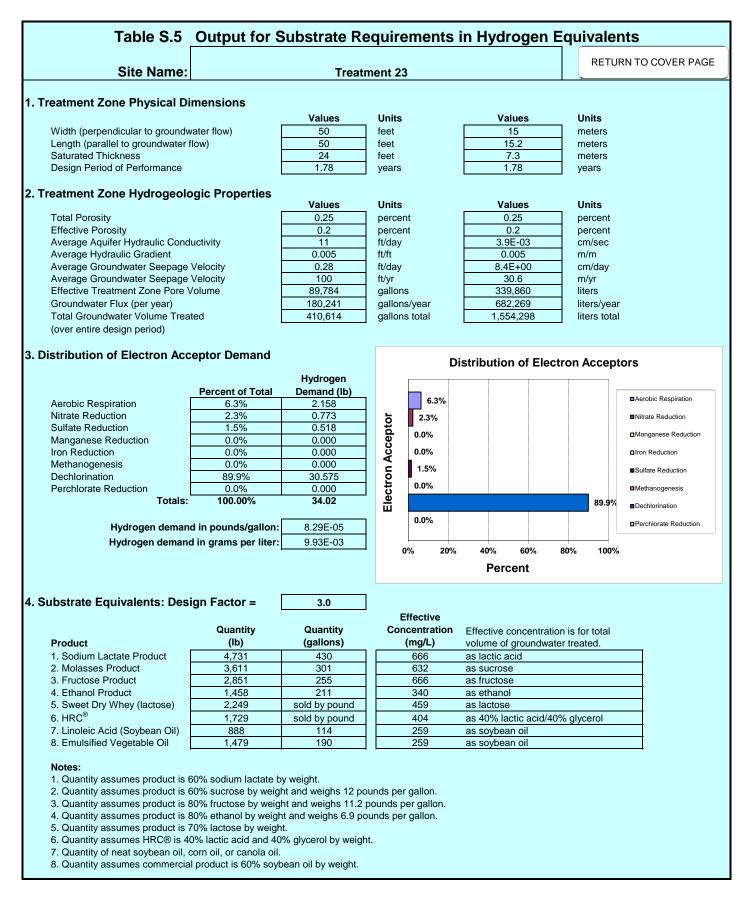
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



### Treatment 23

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		430.0997841
2. Molasses Product		300.881151
3. Fructose Product		254.5642141
4. Ethanol Product		211.2806406
5. Sweet Dry Whey (lactose)	2248.774702	
6. HRC®	1728.66336	
7. Linoleic Acid (Soybean Oil)		113.7964847
8. Emulsified Vegetable Oil		189.6608078
9. Lactoil Product		91.03718775
10. Lactic Acid Product		0
11. Hydrogen Gas	34.02	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

<b>•</b> •••••						
Site Name:	Treatment 24			RETURN TO COVER PAGE		
NOTE: Unshaded boxes are user input.						
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	220	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	140	1-1,000	feet			
Saturated Thickness	24	1-100	feet			
Treatment Zone Cross Sectional Area	5280		ft <sup>2</sup>			
Treatment Zone Volume	739,200		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	1,382,674		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	1,106,139		gallons			
Design Period of Performance	2.6	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
2. Treatment Zone Hydrogeologic Properties						
Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	11	.01-1000	ft/day			
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr			
Average Groundwater Discharge through the Treatment Zone	793,062		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors						
Oxygen	5.0	0.01 to 10	mg/L	Default = 5		
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1		
Sulfate	50	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors						
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

### 4. Contaminant Electron Acceptors

0.000	mg/L	
83.000	mg/L	
80.580	mg/L	
0.000	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.000	mg/L	
0.000	mg/L	
	mg/L	
	mg/L	
	83.000 80.580 0.000 0.000	83.000          mg/L           80.580          mg/L           0.000          mg/L            mg/L             mg/L             mg/L             mg/L             mg/L            0.000          mg/L           0.000          mg/L           0.000          mg/L           0.000          mg/L           0.000          mg/L

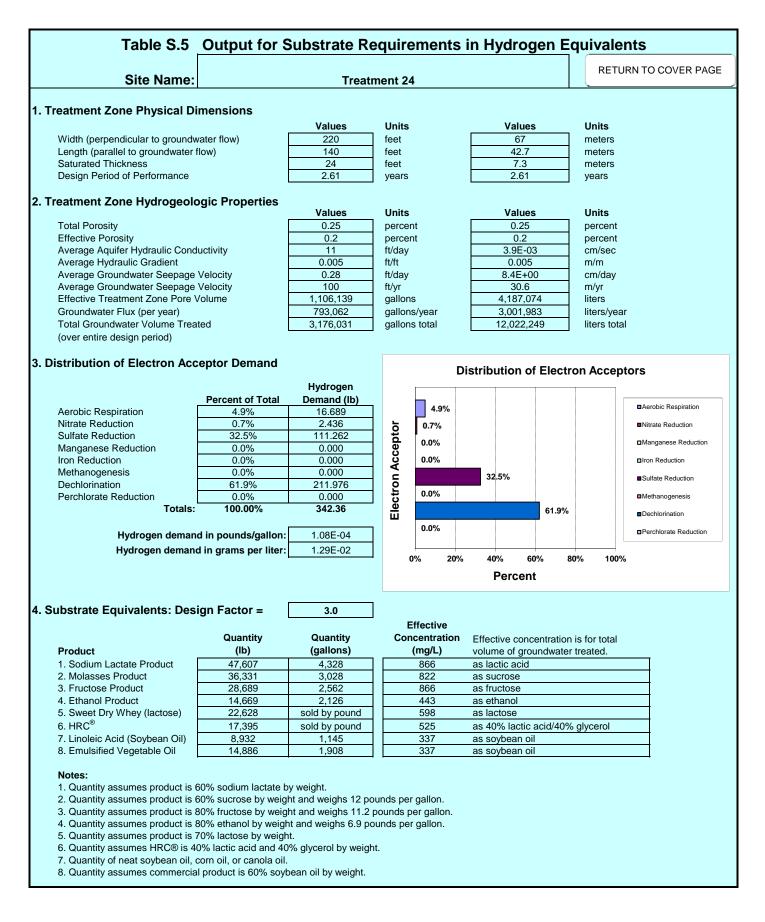
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



### Treatment 24

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		4327.878067
2. Molasses Product		3027.615876
3. Fructose Product		2561.551807
4. Ethanol Product		2126.01095
5. Sweet Dry Whey (lactose)	22628.29016	
6. HRC®	17394.671	
7. Linoleic Acid (Soybean Oil)		1145.07686
8. Emulsified Vegetable Oil		1908.461433
9. Lactoil Product		916.0614879
10. Lactic Acid Product		0
11. Hydrogen Gas	342.36	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents					
Site Name:	Treatment 25			RETURN TO COVER PAGE	
	NOTE: Unshaded	boxes are use	r input.		
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes	
Width (Perpendicular to predominant groundwater flow direction)	170	1-10,000	feet		
Length (Parallel to predominant groundwater flow)	20	1-1,000	feet		
Saturated Thickness	24	1-100	feet		
Treatment Zone Cross Sectional Area	4080		ft <sup>2</sup>		
Treatment Zone Volume	81,600		ft <sup>3</sup>		
Treatment Zone Total Pore Volume (total volume x total porosity)	152,633		gallons		
Treatment Zone Effective Pore Volume (total volume x effective porosity)	122,106		gallons		
Design Period of Performance	2.6	.5 to 5	year		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
2. Treatment Zone Hydrogeologic Properties	· · · · · · · · · · · · · · · · · · ·				
Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity	20%	.05-50	percent	Default = 20%	
Average Aquifer Hydraulic Conductivity	11	.01-1000	ft/day		
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr		
Average Groundwater Discharge through the Treatment Zone	612,821		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors					
Oxygen	5.0	0.01 to 10	mg/L	Default = 5	
Nitrate	2.40	0.1 to- 20	mg/L	Default = 1	
Sulfate	2	10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0	
B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

### 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L	
Trichloroethene (TCE)	14.000	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	170.820	mg/L	
Vinyl Chloride (VC)	0.000	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)	5.200	mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)	0.000	mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA) Trichloroethane (1,1,1-TCA and 1,1,2-TCA) Dichloroethane (1,1-DCA and 1,2-DCA) Chloroethane		mg/L mg/L mg/L mg/L mg/L	

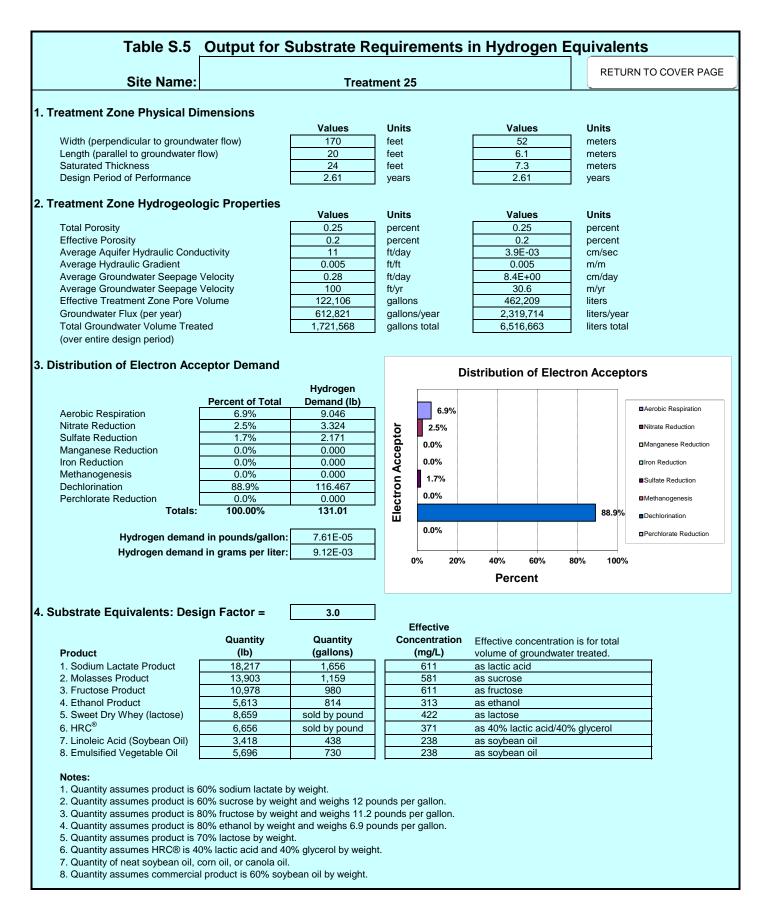
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



## Treatment 25

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		1656.107331
2. Molasses Product		1158.548547
3. Fructose Product		980.2043078
4. Ethanol Product		813.5400916
5. Sweet Dry Whey (lactose)	8658.949406	;
6. HRC®	6656.250874	
7. Linoleic Acid (Soybean Oil)		438.1755106
8. Emulsified Vegetable Oil		730.2925177
9. Lactoil Product		350.5404085
10. Lactic Acid Product		0
11. Hydrogen Gas	131.01	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 26			RETURN TO COVER PAGE		
NOTE: Unshaded boxes are user input.						
. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	170	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	20	1-1,000	feet			
Saturated Thickness	24	1-100	feet			
Treatment Zone Cross Sectional Area	4080		ft <sup>2</sup>			
Treatment Zone Volume	81,600		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	152,633		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	122,106		gallons			
Design Period of Performance	2.6	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
. Treatment Zone Hydrogeologic Properties Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	11	.01-1000	ft/day			
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr			
Average Groundwater Discharge through the Treatment Zone	612,821		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
Native Electron Acceptors     A. Aqueous-Phase Native Electron Acceptors     Overgen	5.0	0.01 to 10	ma/l	Default = 5		
Oxygen Nitrate	0.50	0.01 to 10	mg/L mg/L	Default = 5		
Sulfate	6	10 to 5,000	mg/L	Default = 1 Default = 50		
	0.0	0.1 to 20	v			
Carbon Dioxide (estimated as the amount of Methane produced) B. Solid-Phase Native Electron Acceptors	0.0	0.1 to 20	mg/L	Default = 0		
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

### 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L	
Trichloroethene (TCE)	0.037	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	32.196	mg/L	
Vinyl Chloride (VC)	10.001	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)	0.000	mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)	0.001	mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

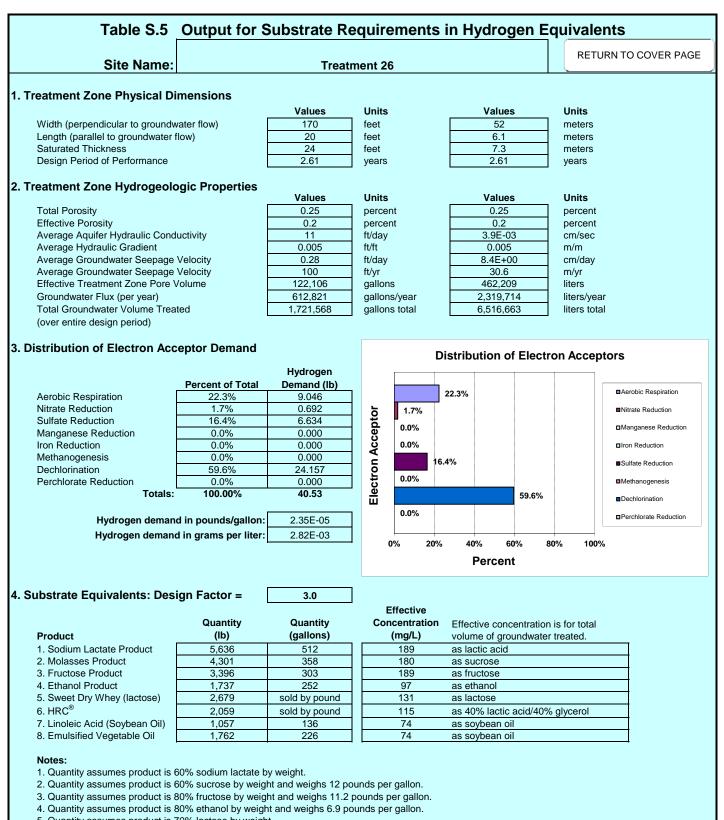
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 µs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes  ${\rm HRC} \ensuremath{\mathbb{R}}$  is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.

### Treatment 26

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		512.3481708
2. Molasses Product		358.4189369
3. Fructose Product		303.2447685
4. Ethanol Product		251.6840365
5. Sweet Dry Whey (lactose)	2678.810006	;
6. HRC®	2059.237282	2
7. Linoleic Acid (Soybean Oil)		135.5578936
8. Emulsified Vegetable Oil		225.9298227
9. Lactoil Product		108.4463149
10. Lactic Acid Product		0
11. Hydrogen Gas	40.53	}

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents							
Site Name:	Treatment 62			RETURN TO COVER PAGE			
NOTE: Unshaded boxes are user input.							
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes			
Width (Perpendicular to predominant groundwater flow direction)	54	1-10,000	feet				
Length (Parallel to predominant groundwater flow)	34	1-1,000	feet				
Saturated Thickness	24	1-100	feet				
Treatment Zone Cross Sectional Area	1296		ft <sup>2</sup>				
Treatment Zone Volume	44,064		ft <sup>3</sup>				
Treatment Zone Total Pore Volume (total volume x total porosity)	82,422		gallons				
Treatment Zone Effective Pore Volume (total volume x effective porosity)	65,937		gallons				
Design Period of Performance	2.5	.5 to 5	year				
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3			
2. Treatment Zone Hydrogeologic Properties							
Total Porosity	25%	.05-50	percent	Default = 25%			
Effective Porosity	20%	.05-50	percent	Default = 20%			
Average Aquifer Hydraulic Conductivity	11	.01-1000	ft/day				
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft				
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day				
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr				
Average Groundwater Discharge through the Treatment Zone	194,661		gallons/year				
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7			
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%			
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors							
Oxygen	5.0	0.01 to 10	mg/L	Default = 5			
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1			
Sulfate	50	10 to 5,000	mg/L	Default = 50			
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0			
B. Solid-Phase Native Electron Acceptors							
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0			
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0			

### 4. Contaminant Electron Acceptors

0.000	mg/L	
1.928	mg/L	
3.535	mg/L	
0.029	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.000	mg/L	
0.000	mg/L	
	mg/L	
	mg/L	
	1.928 3.535 0.029 0.000	1.928      mg/L       3.535      mg/L       0.029      mg/L        mg/L        mg/L        mg/L        mg/L        mg/L        mg/L       0.000        0.000         mg/L       0.000         mg/L       0.000         mg/L       0.000

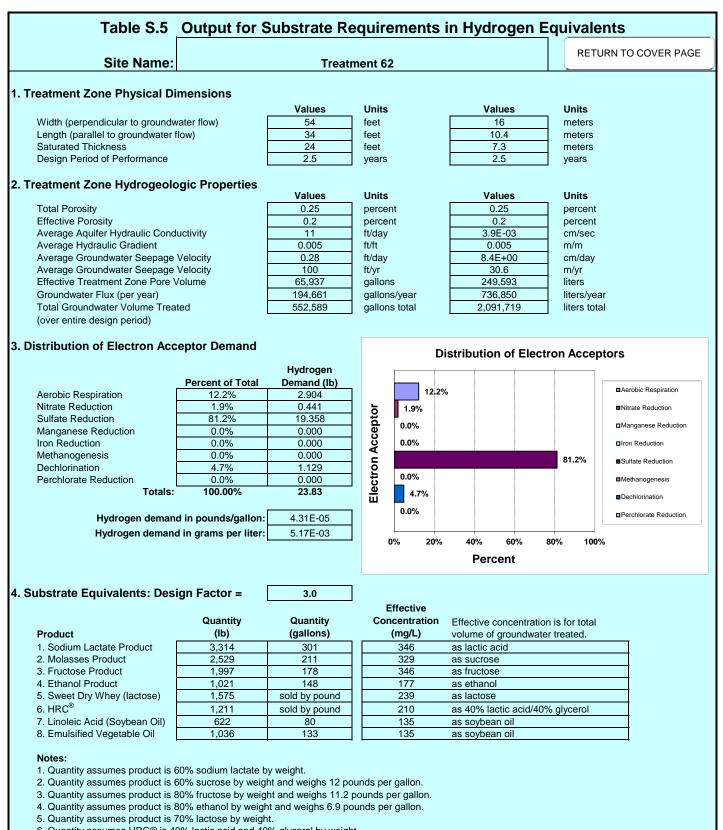
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

8. Quantity assumes commercial product is 60% soybean oil by weight.

## Treatment 62

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		301.2610426
2. Molasses Product		210.7505575
3. Fructose Product		178.3081122
4. Ethanol Product		147.9903697
5. Sweet Dry Whey (lactose)	1575.141947	7
6. HRC®	1210.832801	
7. Linoleic Acid (Soybean Oil)		79.70812562
8. Emulsified Vegetable Oil		132.846876
9. Lactoil Product		63.7665005
10. Lactic Acid Product		0
11. Hydrogen Gas	23.83	3

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents							
Site Name:	Treatment 71			RETURN TO COVER PAGE			
NOTE: Unshaded boxes are user input.							
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes			
Width (Perpendicular to predominant groundwater flow direction)	50	1-10,000	feet				
Length (Parallel to predominant groundwater flow)	50	1-1,000	feet				
Saturated Thickness	24	1-100	feet				
Treatment Zone Cross Sectional Area	1200		ft <sup>2</sup>				
Treatment Zone Volume	60,000		ft <sup>3</sup>				
Treatment Zone Total Pore Volume (total volume x total porosity)	112,230		gallons				
Treatment Zone Effective Pore Volume (total volume x effective porosity)	89,784		gallons				
Design Period of Performance	1.8	.5 to 5	year				
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3			
2. Treatment Zone Hydrogeologic Properties							
Total Porosity	25%	.05-50	percent	Default = 25%			
Effective Porosity	20%	.05-50	percent	Default = 20%			
Average Aquifer Hydraulic Conductivity	11	.01-1000	ft/day				
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft				
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day				
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr				
Average Groundwater Discharge through the Treatment Zone	180,241		gallons/year				
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7			
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%			
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors							
Oxygen	0.6	0.01 to 10	mg/L	Default = 5			
Nitrate	0.00	0.1 to- 20	mg/L	Default = 1			
Sulfate	0	10 to 5,000	mg/L	Default = 50			
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0			
B. Solid-Phase Native Electron Acceptors							
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0			
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0			

### 4. Contaminant Electron Acceptors

0.143	mg/L	
143.467	mg/L	
35.286	mg/L	
1.965	mg/L	
	mg/L	
0.000	mg/L	
0.002	mg/L	
	mg/L	
	mg/L	
	143.467 35.286 1.965 0.000	143.467          mg/L           35.286          mg/L           1.965          mg/L            mg/L             mg/L             mg/L             mg/L             mg/L            0.000          mg/L           0.002          mg/L            mg/L

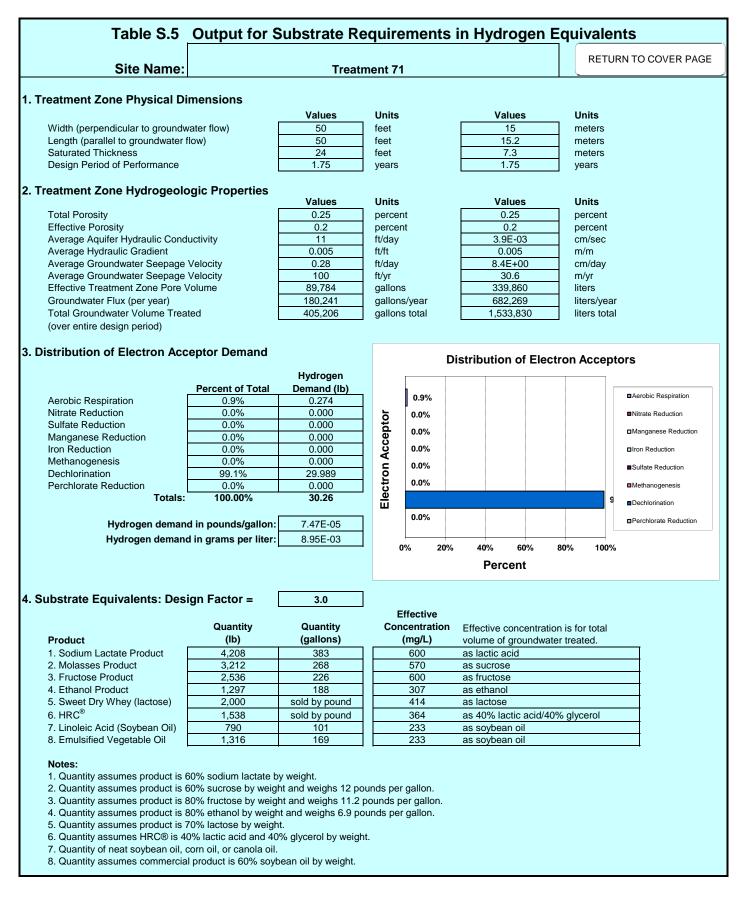
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



## Treatment 71

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		382.5641124
2. Molasses Product		267.6270362
3. Fructose Product		226.4291595
4. Ethanol Product		187.9293916
5. Sweet Dry Whey (lactose)	2000.234666	;
6. HRC®	1537.607292	2
7. Linoleic Acid (Soybean Oil)		101.2194211
8. Emulsified Vegetable Oil		168.6990352
9. Lactoil Product		80.97553688
10. Lactic Acid Product		0
11. Hydrogen Gas	30.26	;

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents							
Site Name:	Treatment 72			RETURN TO COVER PAGE			
	NOTE: Unshaded boxes are user input.						
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes			
Width (Perpendicular to predominant groundwater flow direction)	220	1-10,000	feet				
Length (Parallel to predominant groundwater flow)	140	1-1,000	feet				
Saturated Thickness	24	1-100	feet				
Treatment Zone Cross Sectional Area	5280		ft <sup>2</sup>				
Treatment Zone Volume	739,200		ft <sup>3</sup>				
Treatment Zone Total Pore Volume (total volume x total porosity)	1,382,674		gallons				
Treatment Zone Effective Pore Volume (total volume x effective porosity)	1,106,139		gallons				
Design Period of Performance	2.0	.5 to 5	year				
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3			
2. Treatment Zone Hydrogeologic Properties							
Total Porosity	25%	.05-50	percent	Default = 25%			
Effective Porosity	20%	.05-50	percent	Default = 20%			
Average Aquifer Hydraulic Conductivity	11	.01-1000	ft/day				
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft				
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day				
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr				
Average Groundwater Discharge through the Treatment Zone	793,062		gallons/year				
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7			
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%			
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors							
Oxygen	5.0	0.01 to 10	mg/L	Default = 5			
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1			
Sulfate	50	10 to 5,000	mg/L	Default = 50			
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0			
B. Solid-Phase Native Electron Acceptors							
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0			
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0			

### 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L	
Trichloroethene (TCE)	0.169	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	49.818	mg/L	
Vinyl Chloride (VC)	48.800	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)	0.000	mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)	0.000	mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

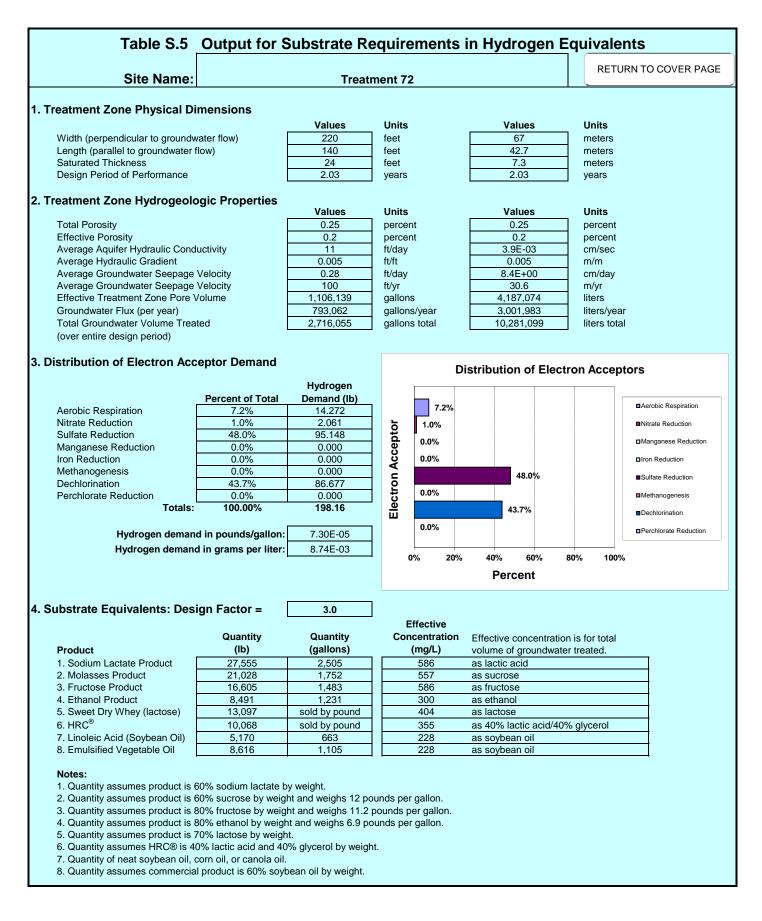
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



# Treatment 72

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		2504.957422
2. Molasses Product		1752.371194
3. Fructose Product		1482.615293
4. Ethanol Product		1230.526097
5. Sweet Dry Whey (lactose)	13097.15812	
6. HRC®	10067.96161	
7. Linoleic Acid (Soybean Oil)		662.7656173
8. Emulsified Vegetable Oil		1104.609362
9. Lactoil Product		530.2124938
10. Lactic Acid Product		0
11. Hydrogen Gas	198.16	j

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 73			RETURN TO COVER PAGE	
NOTE: Unshaded boxes are user input.					
. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes	
Width (Perpendicular to predominant groundwater flow direction)	80	1-10,000	feet		
Length (Parallel to predominant groundwater flow)	130	1-1,000	feet		
Saturated Thickness	24	1-100	feet		
Treatment Zone Cross Sectional Area	1920		ft <sup>2</sup>		
Treatment Zone Volume	249,600		ft <sup>3</sup>		
Treatment Zone Total Pore Volume (total volume x total porosity)	466,877		gallons		
Treatment Zone Effective Pore Volume (total volume x effective porosity)	373,501		gallons		
Design Period of Performance	1.3	.5 to 5	year		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
. Treatment Zone Hydrogeologic Properties	25%	.05-50	porcont	Default = 25%	
	23%		percent		
Effective Porosity Average Aquifer Hydraulic Conductivity	20% 11	.05-50 .01-1000	percent ft/day	Default = 20%	
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr		
Average Groundwater Discharge through the Treatment Zone	288,386		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
Native Electron Acceptors     A. Aqueous-Phase Native Electron Acceptors     Owners	2.0	0.01 to 10		Default = 5	
Oxygen Nitrate	2.0	0.01 to 10	mg/L mg/L	Default = 5	
Sulfate	12	10 to 5,000	mg/L	Default = 1 $Default = 50$	
	0.0	0.1 to 20	ů.		
Carbon Dioxide (estimated as the amount of Methane produced) B. Solid-Phase Native Electron Acceptors	0.0	0.1 to 20	mg/L	Default = 0	
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

### 4. Contaminant Electron Acceptors

mg/L
mg/L
mg/L
mg/L
mg/L
mg/L
mg/L
mg/L
mg/L
mg/L
mg/L
mg/L
mg/L

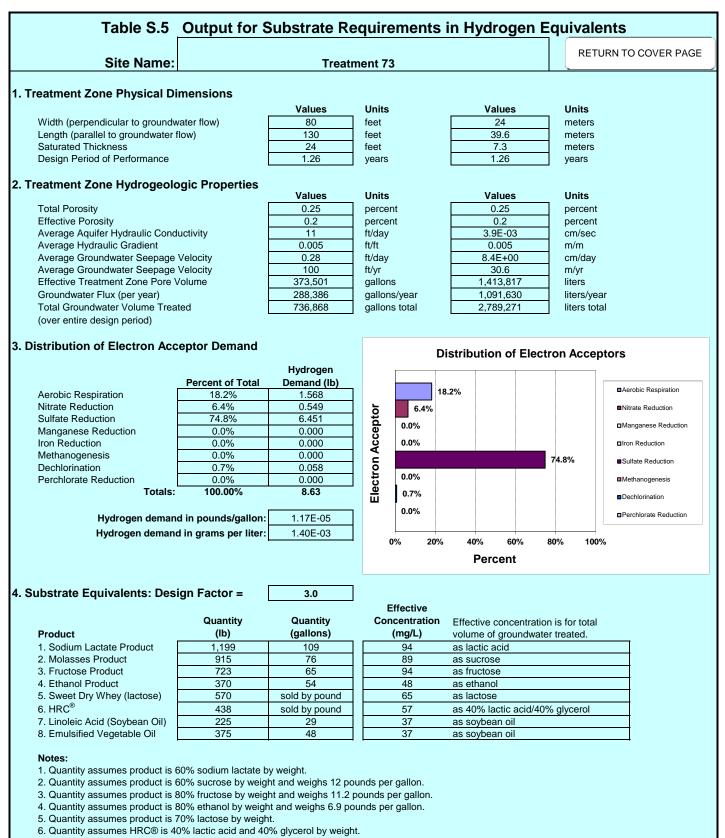
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



7. Quantity of neat soybean oil, corn oil, or canola oil.

8. Quantity assumes commercial product is 60% soybean oil by weight.

## Treatment 73

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		109.0443343
2. Molasses Product		76.28319291
3. Fructose Product		64.54033756
4. Ethanol Product		53.56653885
5. Sweet Dry Whey (lactose)	570.1377898	i
6. HRC®	438.2725875	
7. Linoleic Acid (Soybean Oil)		28.85112332
8. Emulsified Vegetable Oil		48.08520554
9. Lactoil Product		23.08089866
10. Lactic Acid Product		0
11. Hydrogen Gas	8.63	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents					
Site Name:	Treatment 74			RETURN TO COVER PAGE	
	NOTE: Unshaded	boxes are use	r input.		
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes	
Width (Perpendicular to predominant groundwater flow direction)	80	1-10,000	feet		
Length (Parallel to predominant groundwater flow)	40	1-1,000	feet		
Saturated Thickness	24	1-100	feet		
Treatment Zone Cross Sectional Area	1920		ft <sup>2</sup>		
Treatment Zone Volume	76,800		ft <sup>3</sup>		
Treatment Zone Total Pore Volume (total volume x total porosity)	143,654		gallons		
Treatment Zone Effective Pore Volume (total volume x effective porosity)	114,924		gallons		
Design Period of Performance	2.0	.5 to 5	year		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
2. Treatment Zone Hydrogeologic Properties					
Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity	20%	.05-50	percent	Default = 20%	
Average Aquifer Hydraulic Conductivity	11	.01-1000	ft/day		
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr		
Average Groundwater Discharge through the Treatment Zone	288,386		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
3. Native Electron Acceptors <u>A. Aqueous-Phase Native Electron Acceptors</u>	[				
Oxygen	0.8	0.01 to 10	mg/L	Default = 5	
Nitrate	0.00	0.1 to- 20	mg/L	Default = 1	
Sulfate	2	10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0	
B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

## 4. Contaminant Electron Acceptors

0.000	mg/L	
0.000	mg/L	
0.241	mg/L	
0.057	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.000	mg/L	
0.000	mg/L	
	mg/L	
	mg/L	
	0.000 0.241 0.057 0.000 0.000	0.000          mg/L           0.241          mg/L           0.057          mg/L            mg/L         mg/L           0.000          mg/L           0.000          mg/L            mg/L         mg/L

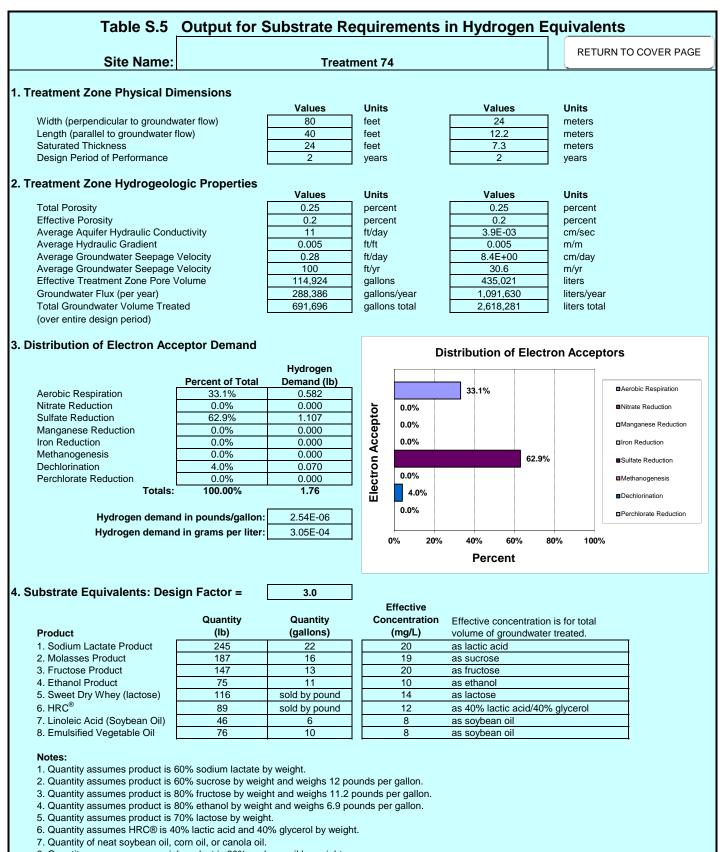
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



## Treatment 74

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		22.2402854
2. Molasses Product		15.55844229
3. Fructose Product		13.16341227
4. Ethanol Product		10.92523623
5. Sweet Dry Whey (lactose)	116.2832278	
6. HRC®	89.38848125	
7. Linoleic Acid (Soybean Oil)		5.88437007
8. Emulsified Vegetable Oil		9.807283449
9. Lactoil Product		4.707496056
10. Lactic Acid Product		0
11. Hydrogen Gas	1.76	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents							
Site Name:	Treatment 75			RETURN TO COVER PAGE			
	NOTE: Unshaded boxes are user input.						
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes			
Width (Perpendicular to predominant groundwater flow direction)	170	1-10,000	feet				
Length (Parallel to predominant groundwater flow)	20	1-1,000	feet				
Saturated Thickness	24	1-100	feet				
Treatment Zone Cross Sectional Area	4080		ft <sup>2</sup>				
Treatment Zone Volume	81,600		ft <sup>3</sup>				
Treatment Zone Total Pore Volume (total volume x total porosity)	152,633		gallons				
Treatment Zone Effective Pore Volume (total volume x effective porosity)	122,106		gallons				
Design Period of Performance	2.0	.5 to 5	year				
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3			
2. Treatment Zone Hydrogeologic Properties							
Total Porosity	25%	.05-50	percent	Default = 25%			
Effective Porosity	20%	.05-50	percent	Default = 20%			
Average Aquifer Hydraulic Conductivity	11	.01-1000	ft/day				
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft				
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day				
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr				
Average Groundwater Discharge through the Treatment Zone	612,821		gallons/year				
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7			
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%			
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors							
Oxygen	5.0	0.01 to 10	mg/L	Default = 5			
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1			
Sulfate	50	10 to 5,000	mg/L	Default = 50			
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0			
B. Solid-Phase Native Electron Acceptors							
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0			
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0			

## 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L	
Trichloroethene (TCE)	4.050	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	48.334	mg/L	
Vinyl Chloride (VC)	0.349	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)	0.925	mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)	0.000	mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

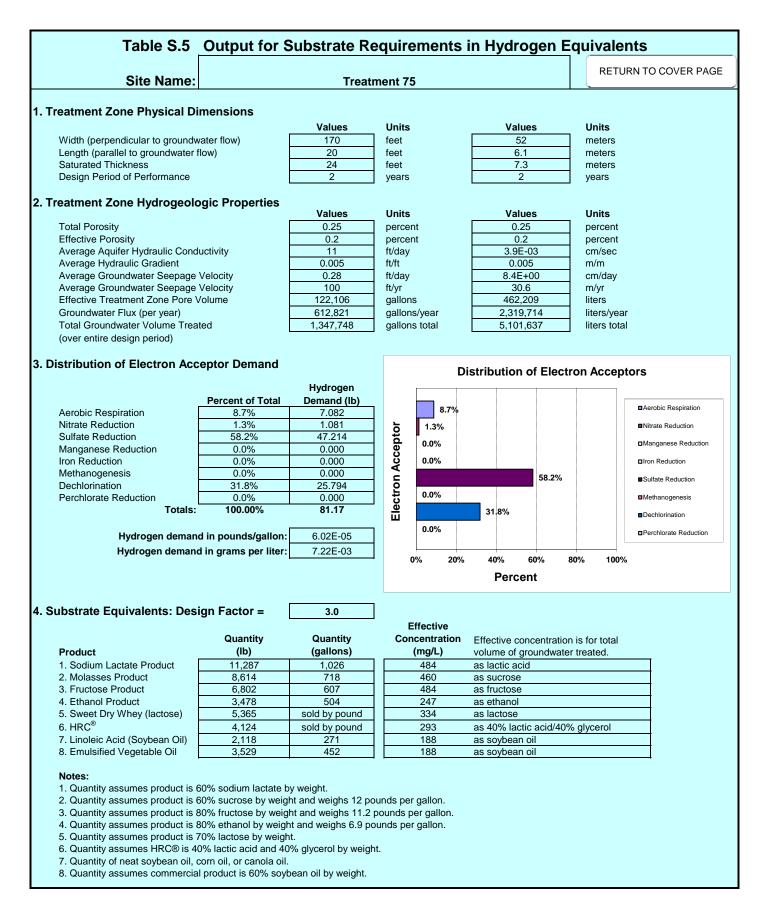
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



## Treatment 75

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		1026.090543
2. Molasses Product		717.8132032
3. Fructose Product		607.3147265
4. Ethanol Product		504.0529554
5. Sweet Dry Whey (lactose)	5364.909589	)
6. HRC®	4124.078161	
7. Linoleic Acid (Soybean Oil)		271.4846673
8. Emulsified Vegetable Oil		452.4744455
9. Lactoil Product		217.1877338
10. Lactic Acid Product		0
11. Hydrogen Gas	81.17	,

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents				
Site Name:	Treatment 76			RETURN TO COVER PAGE
	NOTE: Unshaded	boxes are use	r input.	
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes
Width (Perpendicular to predominant groundwater flow direction)	170	1-10,000	feet	
Length (Parallel to predominant groundwater flow)	30	1-1,000	feet	
Saturated Thickness	24	1-100	feet	
Treatment Zone Cross Sectional Area	4080		ft <sup>2</sup>	
Treatment Zone Volume	122,400		ft <sup>3</sup>	
Treatment Zone Total Pore Volume (total volume x total porosity)	228,949		gallons	
Treatment Zone Effective Pore Volume (total volume x effective porosity)	183,159		gallons	
Design Period of Performance	2.0	.5 to 5	year	
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3
2. Treatment Zone Hydrogeologic Properties				
Total Porosity	25%	.05-50	percent	Default = 25%
Effective Porosity	20%	.05-50	percent	Default = 20%
Average Aquifer Hydraulic Conductivity	11	.01-1000	ft/day	
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft	
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day	
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr	
Average Groundwater Discharge through the Treatment Zone	612,821		gallons/year	
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%
3. Native Electron Acceptors <u>A. Aqueous-Phase Native Electron Acceptors</u>	Γ			
Oxygen	0.8	0.01 to 10	mg/L	Default = 5
Nitrate	0.50	0.1 to- 20	mg/L	Default = 1
Sulfate	6	10 to 5,000	mg/L	Default = 50
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0
B. Solid-Phase Native Electron Acceptors				
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0

## 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L	
Trichloroethene (TCE)	0.001	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	11.036	mg/L	
Vinyl Chloride (VC)	24.655	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)	0.000	mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)	0.000	mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

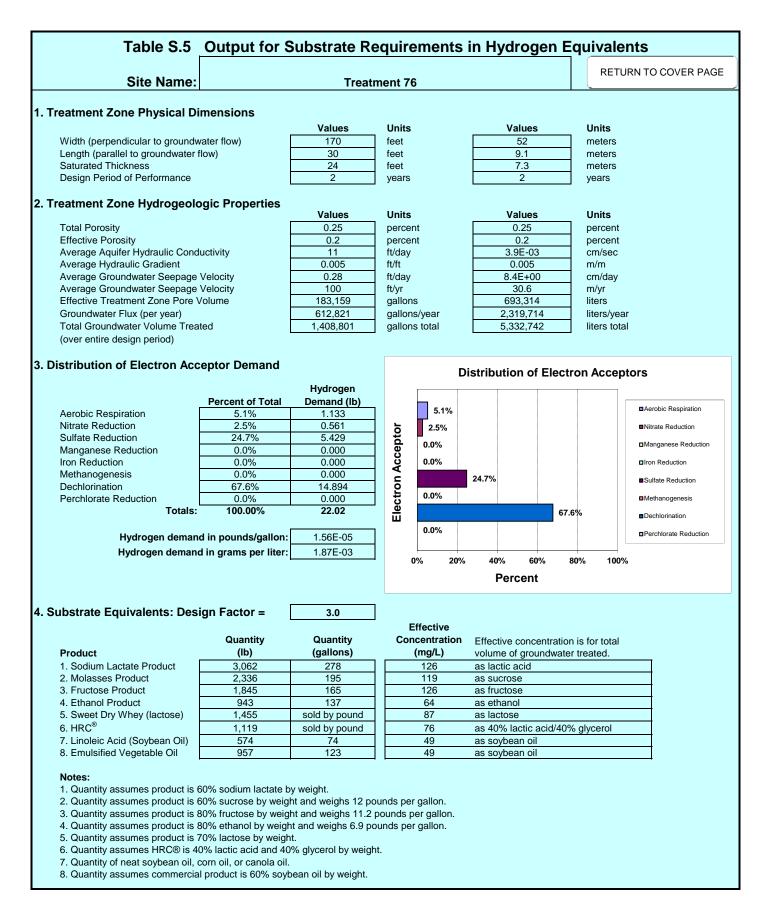
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



## Treatment 76

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		278.3193682
2. Molasses Product		194.7014505
3. Fructose Product		164.7295671
4. Ethanol Product		136.7205857
5. Sweet Dry Whey (lactose)	1455.19151	
6. HRC®	1118.625286	i
7. Linoleic Acid (Soybean Oil)		73.63818092
8. Emulsified Vegetable Oil		122.7303015
9. Lactoil Product		58.91054473
10. Lactic Acid Product		0
11. Hydrogen Gas	22.02	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 77			RETURN TO COVER PAGE		
NOTE: Unshaded boxes are user input.						
. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	200	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	20	1-1,000	feet			
Saturated Thickness	24	1-100	feet			
Treatment Zone Cross Sectional Area	4800		ft <sup>2</sup>			
Treatment Zone Volume	96,000		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	179,568		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	143,654		gallons			
Design Period of Performance	2.0	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
Treatment Zone Hydrogeologic Properties     Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	11	.03-30	ft/day			
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr			
Average Groundwater Discharge through the Treatment Zone	720,966		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
Native Electron Acceptors     A. Aqueous-Phase Native Electron Acceptors     Ovygon	5.0	0.01 to 10	mall	Default = 5		
Oxygen Nitrate	0.60	0.01 to 10	mg/L mg/L	Default = 5		
Sulfate	13	10 to 5,000	mg/L	Default = 1 Default = 50		
	0.0	0.1 to 20				
Carbon Dioxide (estimated as the amount of Methane produced) B. Solid-Phase Native Electron Acceptors	0.0	0.1 to 20	mg/L	Default = 0		
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

## 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L	
Trichloroethene (TCE)	2.993	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	13.467	mg/L	
Vinyl Chloride (VC)	17.592	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)	0.000	mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)	0.000	mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

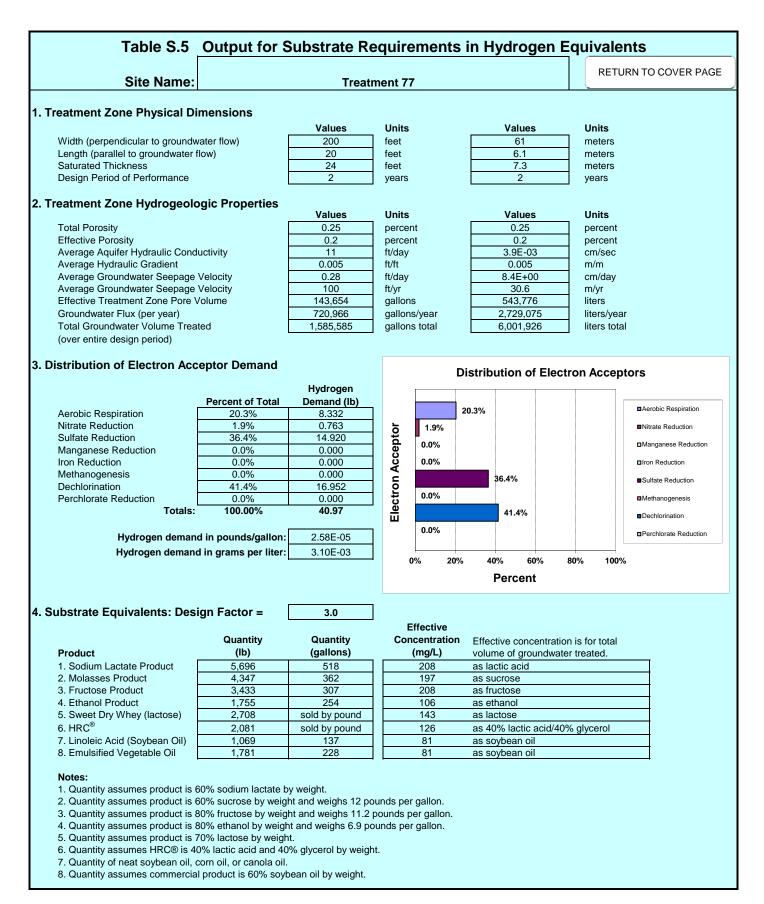
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 µs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



## Treatment 77

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		517.859732
2. Molasses Product		362.2746116
3. Fructose Product		306.5069097
4. Ethanol Product		254.3915156
5. Sweet Dry Whey (lactose)	2707.627178	
6. HRC®	2081.38943	
7. Linoleic Acid (Soybean Oil)		137.0161512
8. Emulsified Vegetable Oil		228.3602521
9. Lactoil Product		109.612921
10. Lactic Acid Product		0
11. Hydrogen Gas	40.97	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 92			RETURN TO COVER PAGE
	NOTE: Unshaded	boxes are use	r input.	
. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes
Width (Perpendicular to predominant groundwater flow direction)	50	1-10,000	feet	
Length (Parallel to predominant groundwater flow)	50	1-1,000	feet	
Saturated Thickness	24	1-100	feet	
Treatment Zone Cross Sectional Area	1200		ft <sup>2</sup>	
Treatment Zone Volume	60,000		ft <sup>3</sup>	
Treatment Zone Total Pore Volume (total volume x total porosity)	112,230		gallons	
Treatment Zone Effective Pore Volume (total volume x effective porosity)	89,784		gallons	
Design Period of Performance	1.1	.5 to 5	year	
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3
Treatment Zone Hydrogeologic Properties     Total Porosity	25%	.05-50	percent	Default = 25%
Effective Porosity	20%	.05-50	percent	Default = 20%
Average Aquifer Hydraulic Conductivity	11	.01-1000	ft/day	
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft	
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day	
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr	
Average Groundwater Discharge through the Treatment Zone	180,241		gallons/year	
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%
. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors				
Oxygen	1.5	0.01 to 10	mg/L	Default = 5
Nitrate	0.04	0.1 to- 20	mg/L	Default = 1
Sulfate	4	10 to 5,000	mg/L	Default = 50
Carbon Dioxide (estimated as the amount of Methane produced) B. Solid-Phase Native Electron Acceptors	0.0	0.1 to 20	mg/L	Default = 0
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0

## 4. Contaminant Electron Acceptors

0.168	mg/L	
8.090	mg/L	
60.938	mg/L	
5.220	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.000	mg/L	
0.002	mg/L	
	mg/L	
	mg/L	
	8.090 60.938 5.220 0.000	8.090          mg/L           60.938          mg/L           5.220          mg/L            mg/L             mg/L             mg/L             mg/L             mg/L            0.000          mg/L           0.002          mg/L            mg/L

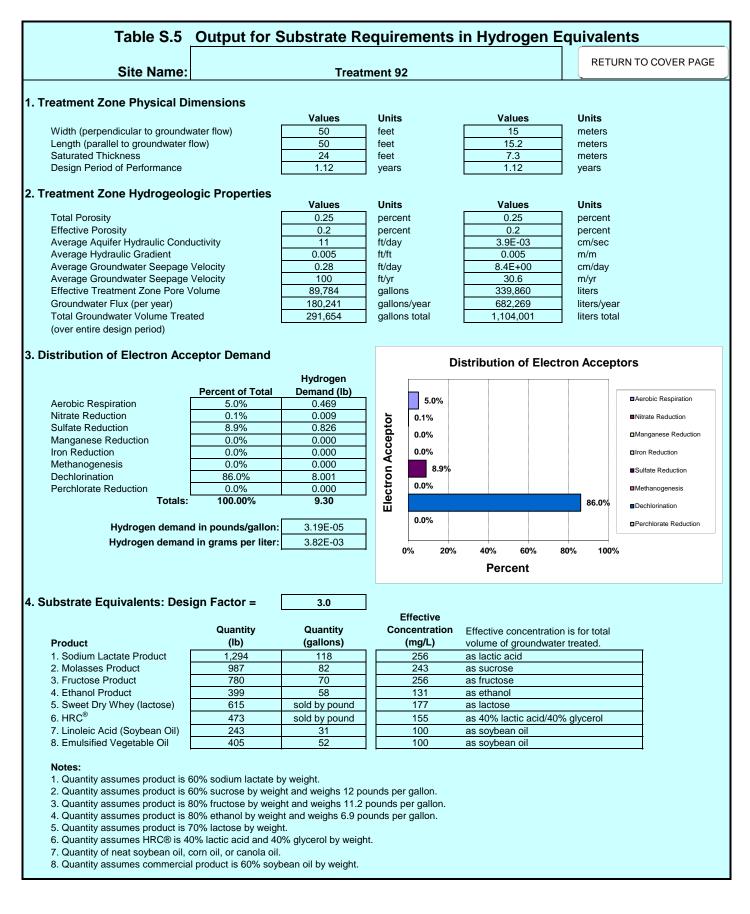
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



## Treatment 92

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		117.6203182
2. Molasses Product		82.28261916
3. Fructose Product		69.61622624
4. Ethanol Product		57.77937377
5. Sweet Dry Whey (lactose)	614.9772818	
6. HRC®	472.7413082	
7. Linoleic Acid (Soybean Oil)		31.1201708
8. Emulsified Vegetable Oil		51.86695134
9. Lactoil Product		24.89613664
10. Lactic Acid Product		0
11. Hydrogen Gas	9.30	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents						
Site Name:	Treatment 93			RETURN TO COVER PAGE		
	NOTE: Unshaded	boxes are use	r input.			
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	220	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	140	1-1,000	feet			
Saturated Thickness	24	1-100	feet			
Treatment Zone Cross Sectional Area	5280		ft <sup>2</sup>			
Treatment Zone Volume	739,200		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	1,382,674		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	1,106,139		gallons			
Design Period of Performance	1.1	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
2. Treatment Zone Hydrogeologic Properties						
Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	11	.01-1000	ft/day			
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr			
Average Groundwater Discharge through the Treatment Zone	793,062		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors						
Oxygen	0.6	0.01 to 10	mg/L	Default = 5		
Nitrate	0.00	0.1 to- 20	mg/L	Default = 1		
Sulfate	0	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors						
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

## 4. Contaminant Electron Acceptors

) mg/L
3 mg/L
I mg/L
7 mg/L
mg/L
mg/L
mg/L
mg/L
mg/L
) mg/L
) mg/L
mg/L
mg/L
03 21 57 57 00

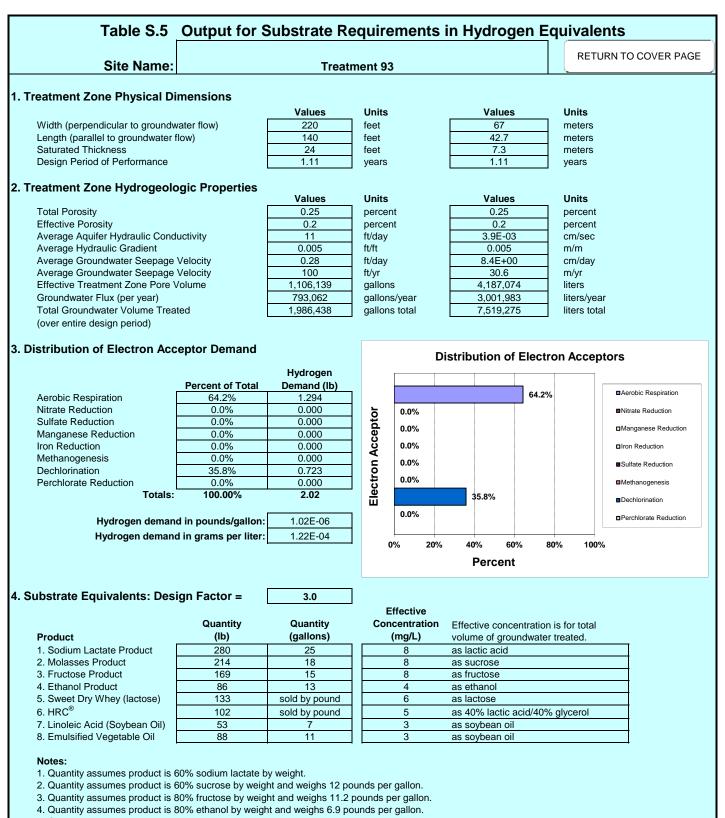
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.

## Treatment 93

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		25.49610943
2. Molasses Product		17.83609068
3. Fructose Product		15.09044482
4. Ethanol Product		12.52461529
5. Sweet Dry Whey (lactose)	133.3062885	5
6. HRC®	102.4743369	)
7. Linoleic Acid (Soybean Oil)		6.745801168
8. Emulsified Vegetable Oil		11.24300195
9. Lactoil Product		5.396640934
10. Lactic Acid Product		0
11. Hydrogen Gas	2.02	2

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 94			RETURN TO COVER PAGE
	NOTE: Unshaded	boxes are use	er input.	
. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes
Width (Perpendicular to predominant groundwater flow direction)	80	1-10,000	feet	
Length (Parallel to predominant groundwater flow)	130	1-1,000	feet	
Saturated Thickness	24	1-100	feet	
Treatment Zone Cross Sectional Area	1920		ft <sup>2</sup>	
Treatment Zone Volume	249,600		ft <sup>3</sup>	
Treatment Zone Total Pore Volume (total volume x total porosity)	466,877		gallons	
Treatment Zone Effective Pore Volume (total volume x effective porosity)	373,501		gallons	
Design Period of Performance	1.1	.5 to 5	year	
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3
. Treatment Zone Hydrogeologic Properties Total Porosity	25%	.05-50	percent	Default = 25%
Effective Porosity	20%	.05-50	I I I I I I I I I I I I I I I I I I I	Default = 20%
Average Aquifer Hydraulic Conductivity	11	.03-50	percent ft/day	
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft	
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day	
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr	
Average Groundwater Discharge through the Treatment Zone	288,386		gallons/year	
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%
. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors	0.9	0.01 to 10	~~~//	Default = 5
Oxygen Nitrate	0.9	0.01 to 10 0.1 to- 20	mg/L mg/L	Default = 5
Sulfate	0.00	10 to 5,000	mg/∟	Default = 1 Default = 50
	0.0	0.1 to 20	ů.	
Carbon Dioxide (estimated as the amount of Methane produced) B. Solid-Phase Native Electron Acceptors	0.0	0.1 to 20	mg/L	Default = 0
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0

## 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L	
Trichloroethene (TCE)	0.000	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.090	mg/L	
Vinyl Chloride (VC)	0.243	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)	0.000	mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)	0.000	mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

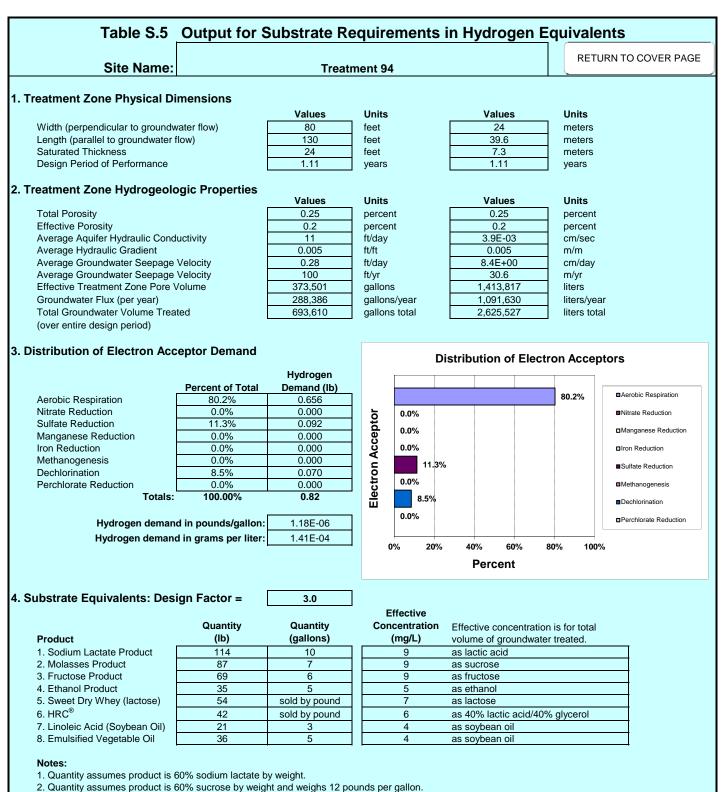
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.

4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.

5. Quantity assumes product is 70% lactose by weight.

6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

8. Quantity assumes commercial product is 60% soybean oil by weight.

## Treatment 94

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		10.34036957
2. Molasses Product		7.233722062
3. Fructose Product		6.120179897
4. Ethanol Product		5.079565223
5. Sweet Dry Whey (lactose)	54.06457374	Ļ
6. HRC®	41.56016502	2
7. Linoleic Acid (Soybean Oil)		2.73587142
8. Emulsified Vegetable Oil		4.559785699
9. Lactoil Product		2.188697136
10. Lactic Acid Product		0
11. Hydrogen Gas	0.82	2

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 95			RETURN TO COVER PAGE		
NOTE: Unshaded boxes are user input.						
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
		-				
Width (Perpendicular to predominant groundwater flow direction)	80	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	40	1-1,000	feet			
Saturated Thickness	24	1-100	feet ft <sup>2</sup>			
Treatment Zone Cross Sectional Area	1920		ft <sup>3</sup>			
Treatment Zone Volume	76,800					
Treatment Zone Total Pore Volume (total volume x total porosity)	143,654		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	114,924		gallons			
Design Period of Performance	1.1	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
2. Treatment Zone Hydrogeologic Properties						
Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	11	.01-1000	ft/day			
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr			
Average Groundwater Discharge through the Treatment Zone	288,386		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors						
Oxygen	0.6	0.01 to 10	mg/L	Default = 5		
Nitrate	0.00	0.1 to- 20	mg/L	Default = 1		
Sulfate	1	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors						
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

## 4. Contaminant Electron Acceptors

0.000	mg/L	
0.000	mg/L	
0.172	mg/L	
0.132	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.000	mg/L	
0.000	mg/L	
	mg/L	
	mg/L	
	0.000 0.172 0.132 0.000 0.000	0.000          mg/L           0.172          mg/L           0.132          mg/L            mg/L             mg/L             mg/L             mg/L             mg/L            0.000          mg/L           0.000          mg/L           0.000          mg/L            mg/L

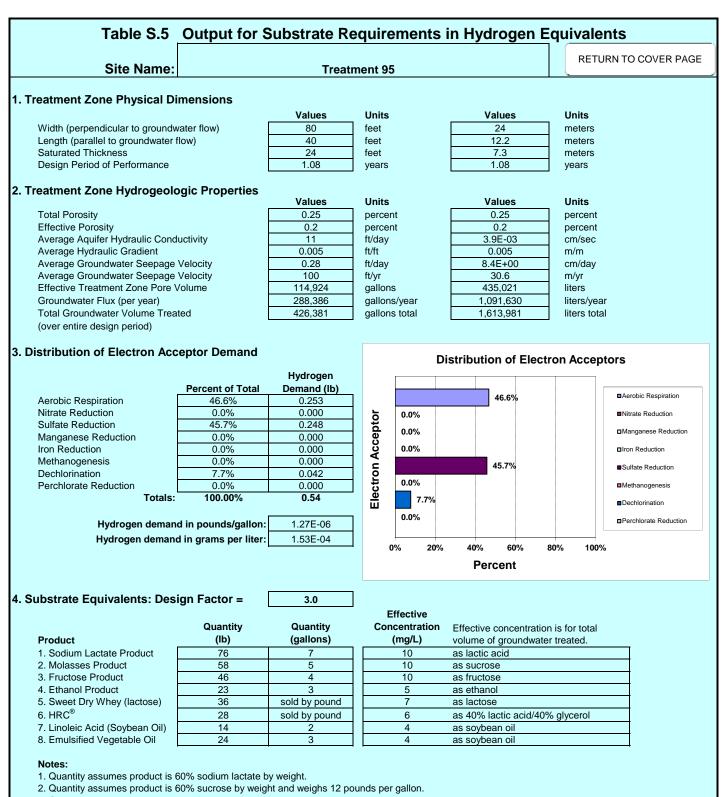
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.

## Treatment 95

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		6.865323417
2. Molasses Product		4.802714366
3. Fructose Product		4.063395809
4. Ethanol Product		3.372496297
5. Sweet Dry Whey (lactose)	35.89531126	3
6. HRC®	27.59320857	7
7. Linoleic Acid (Soybean Oil)		1.816438185
8. Emulsified Vegetable Oil		3.027396974
9. Lactoil Product		1.453150548
10. Lactic Acid Product		0
11. Hydrogen Gas	0.54	1

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents				
Site Name:	Treatment 96			RETURN TO COVER PAGE
	NOTE: Unshaded	boxes are use	r input.	
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes
Width (Perpendicular to predominant groundwater flow direction)	170	1-10,000	feet	
Length (Parallel to predominant groundwater flow)	30	1-1,000	feet	
Saturated Thickness	24	1-100	feet	
Treatment Zone Cross Sectional Area	4080		ft <sup>2</sup>	
Treatment Zone Volume	122,400		ft <sup>3</sup>	
Treatment Zone Total Pore Volume (total volume x total porosity)	228,949		gallons	
Treatment Zone Effective Pore Volume (total volume x effective porosity)	183,159		gallons	
Design Period of Performance	1.1	.5 to 5	year	
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3
2. Treatment Zone Hydrogeologic Properties				
Total Porosity	25%	.05-50	percent	Default = 25%
Effective Porosity	20%	.05-50	percent	Default = 20%
Average Aquifer Hydraulic Conductivity	11	.01-1000	ft/day	
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft	
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day	
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr	
Average Groundwater Discharge through the Treatment Zone	612,821		gallons/year	
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors				
Oxygen	0.5	0.01 to 10	mg/L	Default = 5
Nitrate	0.00	0.1 to- 20	mg/L	Default = 1
Sulfate	25	10 to 5,000	mg/L	Default = 50
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0
B. Solid-Phase Native Electron Acceptors				
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0

## 4. Contaminant Electron Acceptors

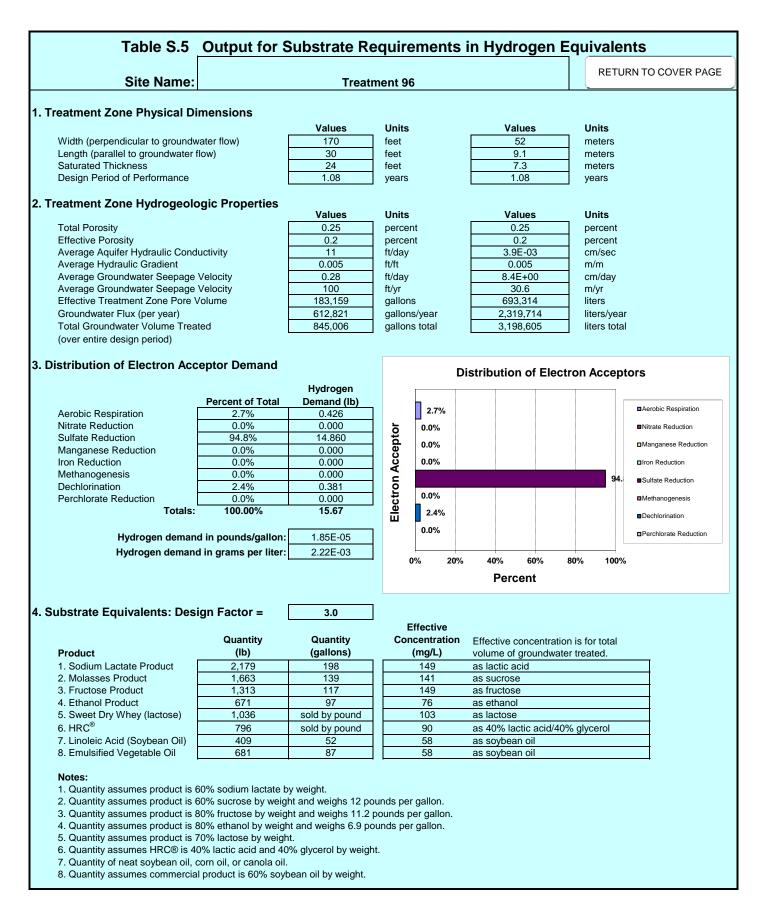
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



## Treatment 96

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		198.0575349
2. Molasses Product		138.5533806
3. Fructose Product		117.224799
4. Ethanol Product		97.2930571
5. Sweet Dry Whey (lactose)	1035.542892	
6. HRC®	796.0357491	
7. Linoleic Acid (Soybean Oil)		52.40237747
8. Emulsified Vegetable Oil		87.33729579
9. Lactoil Product		41.92190198
10. Lactic Acid Product		0
11. Hydrogen Gas	15.67	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents						
Site Name:	Treatment 97			RETURN TO COVER PAGE		
NOTE: Unshaded boxes are user input.						
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	170	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	30	1-1,000	feet			
Saturated Thickness	24	1-100	feet			
Treatment Zone Cross Sectional Area	4080		ft <sup>2</sup>			
Treatment Zone Volume	122,400		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	228,949		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	183,159		gallons			
Design Period of Performance	1.1	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
2. Treatment Zone Hydrogeologic Properties						
Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	11	.01-1000	ft/day			
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr			
Average Groundwater Discharge through the Treatment Zone	612,821		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors						
Oxygen	1.2	0.01 to 10	mg/L	Default = 5		
Nitrate	0.00	0.1 to- 20	mg/L	Default = 1		
Sulfate	4	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors						
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L	
Trichloroethene (TCE)	0.000	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	3.623	mg/L	
Vinyl Chloride (VC)	18.884	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)	0.000	mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)	0.000	mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

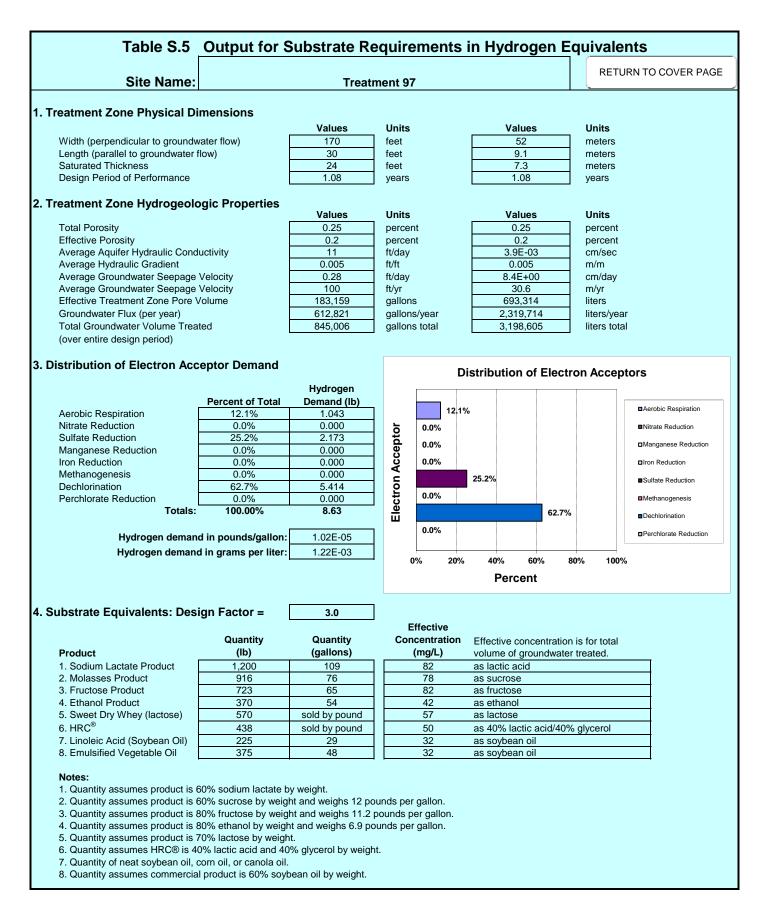
# 5. Aquifer Geochemistry (Optional Screening Parameters)

### A. Aqueous Geochemistry

Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



# Treatment 97

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		109.0919564
2. Molasses Product		76.31650751
3. Fructose Product		64.56852379
4. Ethanol Product		53.58993257
5. Sweet Dry Whey (lactose)	570.3867818	
6. HRC®	438.463991	
7. Linoleic Acid (Soybean Oil)		28.86372326
8. Emulsified Vegetable Oil		48.10620543
9. Lactoil Product		23.09097861
10. Lactic Acid Product		0
11. Hydrogen Gas	8.63	•

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents						
Site Name:	Treatment 98			RETURN TO COVER PAGE		
NOTE: Unshaded boxes are user input.						
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	170	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	30	1-1,000	feet			
Saturated Thickness	24	1-100	feet			
Treatment Zone Cross Sectional Area	4080		ft <sup>2</sup>			
Treatment Zone Volume	122,400		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	228,949		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	183,159		gallons			
Design Period of Performance	1.1	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
2. Treatment Zone Hydrogeologic Properties						
Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	11	.01-1000	ft/day			
Average Hydraulic Gradient	0.005	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.28		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	100.4		ft/yr			
Average Groundwater Discharge through the Treatment Zone	612,821		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors						
Oxygen	0.4	0.01 to 10	mg/L	Default = 5		
Nitrate	0.00	0.1 to- 20	mg/L	Default = 1		
Sulfate	7	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors						
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

## 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L	
Trichloroethene (TCE)	0.014	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	1.240	mg/L	
Vinyl Chloride (VC)	4.357	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)	0.000	mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)	0.000	mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

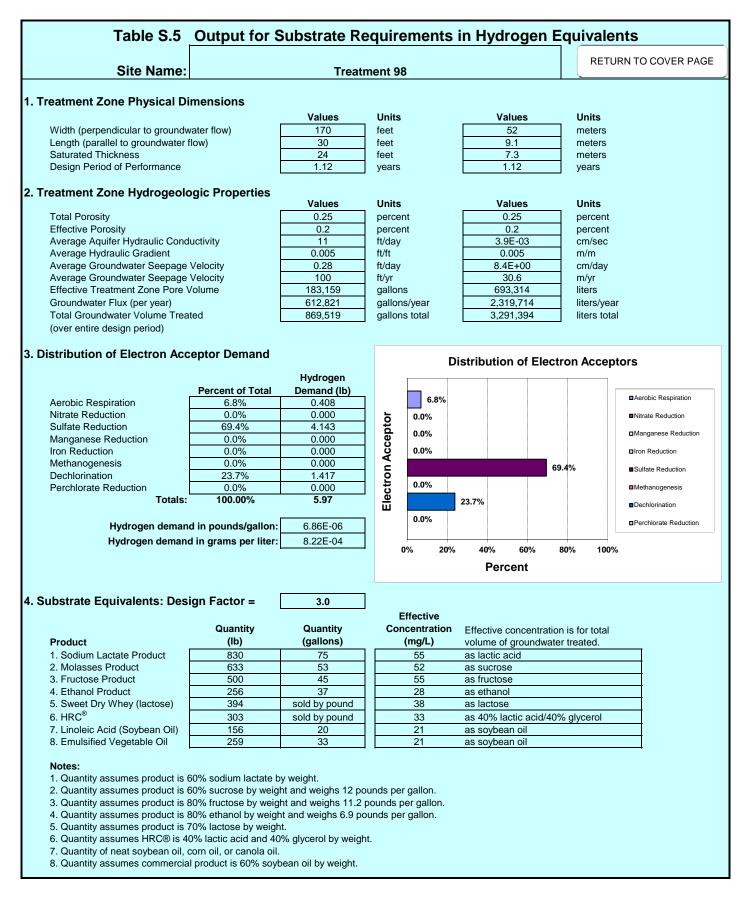
# 5. Aquifer Geochemistry (Optional Screening Parameters)

### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 µs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



## Treatment 98

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		75.44217219
2. Molasses Product		52.77642176
3. Fructose Product		44.65214349
4. Ethanol Product		37.05993599
5. Sweet Dry Whey (lactose)	394.4490432	
6. HRC®	303.2182849	1
7. Linoleic Acid (Soybean Oil)		19.96060986
8. Emulsified Vegetable Oil		33.2676831
9. Lactoil Product		15.96848789
10. Lactic Acid Product		0
11. Hydrogen Gas	5.97	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 115			RETURN TO COVER PAGE	
NOTE: Unshaded boxes are user input.					
. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes	
Width (Perpendicular to predominant groundwater flow direction)	120	1-10.000	feet		
Length (Parallel to predominant groundwater flow)	165	1-1,000	feet		
Saturated Thickness	36.5	1-100	feet		
Treatment Zone Cross Sectional Area	4380		ft <sup>2</sup>		
Treatment Zone Volume	722,700		ft <sup>3</sup>		
Treatment Zone Total Pore Volume (total volume x total porosity)	1,351,810		gallons		
Treatment Zone Effective Pore Volume (total volume x effective porosity)	1,081,448		gallons		
Design Period of Performance	5.5	.5 to 5	year		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
. Treatment Zone Hydrogeologic Properties Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity Average Aquifer Hydraulic Conductivity	20% 60	.05-50	percent ft/day	Default = 20%	
Average Hydraulic Gradient	0.002	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	0.60		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	219.0		ft/yr		
Average Groundwater Discharge through the Treatment Zone	1,435,377		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
3. Native Electron Acceptors					
Oxygen	1.1	0.01 to 10	mg/L	Default = 5	
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1	
Sulfate	13	10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced) B. Solid-Phase Native Electron Acceptors	0.0	0.1 to 20	mg/L	Default = 0	
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

# 4. Contaminant Electron Acceptors

0.358	mg/L	
0.437	mg/L	
2.157	mg/L	
0.056	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.785	mg/L	
0.865	mg/L	
	mg/L	
	mg/L	
	0.437 2.157 0.056 0.785	0.437          mg/L           2.157          mg/L           0.056          mg/L            mg/L             mg/L             mg/L             mg/L             mg/L            0.785          mg/L           0.865          mg/L            mg/L

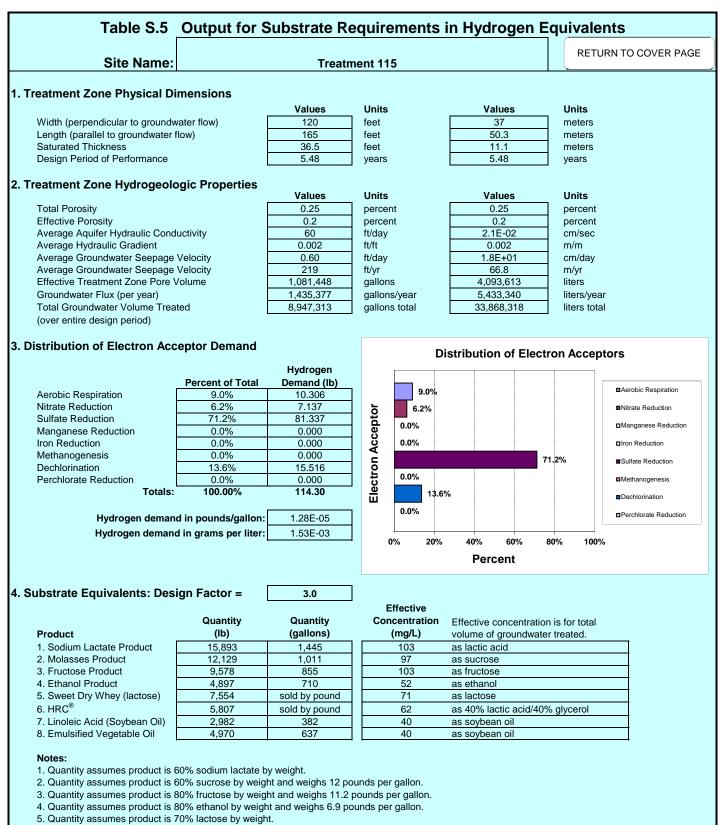
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

# Treatment 115

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		1444.855042
2. Molasses Product		1010.764627
3. Fructose Product		855.1698974
4. Ethanol Product		709.7652919
5. Sweet Dry Whey (lactose)	7554.417804	
6. HRC®	5807.182575	
7. Linoleic Acid (Soybean Oil)		382.2820441
8. Emulsified Vegetable Oil		637.1367402
9. Lactoil Product		305.8256353
10. Lactic Acid Product		0
11. Hydrogen Gas	114.30	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 117			RETURN TO COVER PAGE
	NOTE: Unshaded	boxes are use	r input.	
. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes
Width (Perpendicular to predominant groundwater flow direction)	165	1-10,000	feet	
Length (Parallel to predominant groundwater flow)	120	1-1,000	feet	
Saturated Thickness	36	1-100	feet	
Treatment Zone Cross Sectional Area	5940		ft <sup>2</sup>	
Treatment Zone Volume	712,800		ft <sup>3</sup>	
Treatment Zone Total Pore Volume (total volume x total porosity)	1,333,292		gallons	
Treatment Zone Effective Pore Volume (total volume x effective porosity)	1,066,634		gallons	
Design Period of Performance	2.4	.5 to 5	year	
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3
. Treatment Zone Hydrogeologic Properties Total Porosity	25%	.05-50	percent	Default = 25%
Effective Porosity	20%	.05-50	percent	Default = 20%
Average Aquifer Hydraulic Conductivity	60	.01-1000	ft/day	
Average Hydraulic Gradient	0.002	0.0001-0.1	ft/ft	
Average Groundwater Seepage Velocity through the Treatment Zone	0.60		ft/day	
Average Groundwater Seepage Velocity through the Treatment Zone	219.0		ft/yr	
Average Groundwater Discharge through the Treatment Zone	1,946,607		gallons/year	
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%
. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors				
Oxygen	5.0	0.01 to 10	mg/L	Default = 5
Nitrate	1.12	0.1 to- 20	mg/L	Default = 1
Sulfate	14	10 to 5,000	mg/L	Default = 50
Carbon Dioxide (estimated as the amount of Methane produced) B. Solid-Phase Native Electron Acceptors	0.0	0.1 to 20	mg/L	Default = 0
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.085	mg/L	
Trichloroethene (TCE)	0.394	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	2.499	mg/L	
Vinyl Chloride (VC)	0.038	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)		mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

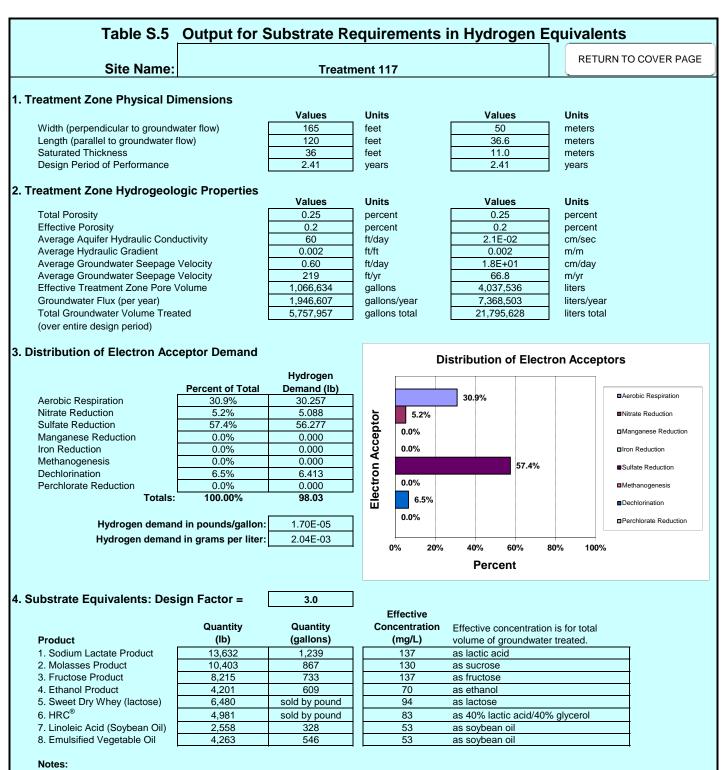
# 5. Aquifer Geochemistry (Optional Screening Parameters)

### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



1. Quantity assumes product is 60% sodium lactate by weight.

2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.

3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.

4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.

5. Quantity assumes product is 70% lactose by weight.

6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

Treatment 117

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		1239.278001
2. Molasses Product		866.9508915
3. Fructose Product		733.4945098
4. Ethanol Product		608.7783801
5. Sweet Dry Whey (lactose)	6479.559209	
6. HRC®	4980.924316	
7. Linoleic Acid (Soybean Oil)		327.8901437
8. Emulsified Vegetable Oil		546.4835728
9. Lactoil Product		262.312115
10. Lactic Acid Product		0
11. Hydrogen Gas	98.03	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents					
Site Name:	Treatment 119			RETURN TO COVER PAGE	
	NOTE: Unshaded	boxes are use	r input.		
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes	
Width (Perpendicular to predominant groundwater flow direction)	165	1-10,000	feet		
Length (Parallel to predominant groundwater flow)	120	1-1,000	feet		
Saturated Thickness	36	1-100	feet		
Treatment Zone Cross Sectional Area	5940		ft <sup>2</sup>		
Treatment Zone Volume	712,800		ft <sup>3</sup>		
Treatment Zone Total Pore Volume (total volume x total porosity)	1,333,292		gallons		
Treatment Zone Effective Pore Volume (total volume x effective porosity)	1,066,634		gallons		
Design Period of Performance	2.8	.5 to 5	year		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
2. Treatment Zone Hydrogeologic Properties					
Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity	20%	.05-50	percent	Default = 20%	
Average Aquifer Hydraulic Conductivity	60	.01-1000	ft/day		
Average Hydraulic Gradient	0.002	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	0.60		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	219.0		ft/yr		
Average Groundwater Discharge through the Treatment Zone	1,946,607		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
3. Native Electron Acceptors <u>A. Aqueous-Phase Native Electron Acceptors</u>	г				
Oxygen	5.0	0.01 to 10	mg/L	Default = 5	
Nitrate	1.58	0.1 to- 20	mg/L	Default = 1	
Sulfate	3	10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0	
B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.398	mg/L
Trichloroethene (TCE)	0.046	mg/L
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.206	mg/L
Vinyl Chloride (VC)	0.065	mg/L
Carbon Tetrachloride (CT)		mg/L
Trichloromethane ( or chloroform) (CF)		mg/L
Dichloromethane (or methylene chloride) (MC)		mg/L
Chloromethane		mg/L
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L
Dichloroethane (1,1-DCA and 1,2-DCA)		mg/L
Chloroethane		mg/L
Perchlorate		mg/L

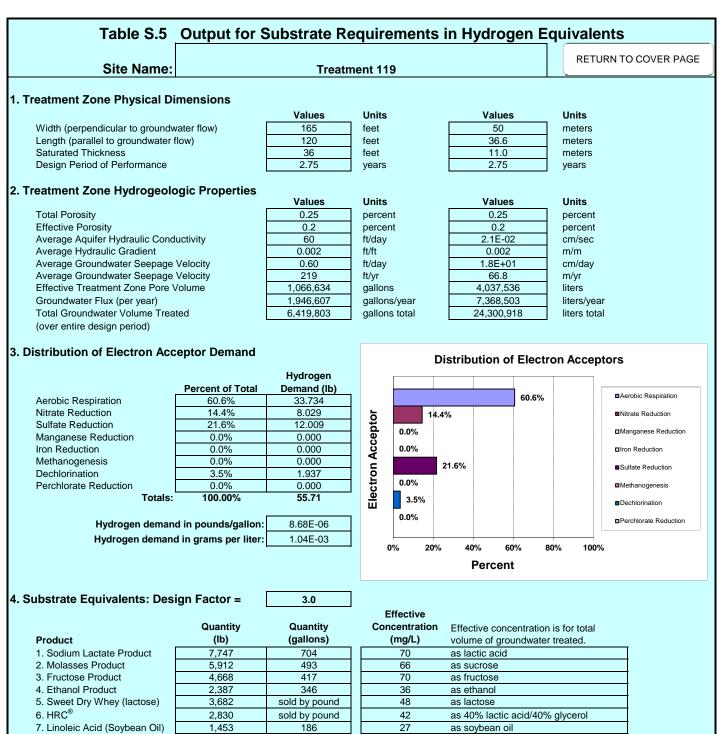
# 5. Aquifer Geochemistry (Optional Screening Parameters)

### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



8. Emulsified Vegetable Oil

#### Notes:

1. Quantity assumes product is 60% sodium lactate by weight.

2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.

311

2.422

3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.

4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.

5. Quantity assumes product is 70% lactose by weight.

6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

8. Quantity assumes commercial product is 60% soybean oil by weight.

27

as soybean oil

# Treatment 119

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		704.2417907
2. Molasses Product		492.6602812
3. Fructose Product		416.8213159
4. Ethanol Product		345.9491545
5. Sweet Dry Whey (lactose)	3682.124897	,
6. HRC®	2830.498934	
7. Linoleic Acid (Soybean Oil)		186.3294126
8. Emulsified Vegetable Oil		310.5490211
9. Lactoil Product		149.0635301
10. Lactic Acid Product		0
11. Hydrogen Gas	55.71	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 123			RETURN TO COVER PAGE
	NOTE: Unshaded	boxes are use	r input.	
. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes
Width (Perpendicular to predominant groundwater flow direction)	165	1-10,000	feet	
Length (Parallel to predominant groundwater flow)	120	1-1,000	feet	
Saturated Thickness	36	1-100	feet	
Treatment Zone Cross Sectional Area	5940		ft <sup>2</sup>	
Treatment Zone Volume	712,800		ft <sup>3</sup>	
Treatment Zone Total Pore Volume (total volume x total porosity)	1,333,292		gallons	
Treatment Zone Effective Pore Volume (total volume x effective porosity)	1,066,634		gallons	
Design Period of Performance	1.1	.5 to 5	year	
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3
. Treatment Zone Hydrogeologic Properties Total Porosity	25%	.05-50	percent	Default = 25%
Effective Porosity	20%	.05-50	percent	Default = 20%
Average Aquifer Hydraulic Conductivity	60	.01-1000	ft/day	
Average Hydraulic Gradient	0.002	0.0001-0.1	ft/ft	
Average Groundwater Seepage Velocity through the Treatment Zone	0.60		ft/day	
Average Groundwater Seepage Velocity through the Treatment Zone	219.0		ft/yr	
Average Groundwater Discharge through the Treatment Zone	1,946,607		gallons/year	
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%
. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors				
Oxygen	5.0	0.01 to 10	mg/L	Default = 5
Nitrate	1.36	0.1 to- 20	mg/L	Default = 1
Sulfate	6	10 to 5,000	mg/L	Default = 50
Carbon Dioxide (estimated as the amount of Methane produced) B. Solid-Phase Native Electron Acceptors	0.0	0.1 to 20	mg/L	Default = 0
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.161	mg/L
Trichloroethene (TCE)	0.020	mg/L
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.094	mg/L
Vinyl Chloride (VC)	0.019	mg/L
Carbon Tetrachloride (CT)		mg/L
Trichloromethane ( or chloroform) (CF)		mg/L
Dichloromethane (or methylene chloride) (MC)		mg/L
Chloromethane		mg/L
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L
Dichloroethane (1,1-DCA and 1,2-DCA)		mg/L
Chloroethane		mg/L
Perchlorate		mg/L

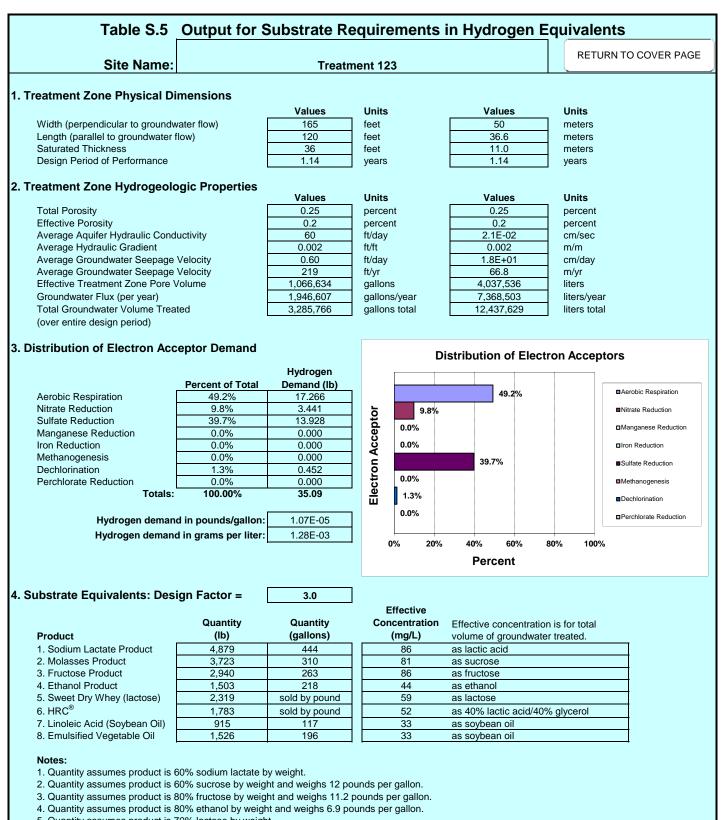
# 5. Aquifer Geochemistry (Optional Screening Parameters)

### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



5. Quantity assumes product is 70% lactose by weight.

6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

# Treatment 123

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		443.5417101
2. Molasses Product		310.2846018
3. Fructose Product		262.5201198
4. Ethanol Product		217.883803
5. Sweet Dry Whey (lactose)	2319.055749	
6. HRC®	1782.689346	
7. Linoleic Acid (Soybean Oil)		117.3529708
8. Emulsified Vegetable Oil		195.5882847
9. Lactoil Product		93.88237666
10. Lactic Acid Product		0
11. Hydrogen Gas	35.09	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 125			RETURN TO COVER PAGE
one name.				
1. Treatment Zone Physical Dimensions	NOTE: Unshaded Values	Range	Units	User Notes
	165			
Width (Perpendicular to predominant groundwater flow direction)		1-10,000	feet	
Length (Parallel to predominant groundwater flow) Saturated Thickness	120 36	1-1,000 1-100	feet	
			feet ft <sup>2</sup>	
Treatment Zone Cross Sectional Area	5940 712,800		ft <sup>3</sup>	
Treatment Zone Volume	1.333.292		gallons	
Treatment Zone Total Pore Volume (total volume x total porosity)	, , -		0	
Treatment Zone Effective Pore Volume (total volume x effective porosity) Design Period of Performance	1,066,634 1.9	 E to E	gallons	
5	3.0	.5 to 5	year	Default = 3
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3
2. Treatment Zone Hydrogeologic Properties		-		
Total Porosity	25%	.05-50	percent	Default = 25%
Effective Porosity	20%	.05-50	percent	Default = 20%
Average Aquifer Hydraulic Conductivity	60	.01-1000	ft/day	
Average Hydraulic Gradient	0.002	0.0001-0.1	ft/ft	
Average Groundwater Seepage Velocity through the Treatment Zone	0.60		ft/day	
Average Groundwater Seepage Velocity through the Treatment Zone	219.0		ft/yr	
Average Groundwater Discharge through the Treatment Zone	1,946,607		gallons/year	
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors		1		
Oxygen	5.0	0.01 to 10	mg/L	Default = 5
Nitrate	1.90	0.1 to- 20	mg/L	Default = 1
Sulfate	3	10 to 5,000	mg/L	Default = 50
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0
B. Solid-Phase Native Electron Acceptors				
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.490	mg/L
Trichloroethene (TCE)	0.031	mg/L
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.166	mg/L
Vinyl Chloride (VC)	0.057	mg/L
Carbon Tetrachloride (CT)		mg/L
Trichloromethane ( or chloroform) (CF)		mg/L
Dichloromethane (or methylene chloride) (MC)		mg/L
Chloromethane		mg/L
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L
Dichloroethane (1,1-DCA and 1,2-DCA)		mg/L
Chloroethane		mg/L
Perchlorate		mg/L

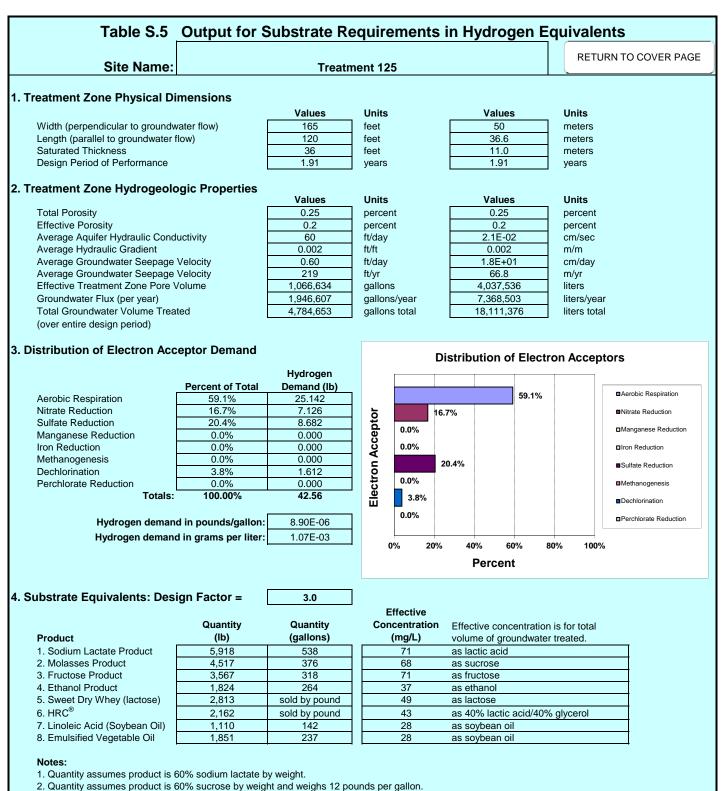
# 5. Aquifer Geochemistry (Optional Screening Parameters)

### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon. 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.

5. Quantity assumes product is 70% lactose by weight.

6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

# Treatment 125

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		538.038964
2. Molasses Product		376.3912207
3. Fructose Product		318.4504412
4. Ethanol Product		264.3042875
5. Sweet Dry Whey (lactose)	2813.134197	,
6. HRC®	2162.494095	j
7. Linoleic Acid (Soybean Oil)		142.3552045
8. Emulsified Vegetable Oil		237.2586742
9. Lactoil Product		113.8841636
10. Lactic Acid Product		0
11. Hydrogen Gas	42.56	;

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 128			RETURN TO COVER PAGE
Site Maille.		heree		
1. Treatment Zone Physical Dimensions	NOTE: Unshaded Values	Range	Units	User Notes
				User Notes
Width (Perpendicular to predominant groundwater flow direction)	80	1-10,000	feet	
Length (Parallel to predominant groundwater flow)	180	1-1,000	feet	
Saturated Thickness	12.5	1-100	feet ft <sup>2</sup>	
Treatment Zone Cross Sectional Area	1000		ft <sup>3</sup>	
Treatment Zone Volume	180,000			
Treatment Zone Total Pore Volume (total volume x total porosity)	336,690		gallons	
Treatment Zone Effective Pore Volume (total volume x effective porosity)	269,352		gallons	
Design Period of Performance	2.1	.5 to 5	year	
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3
2. Treatment Zone Hydrogeologic Properties				
Total Porosity	25%	.05-50	percent	Default = 25%
Effective Porosity	20%	.05-50	percent	Default = 20%
Average Aquifer Hydraulic Conductivity	2.3	.01-1000	ft/day	
Average Hydraulic Gradient	0.01	0.0001-0.1	ft/ft	
Average Groundwater Seepage Velocity through the Treatment Zone	0.12		ft/day	
Average Groundwater Seepage Velocity through the Treatment Zone	42.0		ft/yr	
Average Groundwater Discharge through the Treatment Zone	62,811		gallons/year	
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors	Г	1		
Oxygen	1.2	0.01 to 10	mg/L	Default = 5
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1
Sulfate	1	10 to 5,000	mg/L	Default = 50
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0
B. Solid-Phase Native Electron Acceptors				
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0

## 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L	
Trichloroethene (TCE)	0.006	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	1.513	mg/L	
Vinyl Chloride (VC)	0.008	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)	0.000	mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)	0.044	mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

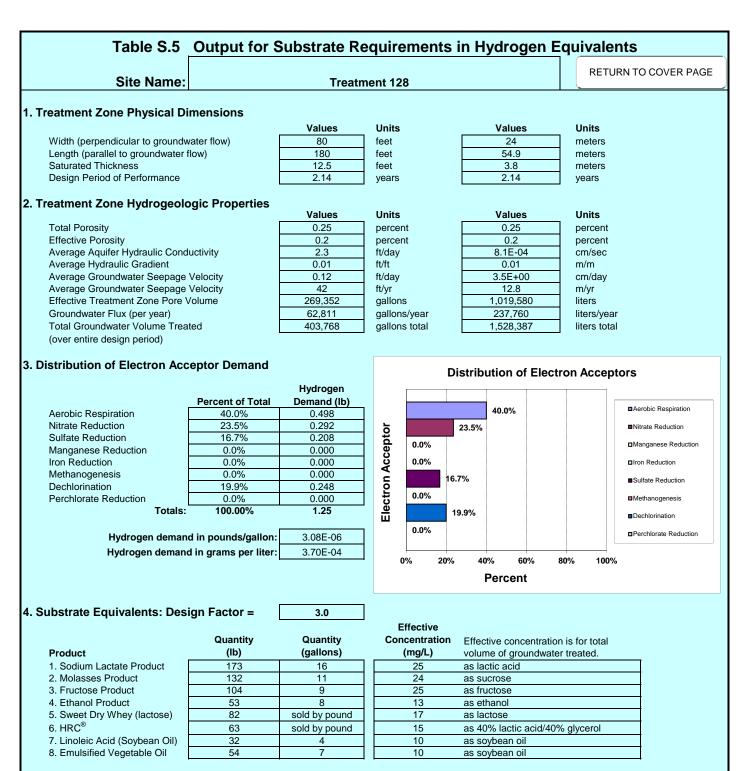
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



#### Notes:

1. Quantity assumes product is 60% sodium lactate by weight.

2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.

3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.

4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.

5. Quantity assumes product is 70% lactose by weight.

6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

# Treatment 128

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		15.740936
2. Molasses Product		11.01174917
3. Fructose Product		9.316626398
4. Ethanol Product		7.732519673
5. Sweet Dry Whey (lactose)	82.30140999	1
6. HRC®	63.26620084	
7. Linoleic Acid (Soybean Oil)		4.164761873
8. Emulsified Vegetable Oil		6.941269789
9. Lactoil Product		3.331809499
10. Lactic Acid Product		0
11. Hydrogen Gas	1.25	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 130			RETURN TO COVER PAGE	
Sile Maille.					
NOTE: Unshaded boxes are user input.					
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes	
Width (Perpendicular to predominant groundwater flow direction)	70	1-10,000	feet		
Length (Parallel to predominant groundwater flow)	355	1-1,000	feet		
Saturated Thickness	12.5	1-100	feet		
Treatment Zone Cross Sectional Area	875		ft <sup>2</sup>		
Treatment Zone Volume	310,625		ft <sup>3</sup>		
Treatment Zone Total Pore Volume (total volume x total porosity)	581,024		gallons		
Treatment Zone Effective Pore Volume (total volume x effective porosity)	464,819		gallons		
Design Period of Performance	2.4	.5 to 5	year		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
2. Treatment Zone Hydrogeologic Properties					
Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity	20%	.05-50	percent	Default = 20%	
Average Aquifer Hydraulic Conductivity	2.3	.01-1000	ft/day		
Average Hydraulic Gradient	0.01	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	0.12		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	42.0		ft/yr		
Average Groundwater Discharge through the Treatment Zone	54,960		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors					
Oxygen	1.2	0.01 to 10	mg/L	Default = 5	
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1	
Sulfate	6	10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0	
B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

# 4. Contaminant Electron Acceptors

0.000	mg/L	
0.005	mg/L	
1.174	mg/L	
0.056	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.000	mg/L	
0.041	mg/L	
	mg/L	
	mg/L	
	0.005 1.174 0.056 0.000	0.005          mg/L           1.174          mg/L           0.056          mg/L            mg/L             mg/L             mg/L             mg/L             mg/L            0.000          mg/L           0.001          mg/L           0.041          mg/L

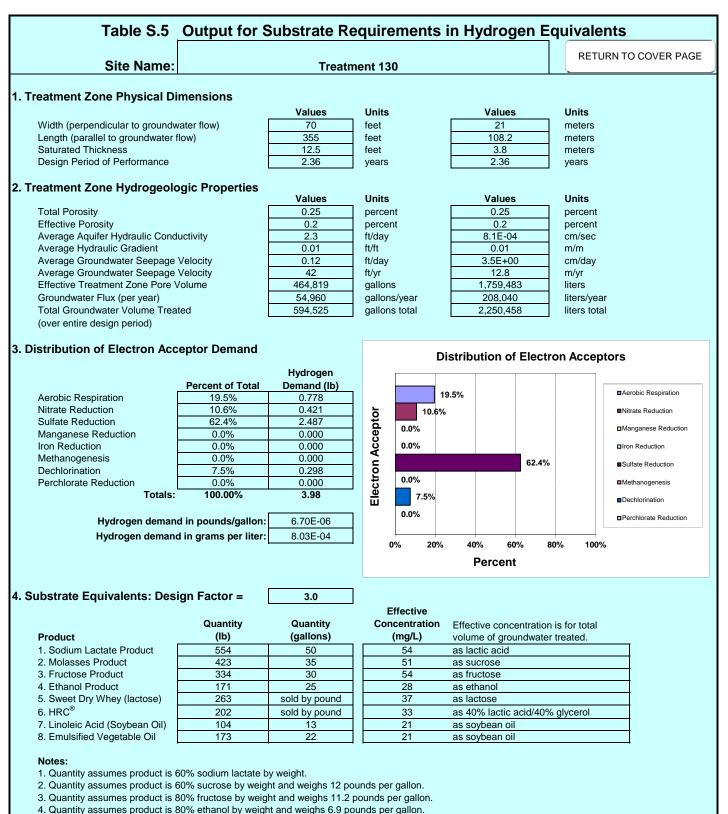
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



5. Quantity assumes product is 70% lactose by weight.

6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

Treatment 130

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		50.35879722
2. Molasses Product		35.22906412
3. Fructose Product		29.80598483
4. Ethanol Product		24.73807086
5. Sweet Dry Whey (lactose)	263.3007349	
6. HRC®	202.4028164	
7. Linoleic Acid (Soybean Oil)		13.32401063
8. Emulsified Vegetable Oil		22.20668439
9. Lactoil Product		10.65920851
10. Lactic Acid Product		0
11. Hydrogen Gas	3.98	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table 5.1 Input for S	Table S.1 Input for Substrate Requirements in Hydrogen Equivalents						
Site Name:	Treatment 131			RETURN TO COVER PAGE			
NOTE: Unshaded boxes are user input.							
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes			
Width (Perpendicular to predominant groundwater flow direction)	200	1-10,000	feet				
Length (Parallel to predominant groundwater flow)	590	1-1,000	feet				
Saturated Thickness	12.5	1-100	feet				
Treatment Zone Cross Sectional Area	2500		ft <sup>2</sup>				
Treatment Zone Volume	1,475,000		ft <sup>3</sup>				
Treatment Zone Total Pore Volume (total volume x total porosity)	2,758,988		gallons				
Treatment Zone Effective Pore Volume (total volume x effective porosity)	2,207,190		gallons				
Design Period of Performance	5.3	.5 to 5	year				
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3			
2. Treatment Zone Hydrogeologic Properties							
Total Porosity	25%	.05-50	percent	Default = 25%			
Effective Porosity	20%	.05-50	percent	Default = 20%			
Average Aquifer Hydraulic Conductivity	2.3	.01-1000	ft/day				
Average Hydraulic Gradient	0.01	0.0001-0.1	ft/ft				
Average Groundwater Seepage Velocity through the Treatment Zone	0.12		ft/day				
Average Groundwater Seepage Velocity through the Treatment Zone	42.0		ft/yr				
Average Groundwater Discharge through the Treatment Zone	157,028		gallons/year				
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7			
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%			
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors							
Oxygen	1.5	0.01 to 10	mg/L	Default = 5			
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1			
Sulfate	8	10 to 5,000	mg/L	Default = 50			
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0			
B. Solid-Phase Native Electron Acceptors							
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0			
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0			

# 4. Contaminant Electron Acceptors

0.000	mg/L	
0.003	mg/L	
0.805	mg/L	
0.073	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.000	mg/L	
0.019	mg/L	
	mg/L	
	mg/L	
	0.003 0.805 0.073	0.003          mg/L           0.805          mg/L           0.073          mg/L            mg/L         mg/L           0.000          mg/L           0.019          mg/L            mg/L         mg/L

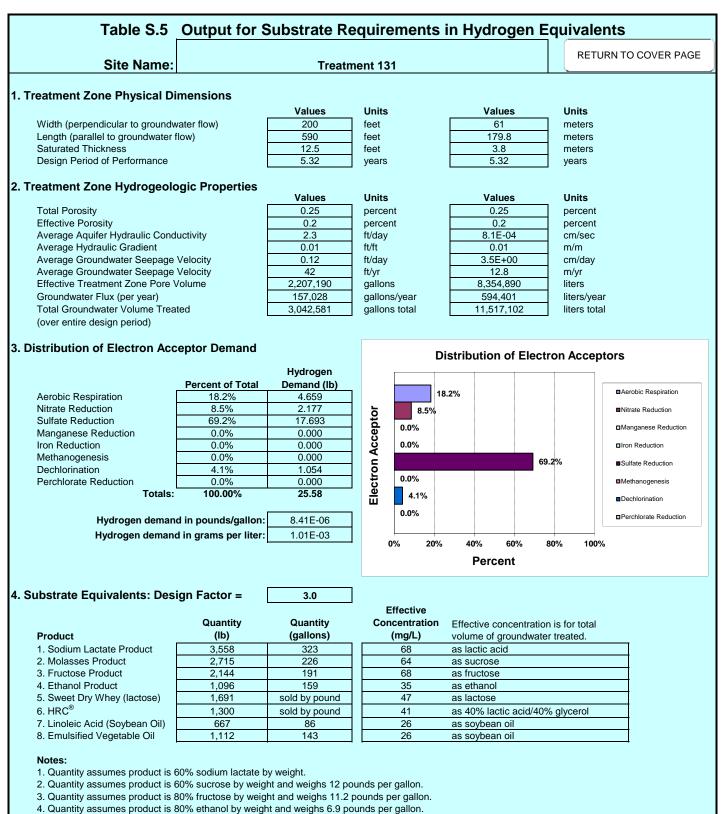
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



5. Quantity assumes product is 70% lactose by weight.

6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

Treatment 131

-		
Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		323.4130151
2. Molasses Product		226.2472195
3. Fructose Product		191.4192545
4. Ethanol Product		158.8722235
5. Sweet Dry Whey (lactose)	1690.963431	
6. HRC®	1299.866334	
7. Linoleic Acid (Soybean Oil)		85.56912973
8. Emulsified Vegetable Oil		142.6152162
9. Lactoil Product		68.45530378
10. Lactic Acid Product		0
11. Hydrogen Gas	25.58	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 135			RETURN TO COVER PAGE		
NOTE: Unshaded boxes are user input.						
. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	120	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	360	1-1,000	feet			
Saturated Thickness	7.5	1-100	feet			
Treatment Zone Cross Sectional Area	900		ft <sup>2</sup>			
Treatment Zone Volume	324,000		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	606,042		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	484,834		gallons			
Design Period of Performance	0.8	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
. Treatment Zone Hydrogeologic Properties Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	2.4	.01-1000	ft/day			
Average Hydraulic Gradient	0.004	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.05		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	17.5		ft/yr			
Average Groundwater Discharge through the Treatment Zone	23,595		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = $0.05\%$		
. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors	Γ					
Oxygen	1.5	0.01 to 10	mg/L	Default = 5		
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1		
Sulfate	1	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced) B. Solid-Phase Native Electron Acceptors	0.0	0.1 to 20	mg/L	Default = 0		
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

# 4. Contaminant Electron Acceptors

0.014	mg/L
0.100	mg/L
2.283	mg/L
0.997	mg/L
	mg/L
	mg/L
	mg/L
	mg/L
	mg/L
0.015	mg/L
0.199	mg/L
	mg/L
	mg/L
	0.100 2.283 0.997

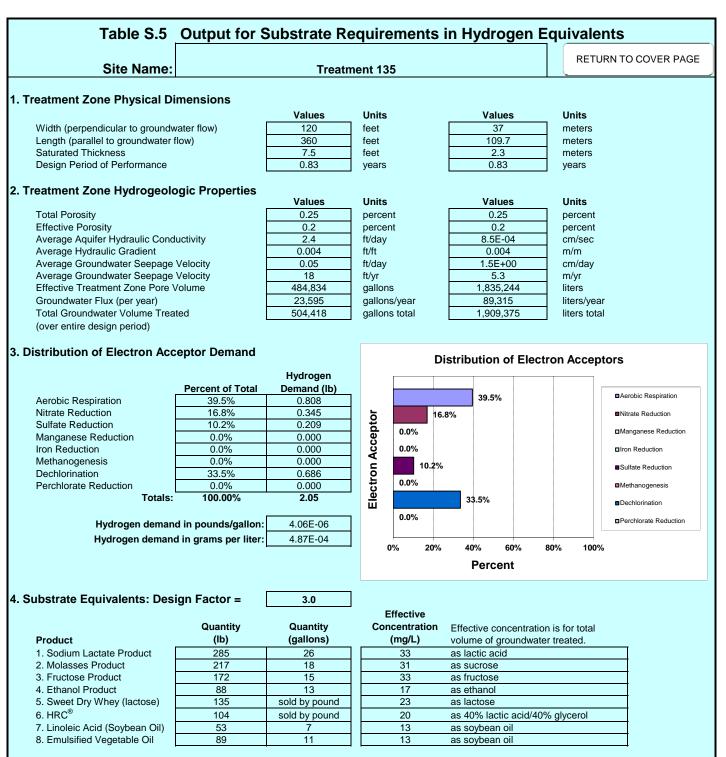
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



### Notes:

1. Quantity assumes product is 60% sodium lactate by weight.

2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.

3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.

4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.

5. Quantity assumes product is 70% lactose by weight.

6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

# Treatment 135

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		25.88954797
2. Molasses Product		18.11132505
3. Fructose Product		15.32331026
4. Ethanol Product		12.7178866
5. Sweet Dry Whey (lactose)	135.36338	
6. HRC®	104.0556509	
7. Linoleic Acid (Soybean Oil)		6.849897763
8. Emulsified Vegetable Oil		11.41649627
9. Lactoil Product		5.47991821
10. Lactic Acid Product		0
11. Hydrogen Gas	2.05	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 139			RETURN TO COVER PAGE		
NOTE: Unshaded boxes are user input.						
. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	320	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	115	1-1,000	feet			
Saturated Thickness	7	1-100	feet			
Treatment Zone Cross Sectional Area	2240		ft <sup>2</sup>			
Treatment Zone Volume	257,600		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	481,841		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	385,473		gallons			
Design Period of Performance	0.5	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
Treatment Zone Hydrogeologic Properties	25%	.05-50	porcont	Default = 25%		
			percent			
Effective Porosity Average Aquifer Hydraulic Conductivity	20% 2.4	.05-50	percent ft/day	Default = 20%		
Average Hydraulic Gradient	0.004	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.004	0.0001-0.1	ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	17.5		ft/yr			
Average Groundwater Discharge through the Treatment Zone	58,726		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors Oxygen	5.0	0.01 to 10	mg/L	Default = 5		
Nitrate	1.00	0.01 to 10	mg/L	Default = 5		
Sulfate	1	10 to 5,000	mg/L	Default = 1 $Default = 50$		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors	0.0	0.110.20	ing/∟			
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L	
Trichloroethene (TCE)	0.017	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.181	mg/L	
Vinyl Chloride (VC)	0.476	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)		mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

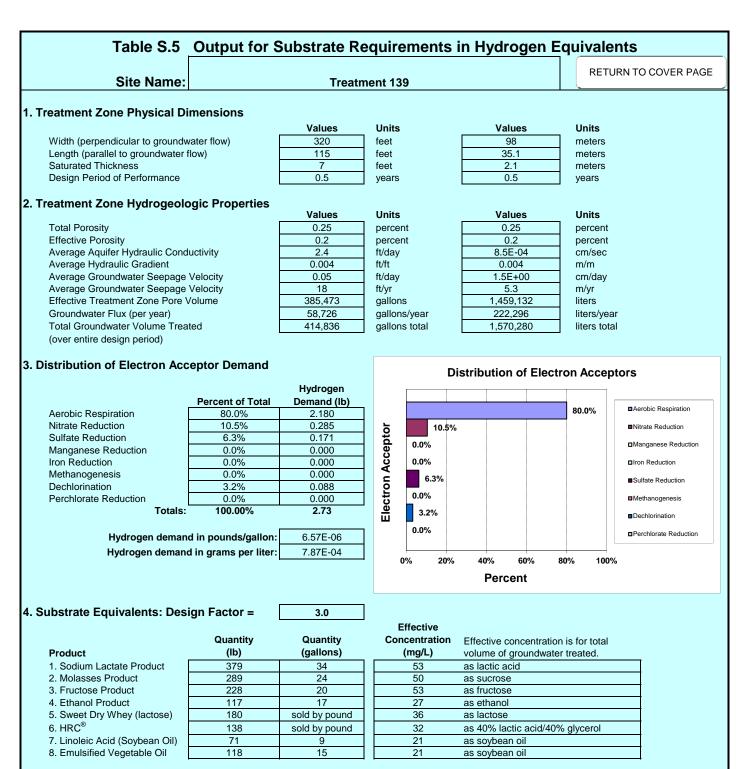
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



#### Notes:

1. Quantity assumes product is 60% sodium lactate by weight.

2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.

3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.

4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.

5. Quantity assumes product is 70% lactose by weight.

6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

# Treatment 139

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		34.44826384
2. Molasses Product		24.09867119
3. Fructose Product		20.38897841
4. Ethanol Product		16.92223879
5. Sweet Dry Whey (lactose)	180.1125858	
6. HRC®	138.4549673	
7. Linoleic Acid (Soybean Oil)		9.114376415
8. Emulsified Vegetable Oil		15.19062736
9. Lactoil Product		7.291501132
10. Lactic Acid Product		0
11. Hydrogen Gas	2.73	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 141			RETURN TO COVER PAGE		
NOTE: Unshaded boxes are user input.						
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	213	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	466	1-1,000	feet			
Saturated Thickness	7.5	1-100	feet			
Treatment Zone Cross Sectional Area	1597.5		ft <sup>2</sup>			
Treatment Zone Volume	744,435		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	1,392,466		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	1,113,973		gallons			
Design Period of Performance	0.8	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
	5.0	2 10 20	unitess	Deladit = 3		
2. Treatment Zone Hydrogeologic Properties		_				
Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	2.4	.01-1000	ft/day			
Average Hydraulic Gradient	0.004	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.05		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	17.5		ft/yr			
Average Groundwater Discharge through the Treatment Zone	41,882		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors						
A. Aqueous-Phase Native Electron Acceptors						
Oxygen	1.8	0.01 to 10	mg/L	Default = 5		
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1		
Sulfate	1	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors						
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L
Trichloroethene (TCE)	0.000	mg/L
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.430	mg/L
Vinyl Chloride (VC)	10.300	mg/L
Carbon Tetrachloride (CT)		mg/L
Trichloromethane ( or chloroform) (CF)		mg/L
Dichloromethane (or methylene chloride) (MC)		mg/L
Chloromethane		mg/L
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)	0.000	mg/L
Dichloroethane (1,1-DCA and 1,2-DCA)	0.134	mg/L
Chloroethane		mg/L
Perchlorate		mg/L
Dichloromethane (or methylene chloride) (MC)         Chloromethane         Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)         Trichloroethane (1,1,1-TCA and 1,1,2-TCA)         Dichloroethane (1,1-DCA and 1,2-DCA)         Chloroethane		mg/L            mg/L            mg/L            mg/L            mg/L            mg/L            mg/L            mg/L

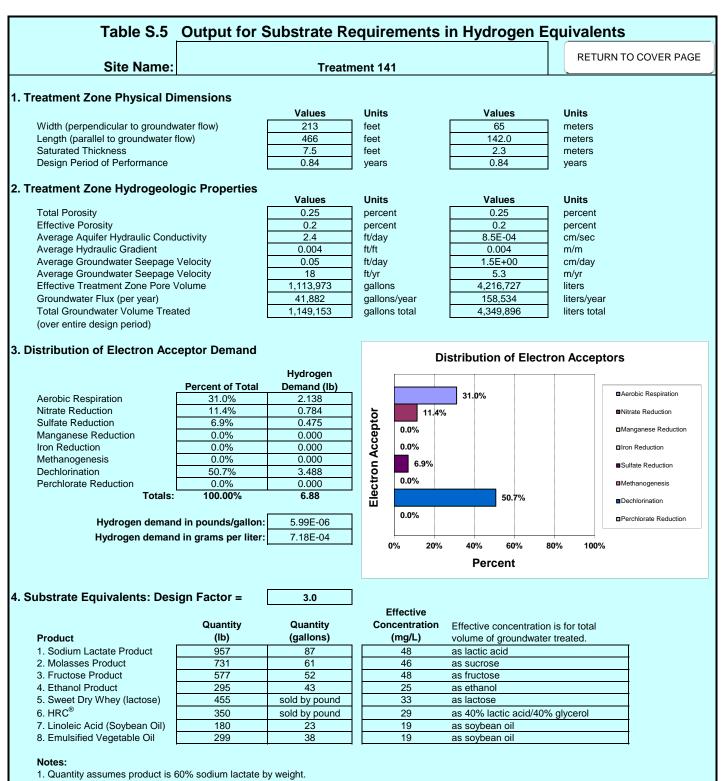
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.

3. Quantity assumes product is 80% fructose by weight and weights 11.2 pounds per gallon.

4. Quantity assumes product is 80% ethanol by weight and weights 6.9 pounds per gallon.

5. Quantity assumes product is 70% lactose by weight.

6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

# Treatment 141

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		87.02912294
2. Molasses Product		60.88220374
3. Fructose Product		51.51014045
4. Ethanol Product		42.75186718
5. Sweet Dry Whey (lactose)	455.0313608	
6. HRC®	349.7887273	•
7. Linoleic Acid (Soybean Oil)		23.02630371
8. Emulsified Vegetable Oil		38.37717285
9. Lactoil Product		18.42104297
10. Lactic Acid Product		0
11. Hydrogen Gas	6.88	•

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 143			RETURN TO COVER PAGE	
	r innut				
1. Treatment Zone Physical Dimensions	NOTE: Unshaded Values	Range	Units	User Notes	
Width (Perpendicular to predominant groundwater flow direction)	200	1-10,000	feet		
Length (Parallel to predominant groundwater flow)	520	1-1,000	feet		
Saturated Thickness	7.5	1-1,000	feet		
Treatment Zone Cross Sectional Area	1500		ft <sup>2</sup>		
Treatment Zone Volume	780.000		ft <sup>3</sup>		
Treatment Zone Total Pore Volume (total volume x total porosity)	1,458,990		gallons		
Treatment Zone Effective Pore Volume (total volume x total porosity)	1,167,192		gallons		
Design Period of Performance	2.9	.5 to 5	year		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
	5.0	2 10 20	unitess	Delaur = 3	
2. Treatment Zone Hydrogeologic Properties					
Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity	20%	.05-50	percent	Default = 20%	
Average Aquifer Hydraulic Conductivity	2.4	.01-1000	ft/day		
Average Hydraulic Gradient	0.004	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	0.05		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	17.5		ft/yr		
Average Groundwater Discharge through the Treatment Zone	39,325		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
3. Native Electron Acceptors					
A. Aqueous-Phase Native Electron Acceptors					
Oxygen	0.5	0.01 to 10	mg/L	Default = 5	
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1	
Sulfate	1	10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0	
B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

# 4. Contaminant Electron Acceptors

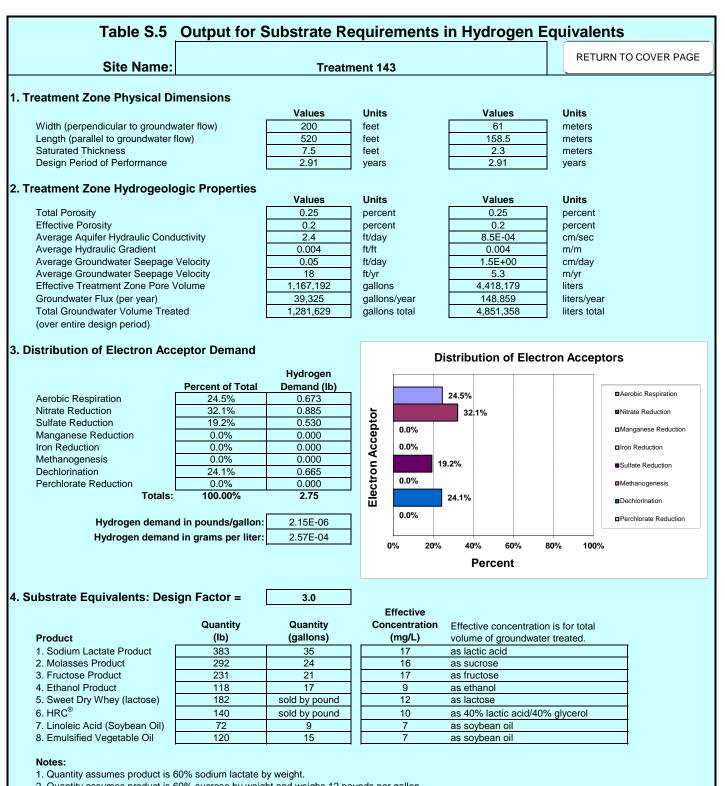
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.

3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.

4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.

5. Quantity assumes product is 70% lactose by weight.

6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

# Treatment 143

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		34.80191225
2. Molasses Product		24.34606992
3. Fructose Product		20.59829316
4. Ethanol Product		17.09596374
5. Sweet Dry Whey (lactose)	181.9616348	5
6. HRC®	139.8763561	
7. Linoleic Acid (Soybean Oil)		9.207945273
8. Emulsified Vegetable Oil		15.34657545
9. Lactoil Product		7.366356218
10. Lactic Acid Product		0
11. Hydrogen Gas	2.75	5

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 145			RETURN TO COVER PAGE
One Name.	haves are use	* input		
. Treatment Zone Physical Dimensions	NOTE: Unshaded Values	Range	Units	User Notes
Width (Perpendicular to predominant groundwater flow direction)	540	1-10,000	feet	
Length (Parallel to predominant groundwater flow)	75	1-1,000	feet	
Saturated Thickness	7	1-1,000	feet	
Treatment Zone Cross Sectional Area	3780		ft <sup>2</sup>	
Treatment Zone Volume	283.500		ft <sup>3</sup>	
Treatment Zone Total Pore Volume (total volume x total porosity)	530.287		gallons	
Treatment Zone Effective Pore Volume (total volume x total porosity)	424,229		gallons	
Design Period of Performance	0.4	.5 to 5	0	
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	year unitless	Default = 3
	5.0	2 10 20	unniess	Default = 5
. Treatment Zone Hydrogeologic Properties		-		
Total Porosity	25%	.05-50	percent	Default = 25%
Effective Porosity	20%	.05-50	percent	Default = 20%
Average Aquifer Hydraulic Conductivity	2.4	.01-1000	ft/day	
Average Hydraulic Gradient	0.004	0.0001-0.1	ft/ft	
Average Groundwater Seepage Velocity through the Treatment Zone	0.05		ft/day	
Average Groundwater Seepage Velocity through the Treatment Zone	17.5		ft/yr	
Average Groundwater Discharge through the Treatment Zone	99,100		gallons/year	
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%
. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors	Γ	1		
Oxygen	5.0	0.01 to 10	mg/L	Default = 5
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1
Sulfate	50	10 to 5,000	mg/L	Default = 50
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0
B. Solid-Phase Native Electron Acceptors				
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L	
Trichloroethene (TCE)	0.000	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.005	mg/L	
Vinyl Chloride (VC)	0.022	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)		mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

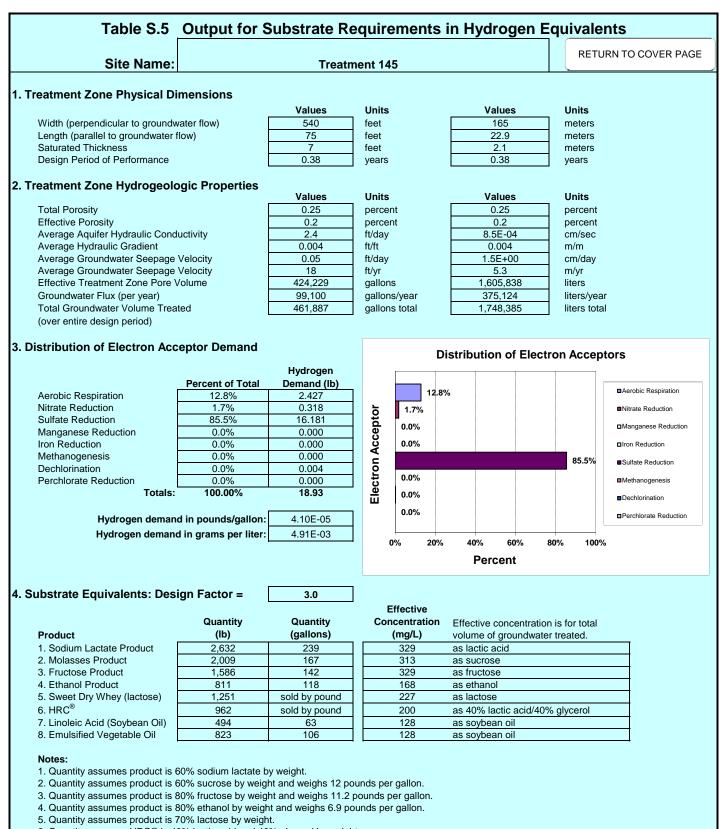
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 µs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

# Treatment 145

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		239.2979247
2. Molasses Product		167.40356
3. Fructose Product		141.6338496
4. Ethanol Product		117.5518349
5. Sweet Dry Whey (lactose)	1251.168076	
6. HRC®	961.7897286	
7. Linoleic Acid (Soybean Oil)		63.31382537
8. Emulsified Vegetable Oil		105.5230423
9. Lactoil Product		50.6510603
10. Lactic Acid Product		0
11. Hydrogen Gas	18.93	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 147			RETURN TO COVER PAGE
	NOTE: Unshaded	boxes are use	r input	
I. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes
Width (Perpendicular to predominant groundwater flow direction)	90	1-10,000	feet	
Length (Parallel to predominant groundwater flow)	75	1-1,000	feet	
Saturated Thickness	23	1-100	feet	
Treatment Zone Cross Sectional Area	2070		ft <sup>2</sup>	
Treatment Zone Volume	155,250		ft <sup>3</sup>	
Treatment Zone Total Pore Volume (total volume x total porosity)	290,395		gallons	
Treatment Zone Effective Pore Volume (total volume x effective porosity)	232,316		gallons	
Design Period of Performance	6.4	.5 to 5	year	
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3
2. Treatment Zone Hydrogeologic Properties				
Total Porosity	25%	.05-50	percent	Default = 25%
Effective Porosity	20%	.05-50	percent	Default = 20%
Average Aquifer Hydraulic Conductivity	10	.01-1000	ft/day	
Average Hydraulic Gradient	0.001	0.0001-0.1	ft/ft	
Average Groundwater Seepage Velocity through the Treatment Zone	0.05		ft/day	
Average Groundwater Seepage Velocity through the Treatment Zone	18.3		ft/yr	
Average Groundwater Discharge through the Treatment Zone	56,530		gallons/year	
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors				
Oxygen	2.2	0.01 to 10	mg/L	Default = 5
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1
Sulfate	336	10 to 5,000	mg/L	Default = 50
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0
B. Solid-Phase Native Electron Acceptors				
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L
Trichloroethene (TCE)	0.000	mg/L
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.361	mg/L
Vinyl Chloride (VC)	1.058	mg/L
Carbon Tetrachloride (CT)		mg/L
Trichloromethane ( or chloroform) (CF)		mg/L
Dichloromethane (or methylene chloride) (MC)		mg/L
Chloromethane		mg/L
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L
Dichloroethane (1,1-DCA and 1,2-DCA)	0.000	mg/L
Chloroethane		mg/L
Perchlorate		mg/L

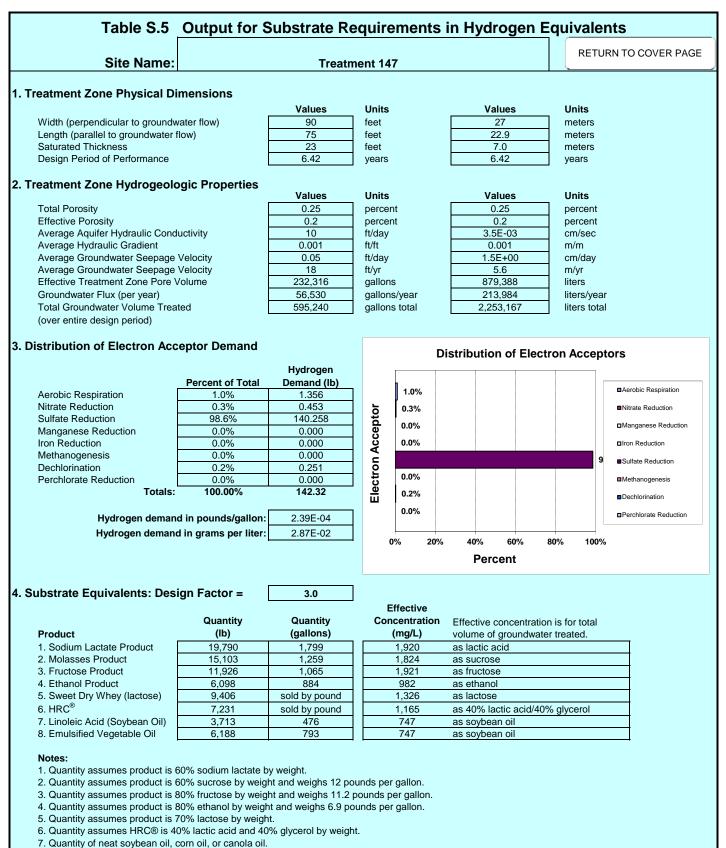
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



# Treatment 147

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		1799.067384
2. Molasses Product		1258.557863
3. Fructose Product		1064.818425
4. Ethanol Product		883.7672638
5. Sweet Dry Whey (lactose)	9406.415372	
6. HRC®	7230.838015	
7. Linoleic Acid (Soybean Oil)		476.0001087
8. Emulsified Vegetable Oil		793.3335145
9. Lactoil Product		380.800087
10. Lactic Acid Product		0
11. Hydrogen Gas	142.32	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 151			RETURN TO COVER PAGE		
NOTE: Unshaded boxes are user input.						
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	50	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	50	1-1,000	feet			
Saturated Thickness	16	1-100	feet			
Treatment Zone Cross Sectional Area	800		ft <sup>2</sup>			
Treatment Zone Volume	40,000		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	74,820		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	59,856		gallons			
Design Period of Performance	1.1	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
2. Treatment Zone Hydrogeologic Properties	r	1				
Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	4.7	.01-1000	ft/day			
Average Hydraulic Gradient	0.003	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.07		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	25.7		ft/yr			
Average Groundwater Discharge through the Treatment Zone	30,805		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors						
A. Aqueous-Phase Native Electron Acceptors						
Oxygen	5.0	0.01 to 10	mg/L	Default = 5		
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1		
Sulfate	32	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
· · · · · · · · · · · · · · · · · · ·	•					
B. Solid-Phase Native Electron Acceptors	t					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L
Trichloroethene (TCE)	3.032	mg/L
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.172	mg/L
Vinyl Chloride (VC)	0.000	mg/L
Carbon Tetrachloride (CT)		mg/L
Trichloromethane ( or chloroform) (CF)		mg/L
Dichloromethane (or methylene chloride) (MC)		mg/L
Chloromethane		mg/L
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L
Dichloroethane (1,1-DCA and 1,2-DCA)	0.000	mg/L
Chloroethane		mg/L
Perchlorate		mg/L

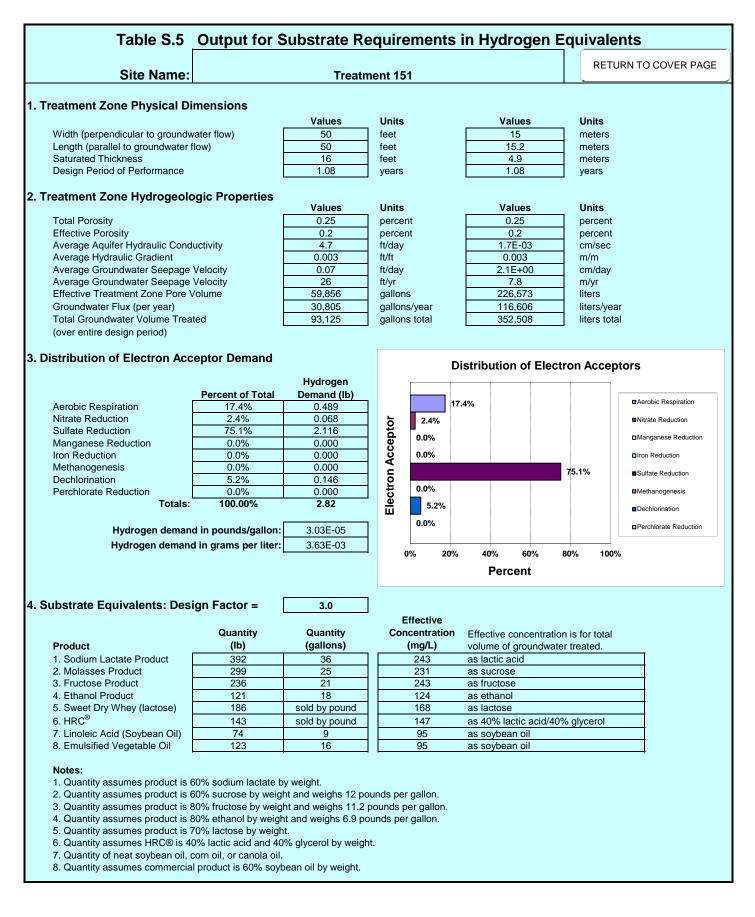
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 µs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



Treatment 151

-		
Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		35.64258716
2. Molasses Product		24.93417353
3. Fructose Product		21.09586548
4. Ethanol Product		17.5089338
5. Sweet Dry Whey (lactose)	186.3570997	
6. HRC®	143.2552091	
7. Linoleic Acid (Soybean Oil)		9.430372378
8. Emulsified Vegetable Oil		15.7172873
9. Lactoil Product		7.544297903
10. Lactic Acid Product		0
11. Hydrogen Gas	2.82	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1       Input for Substrate Requirements in Hydrogen Equivalents					
Site Name:	Treatment 152			RETURN TO COVER PAGE	
	NOTE: Unshaded	l boxes are use	r input.		
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes	
Width (Perpendicular to predominant groundwater flow direction)	50	1-10,000	feet		
Length (Parallel to predominant groundwater flow)	50	1-1,000	feet		
Saturated Thickness	33	1-100	feet		
Treatment Zone Cross Sectional Area	1650		ft <sup>2</sup>		
Treatment Zone Volume	82,500		ft <sup>3</sup>		
Treatment Zone Total Pore Volume (total volume x total porosity)	154,316		gallons		
Treatment Zone Effective Pore Volume (total volume x effective porosity)	123,453		gallons		
Design Period of Performance	1.5	.5 to 5	year		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
2. Treatment Zone Hydrogeologic Properties					
Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity	20%	.05-50	percent	Default = 20%	
Average Aquifer Hydraulic Conductivity	6.9	.01-1000	ft/day		
Average Hydraulic Gradient	0.003	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	0.10		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	37.8		ft/yr		
Average Groundwater Discharge through the Treatment Zone	93,275		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
3. Native Electron Acceptors <u>A. Aqueous-Phase Native Electron Acceptors</u>	1	1			
Oxygen	5.0	0.01 to 10	mg/L	Default = 5	
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1	
Sulfate	207	10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0	
B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L	
Trichloroethene (TCE)	31.596	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.005	mg/L	
Vinyl Chloride (VC)	0.000	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)	0.000	mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

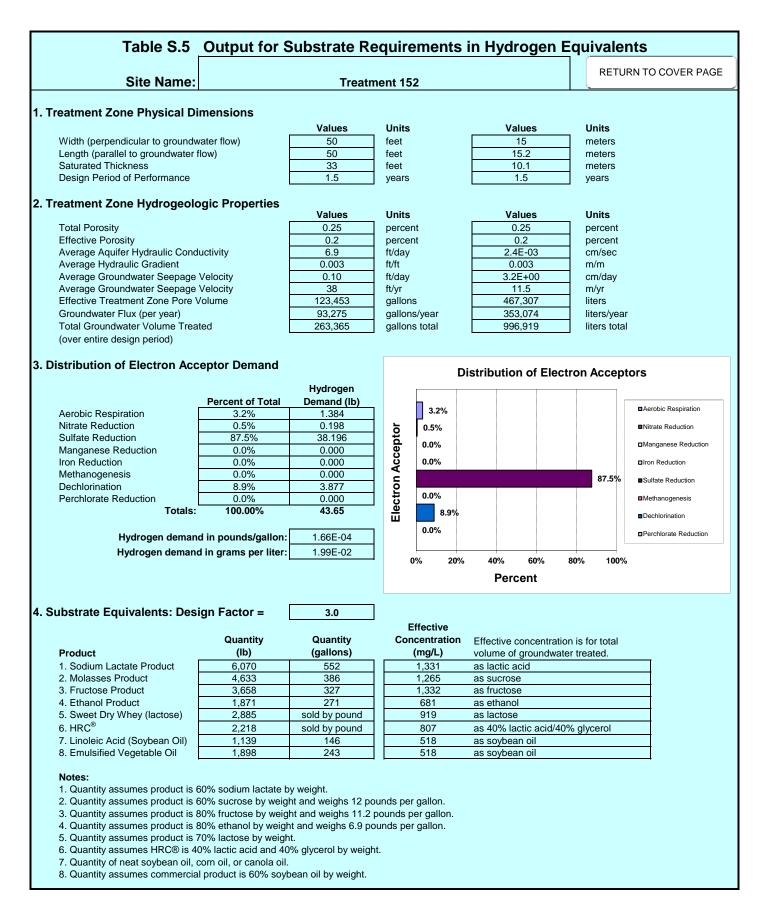
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



# Treatment 152

-		
Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		551.8519082
2. Molasses Product		386.0542216
3. Fructose Product		326.6259424
4. Ethanol Product		271.0897075
5. Sweet Dry Whey (lactose)	2885.355112	
6. HRC®	2218.011283	
7. Linoleic Acid (Soybean Oil)		146.0098553
8. Emulsified Vegetable Oil		243.3497588
9. Lactoil Product		116.8078842
10. Lactic Acid Product		0
11. Hydrogen Gas	43.65	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents						
Site Name:	Treatment 154			RETURN TO COVER PAGE		
	NOTE: Unshaded	l boxes are use	r input.			
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	50	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	50	1-1,000	feet			
Saturated Thickness	30	1-100	feet			
Treatment Zone Cross Sectional Area	1500		ft <sup>2</sup>			
Treatment Zone Volume	75,000		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	140,288		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	112,230		gallons			
Design Period of Performance	3.8	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
2. Treatment Zone Hydrogeologic Properties						
Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	6.9	.01-1000	ft/day			
Average Hydraulic Gradient	0.003	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.10		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	37.8		ft/yr			
Average Groundwater Discharge through the Treatment Zone	84,795		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors <u>A. Aqueous-Phase Native Electron Acceptors</u>	1	1				
Oxygen	5.0	0.01 to 10	mg/L	Default = 5		
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1		
Sulfate	50	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors						
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L	
Trichloroethene (TCE)	0.001	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.002	mg/L	
Vinyl Chloride (VC)	0.000	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)		mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

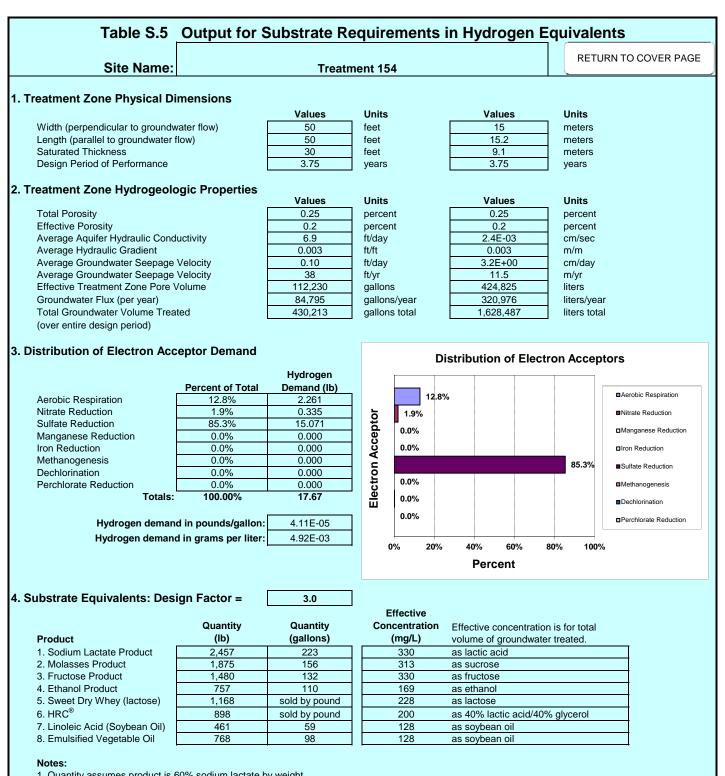
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 µs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

#### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as $CaCO_3$	Default = 10%



1. Quantity assumes product is 60% sodium lactate by weight.

2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.

3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.

4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.

5. Quantity assumes product is 70% lactose by weight.

6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

8. Quantity assumes commercial product is 60% soybean oil by weight.

# Treatment 154

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		223.3355198
2. Molasses Product		156.2368798
3. Fructose Product		132.1861419
4. Ethanol Product		109.7105217
5. Sweet Dry Whey (lactose)	1167.708717	,
6. HRC®	897.6333967	,
7. Linoleic Acid (Soybean Oil)		59.0904669
8. Emulsified Vegetable Oil		98.4841115
9. Lactoil Product		47.27237352
10. Lactic Acid Product		0
11. Hydrogen Gas	17.67	,

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 162			RETURN TO COVER PAGE		
	NOTE: Unshaded		r input.			
. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	40	1-10.000	feet			
Length (Parallel to predominant groundwater flow)	50	1-1.000	feet			
Saturated Thickness	30.3	1-100	feet			
Treatment Zone Cross Sectional Area	1212		ft <sup>2</sup>			
Treatment Zone Volume	60,600		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	113,352		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	90,682		gallons			
Design Period of Performance	5.8	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
. Treatment Zone Hydrogeologic Properties	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	3	.01-1000	ft/day			
Average Hydraulic Gradient	0.001	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.02		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	5.5		ft/yr			
Average Groundwater Discharge through the Treatment Zone	9,930		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors						
Oxygen	1.1	0.01 to 10	mg/L	Default = 5		
Nitrate	2.55	0.1 to- 20	mg/L	Default = 1		
Sulfate	50	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors						
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L
Trichloroethene (TCE)	0.103	mg/L
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	1.216	mg/L
Vinyl Chloride (VC)	0.791	mg/L
Carbon Tetrachloride (CT)		mg/L
Trichloromethane ( or chloroform) (CF)		mg/L
Dichloromethane (or methylene chloride) (MC)		mg/L
Chloromethane		mg/L
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L
Dichloroethane (1,1-DCA and 1,2-DCA)	0.001	mg/L
Chloroethane		mg/L
Perchlorate		mg/L

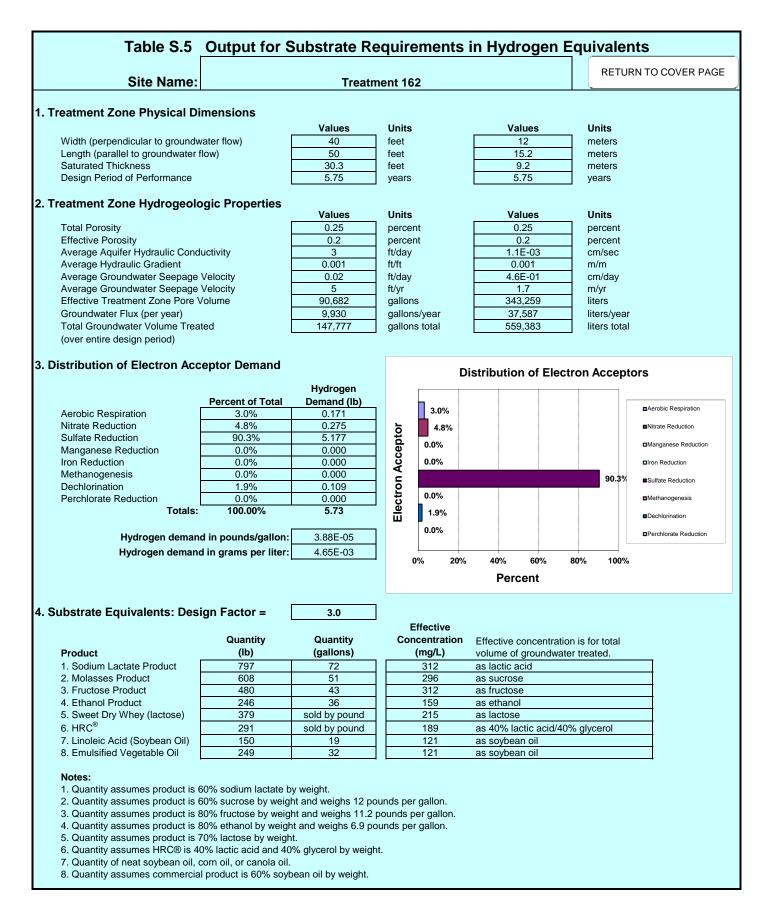
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

#### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



### Treatment 162

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		72.5
2. Molasses Product		50.7
3. Fructose Product		42.9
4. Ethanol Product		35.6
5. Sweet Dry Whey (lactose)	378.9	
6. HRC®	291.3	
7. Linoleic Acid (Soybean Oil)		19.2
8. Emulsified Vegetable Oil		32.0
9. Lactoil Product		15.3
10. Lactic Acid Product		47.0
11. Hydrogen Gas	5.7	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 174			RETURN TO COVER PAGE
	NOTE: Unshaded		r input	
. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes
Width (Perpendicular to predominant groundwater flow direction)	110	1-10,000	feet	
Length (Parallel to predominant groundwater flow)	55	1-1,000	feet	
Saturated Thickness	30.3	1-100	feet	
Treatment Zone Cross Sectional Area	3333		ft <sup>2</sup>	
Treatment Zone Volume	183,315		ft <sup>3</sup>	
Treatment Zone Total Pore Volume (total volume x total porosity)	342,891		gallons	
Treatment Zone Effective Pore Volume (total volume x effective porosity)	274,313		gallons	
Design Period of Performance	0.7	.5 to 5	year	
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3
Treatment Zone Hydrogeologic Properties	05%	05.50		Defectly 0504
Total Porosity	25%	.05-50	percent	Default = 25%
Effective Porosity	20%	.05-50	percent	Default = 20%
Average Aquifer Hydraulic Conductivity	3	.01-1000	ft/day	
Average Hydraulic Gradient	0.001	0.0001-0.1	ft/ft	
Average Groundwater Seepage Velocity through the Treatment Zone	0.02		ft/day	
Average Groundwater Seepage Velocity through the Treatment Zone	5.5		ft/yr	
Average Groundwater Discharge through the Treatment Zone Soil Bulk Density	27,307 1.7	1.4-2.0	gallons/year	Default = 1.7
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	Ŭ.	Default = $1.7$ Default = $0.05\%$
	0.05%	0.01-10	percent	Default = 0.05%
Native Electron Acceptors				
A. Aqueous-Phase Native Electron Acceptors	<b></b>	1		
Oxygen	0.2	0.01 to 10	mg/L	Default = 5
Nitrate	0.70	0.1 to- 20	mg/L	Default = 1
Sulfate	562	10 to 5,000	mg/L	Default = 50
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0
B. Solid-Phase Native Electron Acceptors				
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L
Trichloroethene (TCE)	0.000	mg/L
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.112	mg/L
Vinyl Chloride (VC)	0.350	mg/L
Carbon Tetrachloride (CT)		mg/L
Trichloromethane ( or chloroform) (CF)		mg/L
Dichloromethane (or methylene chloride) (MC)		mg/L
Chloromethane		mg/L
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L
Dichloroethane (1,1-DCA and 1,2-DCA)	0.001	mg/L
Chloroethane		mg/L
Perchlorate		mg/L

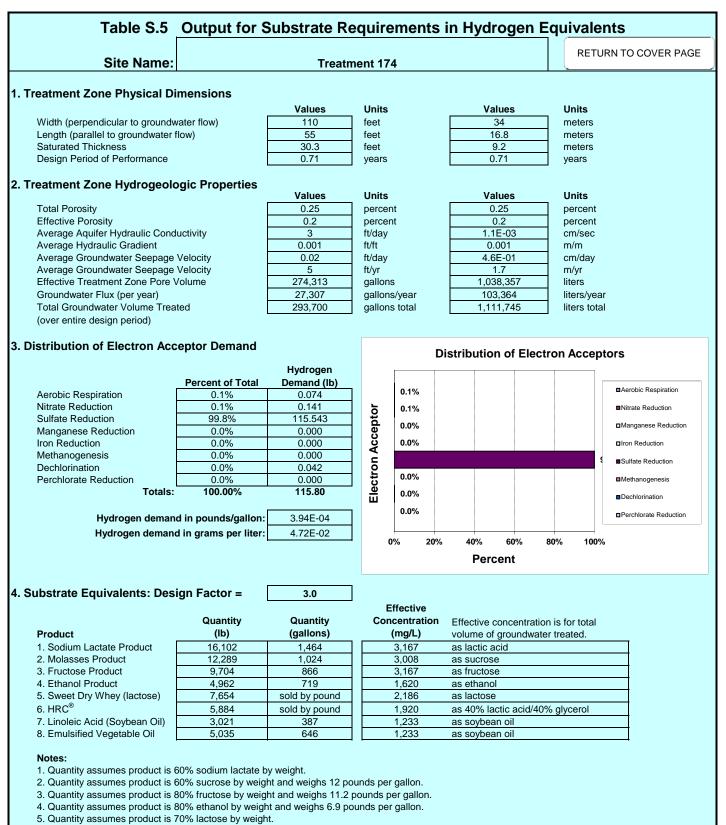
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

#### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

8. Quantity assumes commercial product is 60% soybean oil by weight.

### Treatment 174

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		1463.857672
2. Molasses Product		1024.058131
3. Fructose Product		866.4170305
4. Ethanol Product		719.1000741
5. Sweet Dry Whey (lactose)	7653.772964	
6. HRC®	5883.558223	
7. Linoleic Acid (Soybean Oil)		387.3097901
8. Emulsified Vegetable Oil		645.5163168
9. Lactoil Product		309.8478321
10. Lactic Acid Product		0
11. Hydrogen Gas	115.80	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 181			RETURN TO COVER PAGE		
	NOTE: Unshaded	boxes are use	r input.			
. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	180	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	200	1-1,000	feet			
Saturated Thickness	1	1-100	feet			
Treatment Zone Cross Sectional Area	180		ft <sup>2</sup>			
Treatment Zone Volume	36,000		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	67,338		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	53,870		gallons			
Design Period of Performance	1.0	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
Treatment Zone Hydrogeologic Properties  Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	25%	.05-50		Default = 20%		
Average Aquifer Hydraulic Conductivity	20%	.05-50	percent ft/day	Default = 20%		
Average Hydraulic Gradient	0.054	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.62	0.0001-0.1	ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	226.7		ft/yr			
Average Groundwater Discharge through the Treatment Zone	61,053		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors						
Oxygen	5.4	0.01 to 10	mg/L	Default = 5		
Nitrate Sulfate	1.00 50	0.1 to- 20 10 to 5,000	mg/L mg/L	Default = 1 Default = 50		
	0.0	0.1 to 20	ů.			
Carbon Dioxide (estimated as the amount of Methane produced) B. Solid-Phase Native Electron Acceptors	0.0	0.1 to 20	mg/L	Default = 0		
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

# 4. Contaminant Electron Acceptors

0.060	mg/L	
0.048	mg/L	
0.007	mg/L	
0.011	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.000	mg/L	
0.000	mg/L	
	mg/L	
	mg/L	
	0.048 0.007 0.011 0.000	0.048          mg/L           0.007          mg/L           0.011          mg/L            mg/L             mg/L             mg/L             mg/L             mg/L            0.000          mg/L           0.000          mg/L           0.000          mg/L           0.000          mg/L

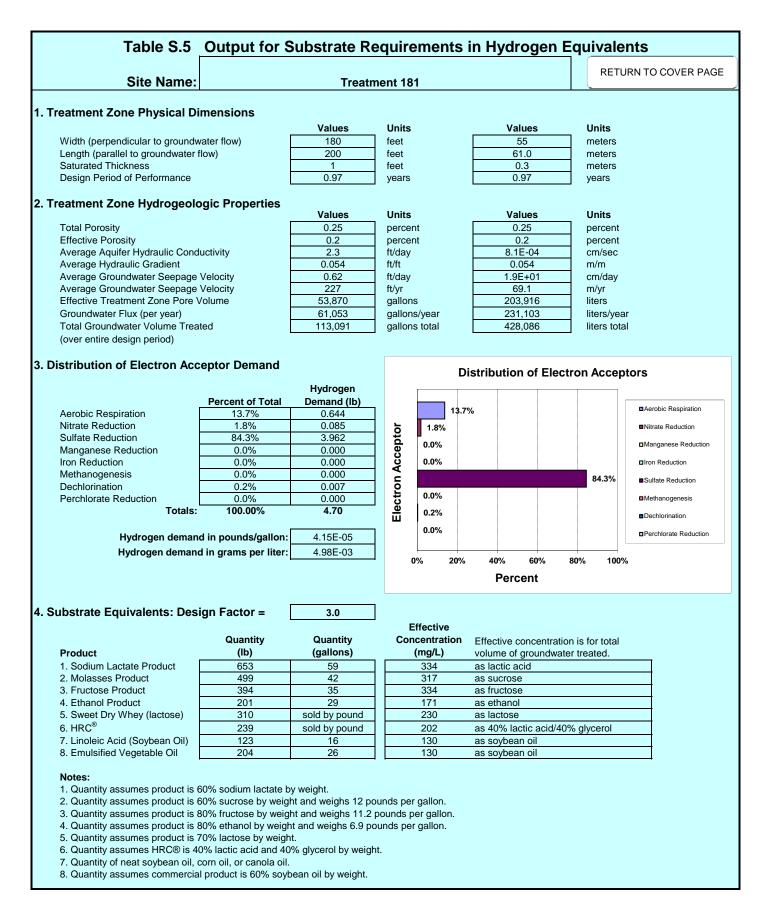
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

#### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



# Treatment 181

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		59.38248016
2. Molasses Product		41.54168321
3. Fructose Product		35.14685418
4. Ethanol Product		29.17083177
5. Sweet Dry Whey (lactose)	310.4810187	7
6. HRC®	238.6709352	2
7. Linoleic Acid (Soybean Oil)		15.71151101
8. Emulsified Vegetable Oil		26.18585168
9. Lactoil Product		12.56920881
10. Lactic Acid Product		0
11. Hydrogen Gas	4.70	)

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 400			RETURN TO COVER PAGE		
Site Name:	Treatment 182	_		KETOKA TO COVER TAGE		
NOTE: Unshaded boxes are user input.						
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	80	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	280	1-1,000	feet			
Saturated Thickness	8.5	1-100	feet			
Treatment Zone Cross Sectional Area	680		ft <sup>2</sup>			
Treatment Zone Volume	190,400		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	356,143		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	284,915		gallons			
Design Period of Performance	0.4	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
2. Treatment Zone Hydrogeologic Properties						
Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	50	.01-1000	ft/day			
Average Hydraulic Gradient	0.011	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	2.75		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	1003.8		ft/yr			
Average Groundwater Discharge through the Treatment Zone	1,021,368		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors <u>A. Aqueous-Phase Native Electron Acceptors</u>						
Oxygen	0.5	0.01 to 10	mg/L	Default = 5		
Nitrate	0.00	0.1 to- 20	mg/L	Default = 1		
Sulfate	40	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors						
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

# 4. Contaminant Electron Acceptors

0.014	mg/L	
0.044	mg/L	
0.968	mg/L	
0.174	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.000	mg/L	
0.000	mg/L	
	mg/L	
	mg/L	
	0.044 0.968 0.174 0.000	0.044          mg/L           0.968          mg/L           0.174          mg/L            mg/L             mg/L             mg/L             mg/L             mg/L            0.000          mg/L           0.000          mg/L           0.000          mg/L           0.000          mg/L

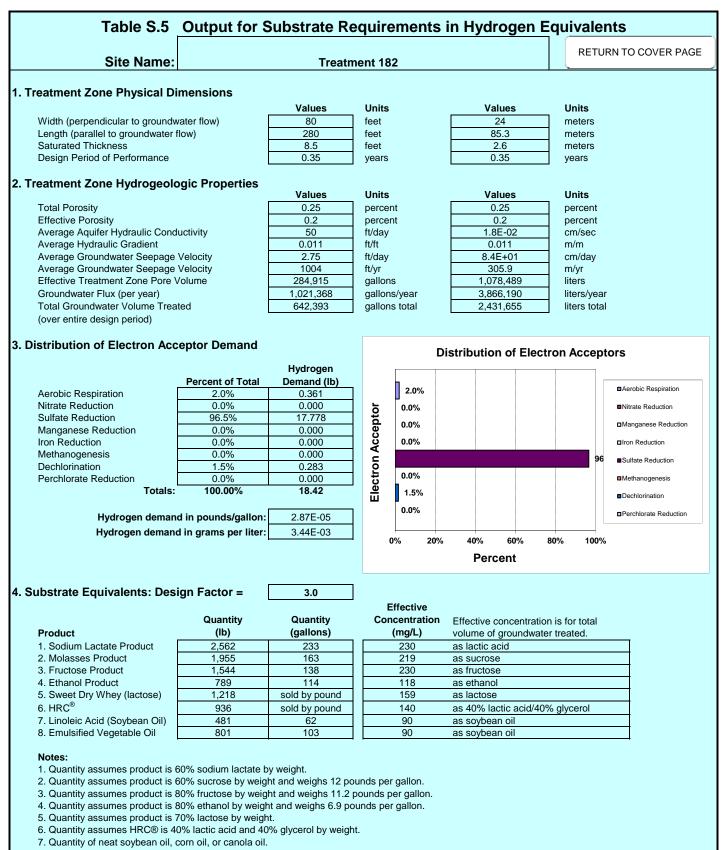
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

#### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



# Treatment 182

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		232.8799049
2. Molasses Product		162.9137619
3. Fructose Product		137.8352005
4. Ethanol Product		114.3990704
5. Sweet Dry Whey (lactose)	1217.611489	)
6. HRC®	935.9943291	
7. Linoleic Acid (Soybean Oil)		61.61573547
8. Emulsified Vegetable Oil		102.6928925
9. Lactoil Product		49.29258838
10. Lactic Acid Product		0
11. Hydrogen Gas	18.42	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 195			RETURN TO COVER PAGE		
NOTE: Unshaded boxes are user input.						
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	150	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	180	1-1,000	feet			
Saturated Thickness	12	1-100	feet			
Treatment Zone Cross Sectional Area	1800		ft <sup>2</sup>			
Treatment Zone Volume	324,000		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	606,042		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	484,834		gallons			
Design Period of Performance	0.4	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
2. Treatment Zone Hydrogeologic Properties						
Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	72	.01-1000	ft/day			
Average Hydraulic Gradient	0.007	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	2.52		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	919.8		ft/yr			
Average Groundwater Discharge through the Treatment Zone	2,477,500		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors						
Oxygen	5.0	0.01 to 10	mg/L	Default = 5		
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1		
Sulfate	25	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors						
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L	
Trichloroethene (TCE)	3.770	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.662	mg/L	
Vinyl Chloride (VC)	0.034	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)		mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

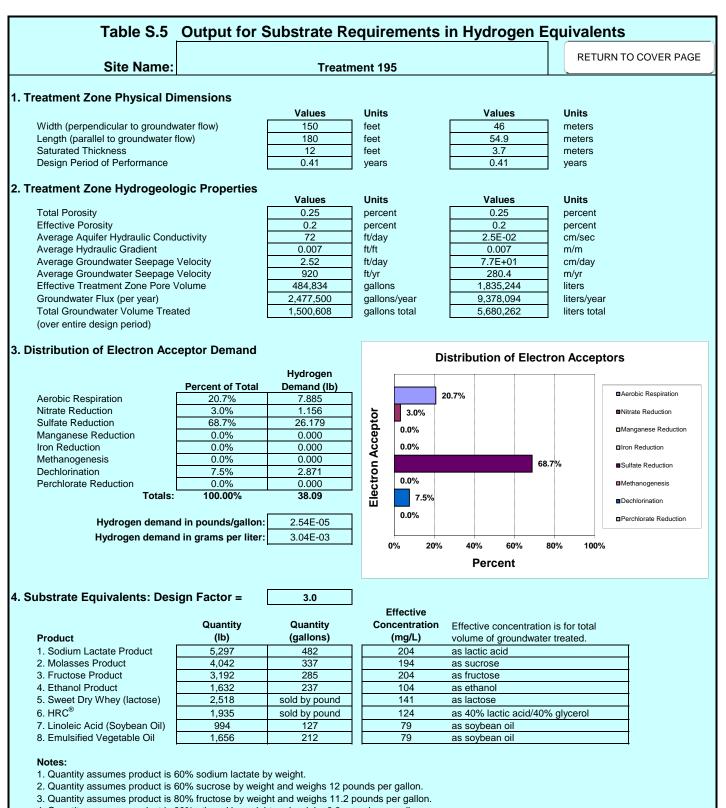
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

#### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.

5. Quantity assumes product is 70% lactose by weight.

6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

8. Quantity assumes commercial product is 60% soybean oil by weight.

# Treatment 195

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		481.5284322
2. Molasses Product		336.8586414
3. Fructose Product		285.0034142
4. Ethanol Product		236.5442611
5. Sweet Dry Whey (lactose)	2517.669148	
6. HRC®	1935.366137	
7. Linoleic Acid (Soybean Oil)		127.4035582
8. Emulsified Vegetable Oil		212.3392636
9. Lactoil Product		101.9228465
10. Lactic Acid Product		0
11. Hydrogen Gas	38.09	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents					
Site Name:	Treatment 197			RETURN TO COVER PAGE	
	NOTE: Unshaded	boxes are use	r input.		
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes	
Width (Perpendicular to predominant groundwater flow direction)	164	1-10,000	feet		
Length (Parallel to predominant groundwater flow)	750	1-1,000	feet		
Saturated Thickness	12	1-100	feet		
Treatment Zone Cross Sectional Area	1968		ft <sup>2</sup>		
Treatment Zone Volume	1,476,000		ft <sup>3</sup>		
Treatment Zone Total Pore Volume (total volume x total porosity)	2,760,858		gallons		
Treatment Zone Effective Pore Volume (total volume x effective porosity)	2,208,686		gallons		
Design Period of Performance	2.6	.5 to 5	year		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
2. Treatment Zone Hydrogeologic Properties					
Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity	20%	.05-50	percent	Default = 20%	
Average Aquifer Hydraulic Conductivity	72	.01-1000	ft/day		
Average Hydraulic Gradient	0.007	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	2.52		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	919.8		ft/yr		
Average Groundwater Discharge through the Treatment Zone	2,708,733		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors					
Oxygen	5.0	0.01 to 10	mg/L	Default = 5	
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1	
Sulfate	35	10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0	
B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

# 4. Contaminant Electron Acceptors

0.000	mg/L	
0.907	mg/L	
1.086	mg/L	
0.019	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	0.907 1.086	0.907          mg/L           1.086          mg/L           0.019          mg/L            mg/L             mg/L             mg/L             mg/L             mg/L             mg/L             mg/L             mg/L             mg/L             mg/L             mg/L             mg/L             mg/L             mg/L

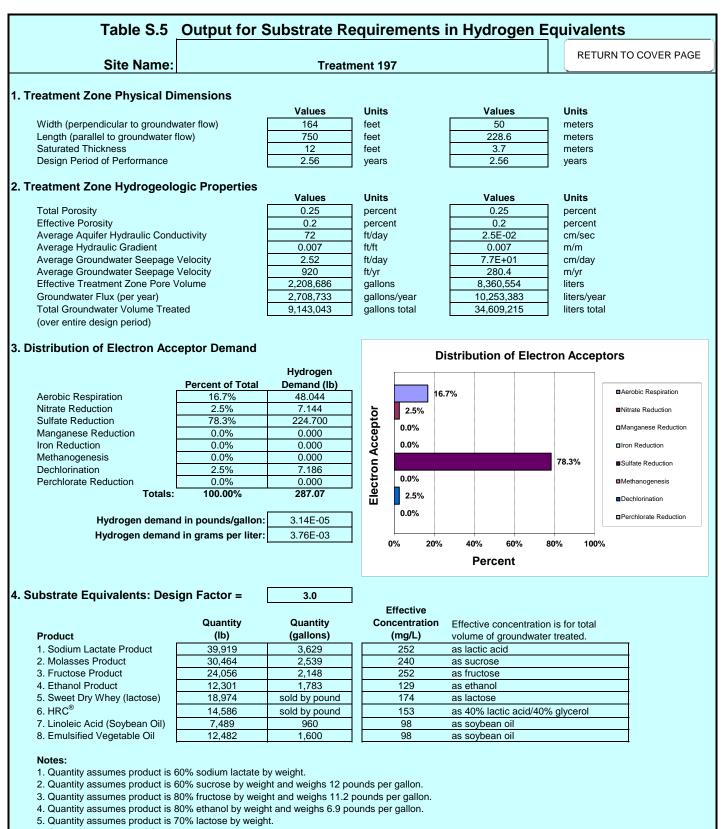
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

#### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

8. Quantity assumes commercial product is 60% soybean oil by weight.

Treatment 197

978
706
244
306
939
323
751
0
22 30 32

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 220			RETURN TO COVER PAGE		
NOTE: Unshaded boxes are user input.						
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	120	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	210	1-1,000	feet			
Saturated Thickness	12	1-100	feet			
Treatment Zone Cross Sectional Area	1440		ft <sup>2</sup>			
Treatment Zone Volume	302,400		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	565,639		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	452,511		gallons			
Design Period of Performance	1.9	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
2. Treatment Zone Hydrogeologic Properties						
Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	10	.01-1000	ft/day			
Average Hydraulic Gradient	0.007	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.35		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	127.8		ft/yr			
Average Groundwater Discharge through the Treatment Zone	275,278		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors	Г	1				
Oxygen	0.6	0.01 to 10	mg/L	Default = 5		
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1		
Sulfate	403	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors						
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

# 4. Contaminant Electron Acceptors

0.905	mg/L	
0.103	mg/L	
0.082	mg/L	
0.007	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	0.103 0.082	0.103      mg/L       0.082      mg/L       0.007      mg/L        mg/L

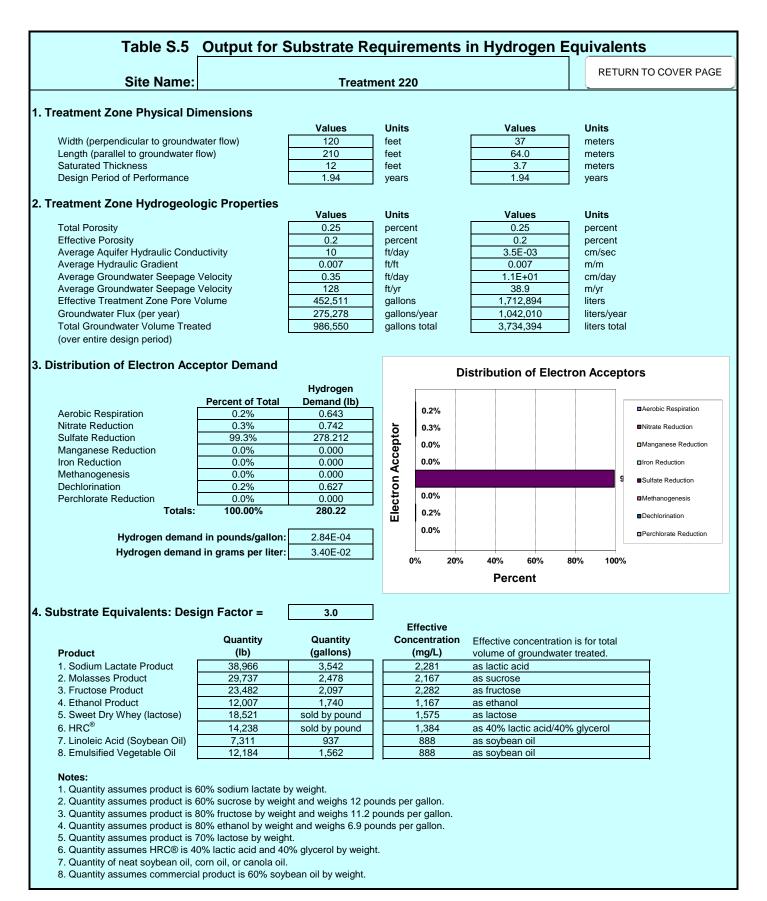
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 µs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

#### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



### Treatment 220

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		3542.375818
2. Molasses Product		2478.108924
3. Fructose Product		2096.634663
4. Ethanol Product		1740.143705
5. Sweet Dry Whey (lactose)	18521.29533	
6. HRC®	14237.56884	
7. Linoleic Acid (Soybean Oil)		937.2474258
8. Emulsified Vegetable Oil		1562.079043
9. Lactoil Product		749.7979406
10. Lactic Acid Product		2296.048482
11. Hydrogen Gas	280.22	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents						
Site Name:	Treatment 221			RETURN TO COVER PAGE		
	NOTE: Unshaded	boxes are use	r input.			
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	90	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	170	1-1,000	feet			
Saturated Thickness	12	1-100	feet			
Treatment Zone Cross Sectional Area	1080		ft <sup>2</sup>			
Treatment Zone Volume	183,600		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	343,424		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	274,739		gallons			
Design Period of Performance	1.6	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
2. Treatment Zone Hydrogeologic Properties						
Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	10	.01-1000	ft/day			
Average Hydraulic Gradient	0.007	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.35		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	127.8		ft/yr			
Average Groundwater Discharge through the Treatment Zone	206,458		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors						
Oxygen	5.0	0.01 to 10	mg/L	Default = 5		
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1		
Sulfate	361	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors						
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	6.450	mg/L
Trichloroethene (TCE)	2.560	mg/L
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	8.047	mg/L
Vinyl Chloride (VC)	0.039	mg/L
Carbon Tetrachloride (CT)		mg/L
Trichloromethane ( or chloroform) (CF)		mg/L
Dichloromethane (or methylene chloride) (MC)		mg/L
Chloromethane		mg/L
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L
Dichloroethane (1,1-DCA and 1,2-DCA)		mg/L
Chloroethane		mg/L
Perchlorate		mg/L

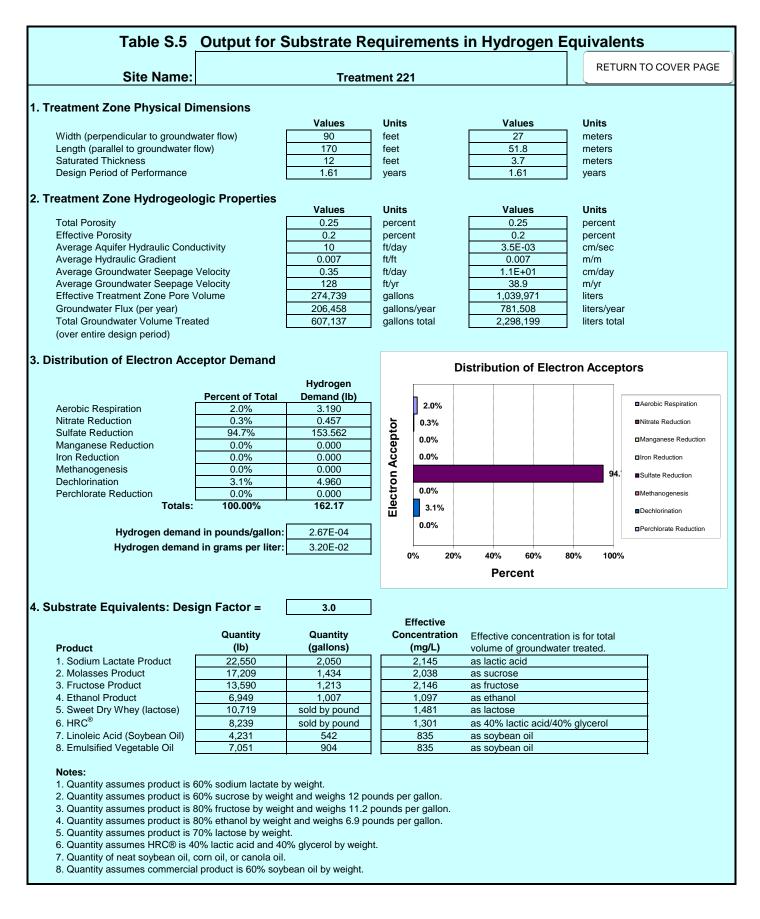
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

#### **B. Aquifer Matrix**

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as $CaCO_3$	Default = 10%



Treatment 221

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		2050.025539
2. Molasses Product		1434.118469
3. Fructose Product		1213.353644
4. Ethanol Product		1007.047028
5. Sweet Dry Whey (lactose)	10718.54891	
6. HRC®	8239.492714	
7. Linoleic Acid (Soybean Oil)		542.3990165
8. Emulsified Vegetable Oil		903.9983608
9. Lactoil Product		433.9192132
10. Lactic Acid Product		1328.757385
11. Hydrogen Gas	162.17	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 223			RETURN TO COVER PAGE	
NOTE: Unshaded boxes are user input.					
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes	
Width (Perpendicular to predominant groundwater flow direction)	120	1-10,000	feet		
Length (Parallel to predominant groundwater flow)	210	1-1,000	feet		
Saturated Thickness	12	1-100	feet		
Treatment Zone Cross Sectional Area	1440		ft <sup>2</sup>		
Treatment Zone Volume	302,400		ft <sup>3</sup>		
Treatment Zone Total Pore Volume (total volume x total porosity)	565,639		gallons		
Treatment Zone Effective Pore Volume (total volume x effective porosity)	452,511		gallons		
Design Period of Performance	0.0	.5 to 5	year		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
2. Treatment Zone Hydrogeologic Properties	25%	05 50		Default 25%	
Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity	20% 10	.05-50	percent ft/dov/	Default = 20%	
Average Aquifer Hydraulic Conductivity	0.007	.01-1000	ft/day ft/ft		
Average Hydraulic Gradient Average Groundwater Seepage Velocity through the Treatment Zone	0.35	0.0001-0.1	ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	127.8		ft/yr		
Average Groundwater Discharge through the Treatment Zone	275,278		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
	0.0378	0.01-10	percent	Default = 0.0370	
3. Native Electron Acceptors					
A. Aqueous-Phase Native Electron Acceptors	r	1			
Oxygen	0.6	0.01 to 10	mg/L	Default = 5	
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1	
Sulfate	476	10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0	
B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

### 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)         165.011          mg/L           Trichloroethene (TCE)         0.457          mg/L           Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)         0.456          mg/L           Vinyl Chloride (VC)         0.000          mg/L           Carbon Tetrachloride (CT)          mg/L            Trichloromethane ( or chloroform) (CF)          mg/L            Dichloromethane (or methylene chloride) (MC)          mg/L            Chloromethane          mg/L          mg/L           Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)          mg/L          mg/L           Dichloroethane (1,1,1-TCA and 1,1,2-TCA)          mg/L          mg/L           Dichloroethane (1,1,1-DCA and 1,2-DCA)          mg/L          mg/L           Dichloroethane          mg/L          mg/L            Dichloroethane (1,1,1-DCA and 1,2-DCA)          mg/L          mg/L           Dichloroethane          mg/L          mg/L            Dichlor			
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)0.456mg/LVinyl Chloride (VC)0.000mg/LCarbon Tetrachloride (CT)mg/LTrichloromethane ( or chloroform) (CF)mg/LDichloromethane (or methylene chloride) (MC)mg/LChloromethanemg/LTetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)mg/LTrichloroethane (1,1,1-TCA and 1,1,2-TCA)mg/LDichloroethane (1,1-DCA and 1,2-DCA)mg/LChloroethanemg/LChloroethanemg/LDichloroethanemg/LDichloroethanemg/LDichloroethanemg/LDichloroethanemg/LDichloroethanemg/LDichloroethanemg/LDichloroethanemg/LDichloroethanemg/LChloroethanemg/LChloroethanemg/LChloroethanemg/LChloroethanemg/LChloroethanemg/LChloroethanemg/LChloroethanemg/L	Tetrachloroethene (PCE)	165.011	mg/L
Vinyl Chloride (VC)0.000mg/LCarbon Tetrachloride (CT)mg/LTrichloromethane ( or chloroform) (CF)mg/LDichloromethane (or methylene chloride) (MC)mg/LChloromethanemg/LChloromethanemg/LTetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)mg/LTrichloroethane (1,1,1-TCA and 1,1,2-TCA)mg/LDichloroethane (1,1-DCA and 1,2-DCA)mg/LChloroethanemg/LChloroethanemg/LDichloroethanemg/LDichloroethanemg/LDichloroethanemg/L	Trichloroethene (TCE)	0.457	mg/L
Carbon Tetrachloride (CT)mg/LTrichloromethane ( or chloroform) (CF)mg/LDichloromethane (or methylene chloride) (MC)mg/LChloromethanemg/LTetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)mg/LTrichloroethane (1,1,1-TCA and 1,1,2-TCA)mg/LDichloroethane (1,1-DCA and 1,2-DCA)mg/LChloroethanemg/LDichloroethanemg/LDichloroethanemg/L	Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.456	mg/L
Trichloromethane ( or chloroform) (CF)mg/LDichloromethane (or methylene chloride) (MC)mg/LChloromethanemg/LTetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)mg/LTrichloroethane (1,1,1-TCA and 1,1,2-TCA)mg/LDichloroethane (1,1-DCA and 1,2-DCA)mg/LChloroethane (1,1-DCA and 1,2-DCA)mg/LDichloroethane (1,1-DCA and 1,2-DCA)mg/LOthoroethane g/L	Vinyl Chloride (VC)	0.000	mg/L
Dichloromethane (or methylene chloride) (MC)mg/LChloromethanemg/LTetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)mg/LTrichloroethane (1,1,1-TCA and 1,1,2-TCA)mg/LDichloroethane (1,1-DCA and 1,2-DCA)mg/LChloroethane (1,1-DCA and 1,2-DCA)mg/LDichloroethane (1,1-DCA and 1,2-DCA)mg/LOthoroethanemg/L	Carbon Tetrachloride (CT)		mg/L
Chloromethanemg/LTetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)mg/LTrichloroethane (1,1,1-TCA and 1,1,2-TCA)mg/LDichloroethane (1,1-DCA and 1,2-DCA)mg/LChloroethanemg/L	Trichloromethane ( or chloroform) (CF)		mg/L
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)        mg/L         Trichloroethane (1,1,1-TCA and 1,1,2-TCA)        mg/L         Dichloroethane (1,1-DCA and 1,2-DCA)        mg/L         Chloroethane        mg/L	Dichloromethane (or methylene chloride) (MC)		mg/L
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)      mg/L       Dichloroethane (1,1-DCA and 1,2-DCA)      mg/L       Chloroethane      mg/L	Chloromethane		mg/L
Dichloroethane (1,1-DCA and 1,2-DCA)      mg/L       Chloroethane      mg/L	Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L
Chloroethane mg/L	Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L
	Dichloroethane (1,1-DCA and 1,2-DCA)		mg/L
	Chloroethane		mg/L
Perchlorate mg/L	Perchlorate		mg/L

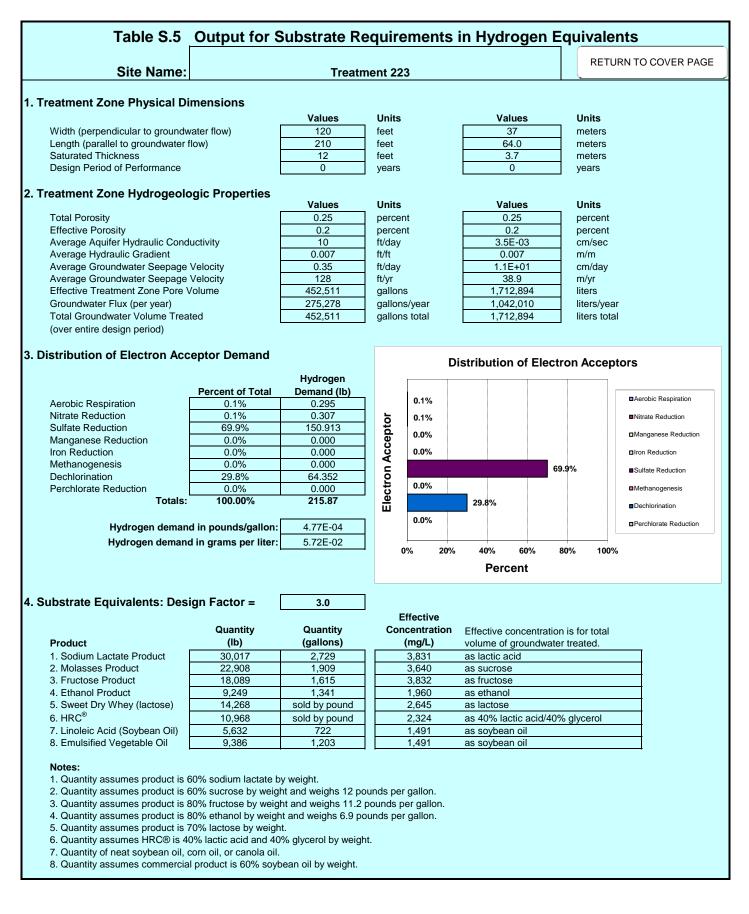
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 µs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



### Treatment 223

-		
Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		2728.830464
2. Molasses Product		1908.984104
3. Fructose Product		1615.119579
4. Ethanol Product		1340.500669
5. Sweet Dry Whey (lactose)	14267.67727	
6. HRC®	10967.75542	
7. Linoleic Acid (Soybean Oil)		721.9983025
8. Emulsified Vegetable Oil		1203.330504
9. Lactoil Product		577.598642
10. Lactic Acid Product		0
11. Hydrogen Gas	215.87	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 224			RETURN TO COVER PAGE	
Olte Name.	NOTE: Unshaded	hoves are use	ripput		
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes	
· · · · · · · · · · · · · · · · · · ·	1	1			
Width (Perpendicular to predominant groundwater flow direction)	75 30	1-10,000	feet		
Length (Parallel to predominant groundwater flow) Saturated Thickness		1-1,000 1-100	feet		
	900		feet ft <sup>2</sup>		
Treatment Zone Cross Sectional Area	27.000		ft <sup>3</sup>		
Treatment Zone Volume Treatment Zone Total Pore Volume (total volume x total porosity)	50,504		gallons		
· · · · · · · · · · · · · · · · · · ·			0		
Treatment Zone Effective Pore Volume (total volume x effective porosity)	40,403		gallons		
Design Period of Performance	1.6 3.0	.5 to 5	year	Default = 3	
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
2. Treatment Zone Hydrogeologic Properties					
Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity	20%	.05-50	percent	Default = 20%	
Average Aquifer Hydraulic Conductivity	10	.01-1000	ft/day		
Average Hydraulic Gradient	0.007	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	0.35		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	127.8		ft/yr		
Average Groundwater Discharge through the Treatment Zone	172,049		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors					
Oxygen		0.01 to 10	mg/L	Default = 5	
Nitrate		0.1 to- 20	mg/L	Default = 1	
Sulfate		10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced)		0.1 to 20	mg/L	Default = 0	
B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

### 4. Contaminant Electron Acceptors

mg/L
mg/L
mg/L
mg/L
mg/L
mg/L
mg/L
mg/L
mg/L
mg/L
mg/L
mg/L
mg/L

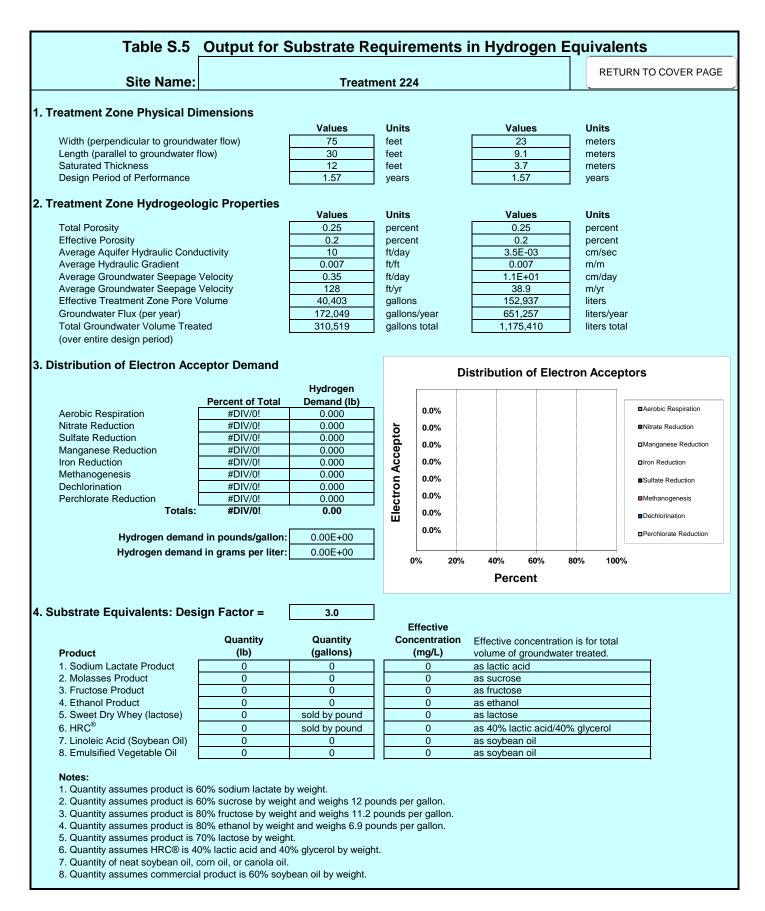
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



### Treatment 224

Substrate	Quantity(lb)	Quantity (gallons)	
1. Sodium Lactate Product		C	)
2. Molasses Product		C	)
3. Fructose Product		C	)
4. Ethanol Product		C	)
5. Sweet Dry Whey (lactose)		0	
6. HRC®		0	
7. Linoleic Acid (Soybean Oil)		C	)
8. Emulsified Vegetable Oil		C	)
9. Lactoil Product		C	)
10. Lactic Acid Product		C	)
11. Hydrogen Gas	(	0.00	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 229			RETURN TO COVER PAGE	
	NOTE: Unshaded	boxes are use	r input		
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes	
Width (Perpendicular to predominant groundwater flow direction)	170	1-10,000	feet		
Length (Parallel to predominant groundwater flow)	340	1-1,000	feet		
Saturated Thickness	12	1-100	feet		
Treatment Zone Cross Sectional Area	2040		ft <sup>2</sup>		
Treatment Zone Volume	693,600		ft <sup>3</sup>		
Treatment Zone Total Pore Volume (total volume x total porosity)	1,297,379		gallons		
Treatment Zone Effective Pore Volume (total volume x effective porosity)	1,037,903		gallons		
Design Period of Performance	2.0	.5 to 5	year		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
2. Treatment Zone Hydrogeologic Properties					
Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity	20%	.05-50	percent	Default = 20%	
Average Aquifer Hydraulic Conductivity	10	.01-1000	ft/day		
Average Hydraulic Gradient	0.007	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	0.35		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	127.8		ft/yr		
Average Groundwater Discharge through the Treatment Zone	389,977		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors					
Oxygen	5.0	0.01 to 10	mg/L	Default = 5	
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1	
Sulfate	215	10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0	
B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

### 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.157	mg/L
Trichloroethene (TCE)	0.015	mg/L
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.486	mg/L
Vinyl Chloride (VC)	0.227	mg/L
Carbon Tetrachloride (CT)		mg/L
Trichloromethane ( or chloroform) (CF)		mg/L
Dichloromethane (or methylene chloride) (MC)		mg/L
Chloromethane		mg/L
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L
Dichloroethane (1,1-DCA and 1,2-DCA)	0.000	mg/L
Chloroethane		mg/L
Perchlorate		mg/L

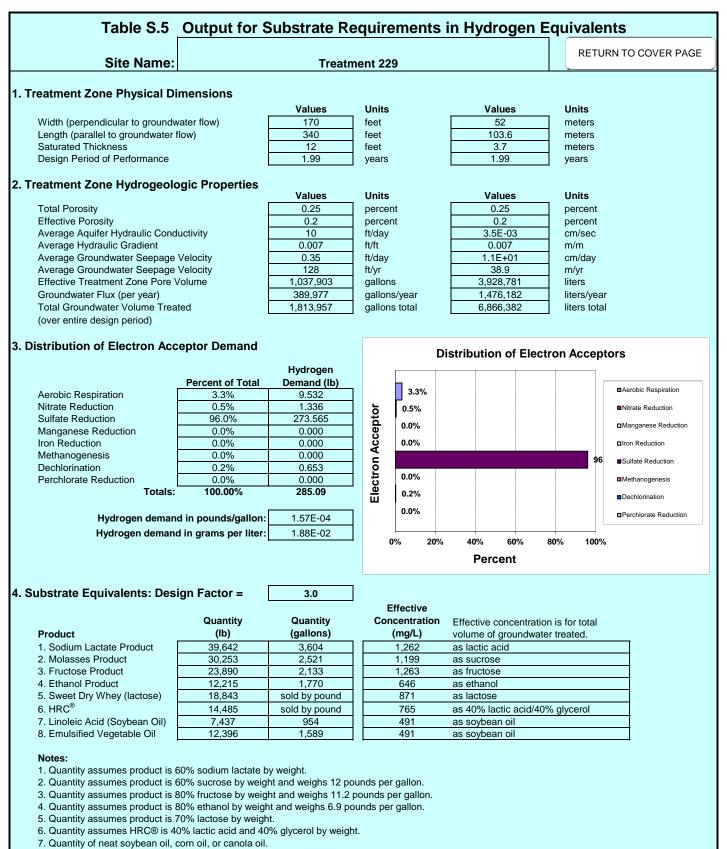
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 µs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



### Treatment 229

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		3603.840667
2. Molasses Product		2521.107352
3. Fructose Product		2133.014014
4. Ethanol Product		1770.337472
5. Sweet Dry Whey (lactose)	18842.66401	
6. HRC®	14484.60926	i
7. Linoleic Acid (Soybean Oil)		953.5098932
8. Emulsified Vegetable Oil		1589.183155
9. Lactoil Product		762.8079146
10. Lactic Acid Product		2335.887924
11. Hydrogen Gas	285.09	)

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

		1 (			
Site Name:	Treatment 230			RETURN TO COVER PAGE	
NOTE: Unshaded boxes are user input.					
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes	
Width (Perpendicular to predominant groundwater flow direction)	75	1-10,000	feet		
Length (Parallel to predominant groundwater flow)	30	1-1,000	feet		
Saturated Thickness	12	1-100	feet		
Treatment Zone Cross Sectional Area	900		ft <sup>2</sup>		
Treatment Zone Volume	27,000		ft <sup>3</sup>		
Treatment Zone Total Pore Volume (total volume x total porosity)	50,504		gallons		
Treatment Zone Effective Pore Volume (total volume x effective porosity)	40,403		gallons		
Design Period of Performance	0.8	.5 to 5	year		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
2. Treatment Zone Hydrogeologic Properties					
Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity	20%	.05-50	percent	Default = 20%	
Average Aquifer Hydraulic Conductivity	10	.01-1000	ft/day		
Average Hydraulic Gradient	0.007	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	0.35		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	127.8		ft/yr		
Average Groundwater Discharge through the Treatment Zone	172,049		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
3. Native Electron Acceptors <u>A. Aqueous-Phase Native Electron Acceptors</u>					
Oxygen	5.0	0.01 to 10	mg/L	Default = 5	
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1	
Sulfate	174	10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0	
B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

### 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L
Trichloroethene (TCE)	0.000	mg/L
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.060	mg/L
Vinyl Chloride (VC)	0.009	mg/L
Carbon Tetrachloride (CT)		mg/L
Trichloromethane ( or chloroform) (CF)		mg/L
Dichloromethane (or methylene chloride) (MC)		mg/L
Chloromethane		mg/L
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L
Dichloroethane (1,1-DCA and 1,2-DCA)		mg/L
Chloroethane		mg/L
Perchlorate		mg/L

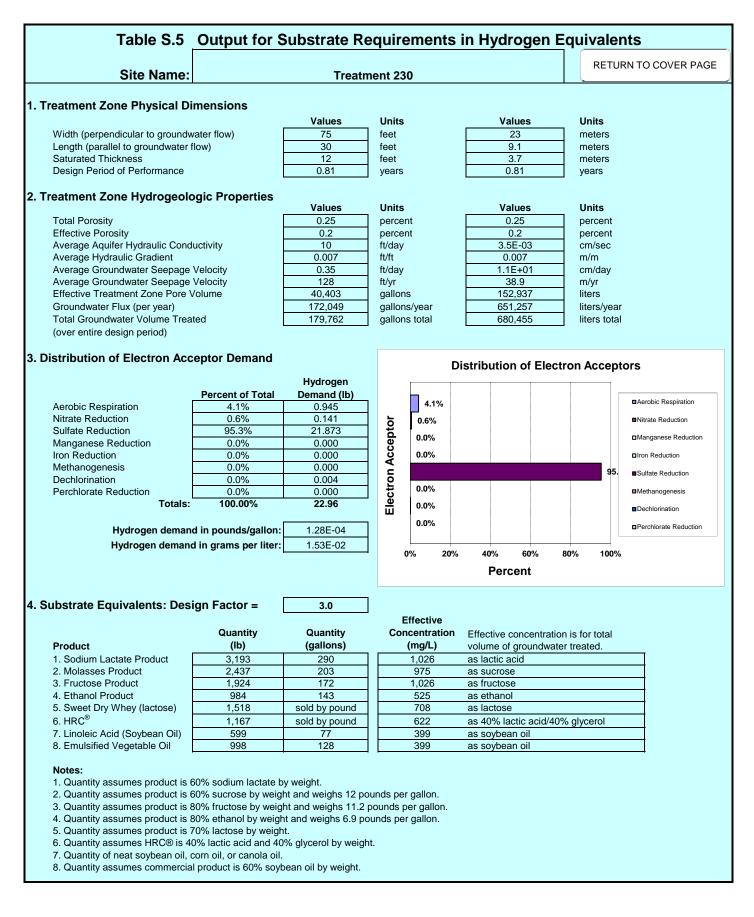
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



Treatment 230

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		290.2815873
2. Molasses Product		203.0697557
3. Fructose Product		171.809675
4. Ethanol Product		142.5968624
5. Sweet Dry Whey (lactose)	1517.735916	
6. HRC®	1166.704013	
7. Linoleic Acid (Soybean Oil)		76.80316387
8. Emulsified Vegetable Oil		128.0052731
9. Lactoil Product		61.4425311
10. Lactic Acid Product		0
11. Hydrogen Gas	22.96	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 231			RETURN TO COVER PAGE		
NOTE: Unshaded boxes are user input.						
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	50	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	170	1-1,000	feet			
Saturated Thickness	12	1-100	feet			
Treatment Zone Cross Sectional Area	600		ft <sup>2</sup>			
Treatment Zone Volume	102,000		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	190,791		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	152,633		gallons			
Design Period of Performance	2.0	.5 to 5	vear			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
	0.0	2 10 20	unidoco	Boldur - 0		
2. Treatment Zone Hydrogeologic Properties						
Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	10	.01-1000	ft/day			
Average Hydraulic Gradient	0.007	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.35		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	127.8		ft/yr			
Average Groundwater Discharge through the Treatment Zone	114,699		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors						
A. Aqueous-Phase Native Electron Acceptors						
Oxygen	5.0	0.01 to 10	mg/L	Default = 5		
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1		
Sulfate	104	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors	1					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

### 4. Contaminant Electron Acceptors

0.000	mg/L	
0.000	mg/L	
0.090	mg/L	
0.060	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.000	mg/L	
	mg/L	
	mg/L	
	0.000 0.090 0.060	0.000          mg/L           0.090          mg/L           0.060          mg/L            mg/L         mg/L

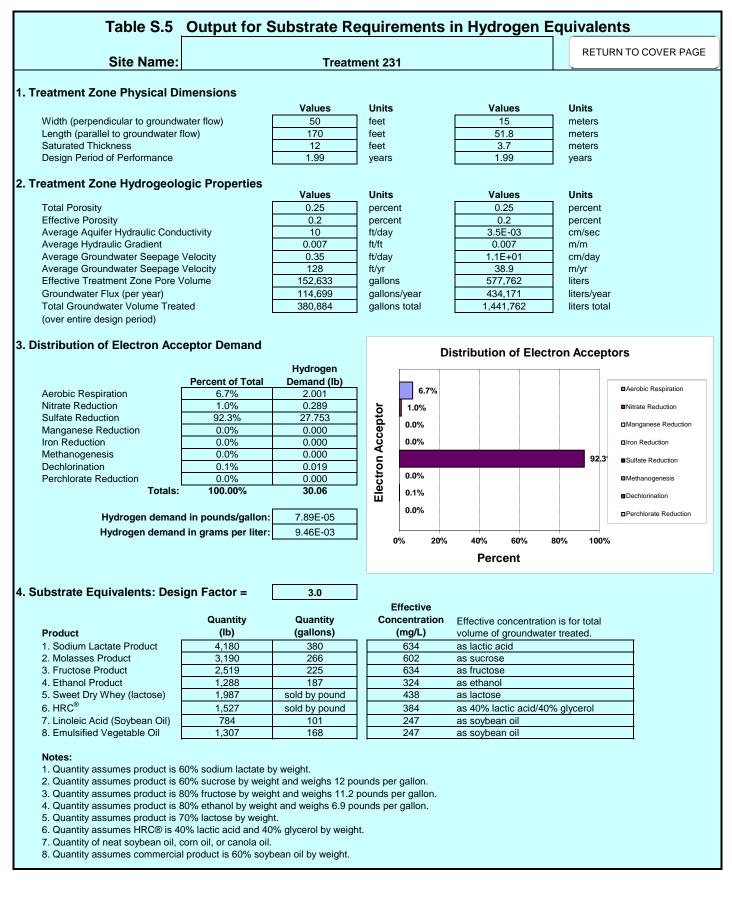
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



Treatment 231

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		380.0375407
2. Molasses Product		265.8595445
3. Fructose Product		224.9337513
4. Ethanol Product		186.6882477
5. Sweet Dry Whey (lactose)	1987.024498	
6. HRC®	1527.452457	
7. Linoleic Acid (Soybean Oil)		100.5509367
8. Emulsified Vegetable Oil		167.5848945
9. Lactoil Product		80.44074938
10. Lactic Acid Product		0
11. Hydrogen Gas	30.06	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for S	ubstrate Re	quiremen	ts in Hydro	ogen Equivalents
Site Name:	Treatment 233			RETURN TO COVER PAGE
	NOTE: Unshaded	boxes are use	r input.	
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes
Width (Perpendicular to predominant groundwater flow direction)	420	1-10,000	feet	
Length (Parallel to predominant groundwater flow)	655	1-1,000	feet	
Saturated Thickness	28.1	1-100	feet	
Treatment Zone Cross Sectional Area	11802		ft <sup>2</sup>	
Treatment Zone Volume	7,730,310		ft <sup>3</sup>	
Treatment Zone Total Pore Volume (total volume x total porosity)	14,459,545		gallons	
Treatment Zone Effective Pore Volume (total volume x effective porosity)	11,567,636		gallons	
Design Period of Performance	4.0	.5 to 5	year	
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3
2. Treatment Zone Hydrogeologic Properties				
Total Porosity	25%	.05-50	percent	Default = 25%
Effective Porosity	20%	.05-50	percent	Default = 20%
Average Aquifer Hydraulic Conductivity	50	.01-1000	ft/day	
Average Hydraulic Gradient	0.023	0.0001-0.1	ft/ft	
Average Groundwater Seepage Velocity through the Treatment Zone	5.75		ft/day	
Average Groundwater Seepage Velocity through the Treatment Zone	2098.8		ft/yr	
Average Groundwater Discharge through the Treatment Zone	37,065,001		gallons/year	
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%
3. Native Electron Acceptors <u>A. Aqueous-Phase Native Electron Acceptors</u>				
Oxygen	4.1	0.01 to 10	mg/L	Default = 5
Nitrate	5.40	0.1 to- 20	mg/L	Default = 1
Sulfate	102	10 to 5,000	mg/L	Default = 50
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0
B. Solid-Phase Native Electron Acceptors				
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0

### 4. Contaminant Electron Acceptors

0.002	mg/L	
5.087	mg/L	
0.020	mg/L	
0.000	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.000	mg/L	
0.000	mg/L	
	mg/L	
	mg/L	
	5.087 0.020 0.000 0.000	5.087          mg/L           0.020          mg/L           0.000          mg/L            mg/L             mg/L             mg/L             mg/L             mg/L             mg/L            0.000          mg/L           0.000          mg/L           0.000          mg/L           0.000          mg/L           0.000          mg/L

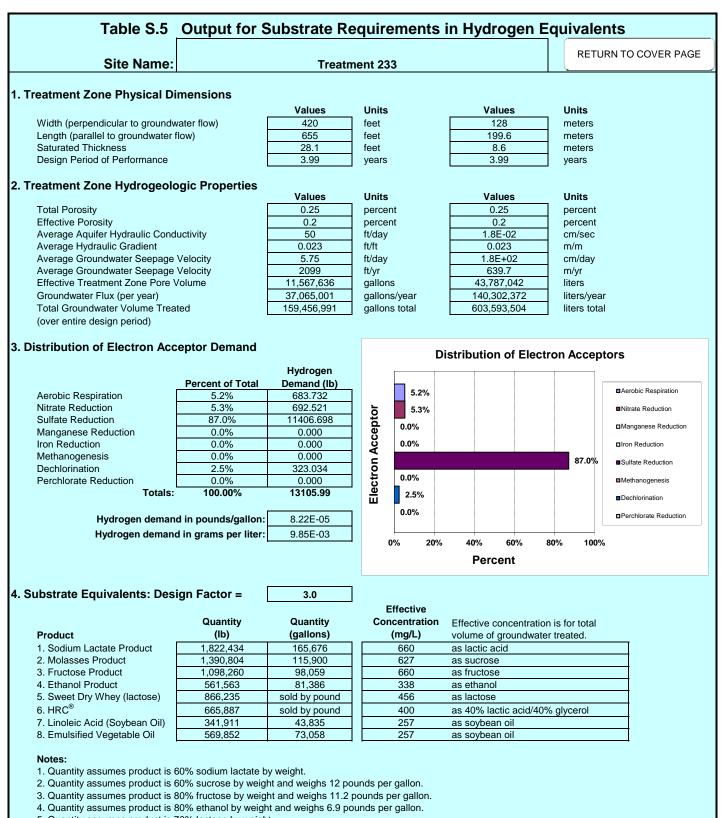
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.

### Treatment 233

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		165675.7964
2. Molasses Product		115900.3705
3. Fructose Product		98058.93996
4. Ethanol Product		81385.97063
5. Sweet Dry Whey (lactose)	866235.1238	
6. HRC®	665886.5908	
7. Linoleic Acid (Soybean Oil)		43834.76562
8. Emulsified Vegetable Oil		73057.94269
9. Lactoil Product		35067.81249
10. Lactic Acid Product		0
11. Hydrogen Gas	13105.99	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 234			RETURN TO COVER PAGE
	NOTE: Unshaded	hoxes are use	r input	
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes
Width (Perpendicular to predominant groundwater flow direction)	200	1-10,000	feet	
Length (Parallel to predominant groundwater flow)	240	1-1,000	feet	
Saturated Thickness	37	1-100	feet	
Treatment Zone Cross Sectional Area	7400		ft <sup>2</sup>	
Treatment Zone Volume	1,776,000		ft <sup>3</sup>	
Treatment Zone Total Pore Volume (total volume x total porosity)	3,322,008		gallons	
Treatment Zone Effective Pore Volume (total volume x effective porosity)	2,657,606		gallons	
Design Period of Performance	0.7	.5 to 5	year	
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3
2. Treatment Zone Hydrogeologic Properties				
Total Porosity	25%	.05-50	percent	Default = 25%
Effective Porosity	20%	.05-50	percent	Default = 20%
Average Aquifer Hydraulic Conductivity	50	.01-1000	ft/day	
Average Hydraulic Gradient	0.023	0.0001-0.1	ft/ft	
Average Groundwater Seepage Velocity through the Treatment Zone	5.75		ft/day	
Average Groundwater Seepage Velocity through the Treatment Zone	2098.8		ft/yr	
Average Groundwater Discharge through the Treatment Zone	23,240,214		gallons/year	
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors				
Oxygen	5.7	0.01 to 10	mg/L	Default = 5
Nitrate	2.44	0.1 to- 20	mg/L	Default = 1
Sulfate	214	10 to 5,000	mg/L	Default = 50
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0
B. Solid-Phase Native Electron Acceptors				
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0

### 4. Contaminant Electron Acceptors

0.004	mg/L	
12.580	mg/L	
0.037	mg/L	
0.002	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.000	mg/L	
0.000	mg/L	
	mg/L	
	mg/L	
	12.580 0.037 0.002 0.000	12.580          mg/L           0.037          mg/L           0.002          mg/L            mg/L             mg/L             mg/L             mg/L             mg/L            0.000          mg/L           0.000          mg/L           0.000          mg/L           0.000          mg/L

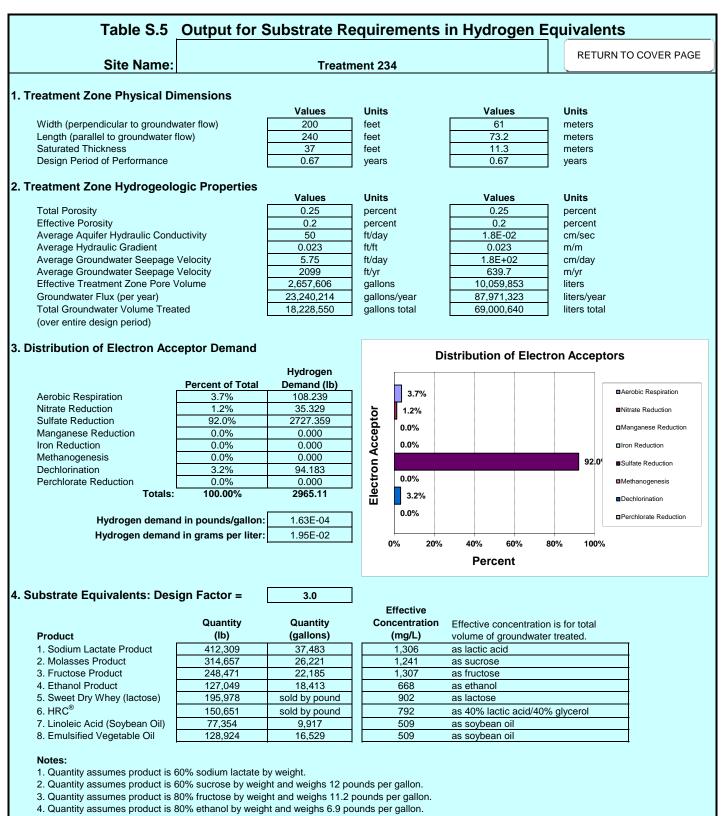
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.

### Treatment 234

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		37482.64734
2. Molasses Product		26221.40836
3. Fructose Product		22184.94641
4. Ethanol Product		18412.83821
5. Sweet Dry Whey (lactose)	195977.8457	
6. HRC®	150650.8059	1
7. Linoleic Acid (Soybean Oil)		9917.218425
8. Emulsified Vegetable Oil		16528.69738
9. Lactoil Product		7933.77474
10. Lactic Acid Product		0
11. Hydrogen Gas	2965.11	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 235			RETURN TO COVER PAGE
	NOTE: Unshaded	boxes are use	r input.	
. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes
Width (Perpendicular to predominant groundwater flow direction)	200	1-10,000	feet	
Length (Parallel to predominant groundwater flow)	240	1-1,000	feet	
Saturated Thickness	37	1-100	feet	
Treatment Zone Cross Sectional Area	7400		ft <sup>2</sup>	
Treatment Zone Volume	1,776,000		ft <sup>3</sup>	
Treatment Zone Total Pore Volume (total volume x total porosity)	3,322,008		gallons	
Treatment Zone Effective Pore Volume (total volume x effective porosity)	2,657,606		gallons	
Design Period of Performance	3.0	.5 to 5	year	
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3
. Treatment Zone Hydrogeologic Properties	25%	.05-50	percent	Default = 25%
Effective Porosity	20%	.05-50	percent	Default = 20%
Average Aquifer Hydraulic Conductivity	50	.01-1000	ft/day	
Average Hydraulic Gradient	0.023	0.0001-0.1	ft/ft	
Average Groundwater Seepage Velocity through the Treatment Zone	5.75		ft/day	
Average Groundwater Seepage Velocity through the Treatment Zone	2098.8		ft/yr	
Average Groundwater Discharge through the Treatment Zone	23,240,214		gallons/year	
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%
Native Electron Acceptors     A. Aqueous-Phase Native Electron Acceptors	2.9	0.01 to 10		Default = 5
Oxygen Nitrate	0.00	0.01 to 10 0.1 to- 20	mg/L	Default = 5
Sulfate	47		mg/L	Default = 1 Default = 50
		10 to 5,000	mg/L	
Carbon Dioxide (estimated as the amount of Methane produced) B. Solid-Phase Native Electron Acceptors	0.0	0.1 to 20	mg/L	Default = 0
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0

### 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L	
Trichloroethene (TCE)	7.509	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	4.877	mg/L	
Vinyl Chloride (VC)	0.004	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)	0.000	mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)	0.000	mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

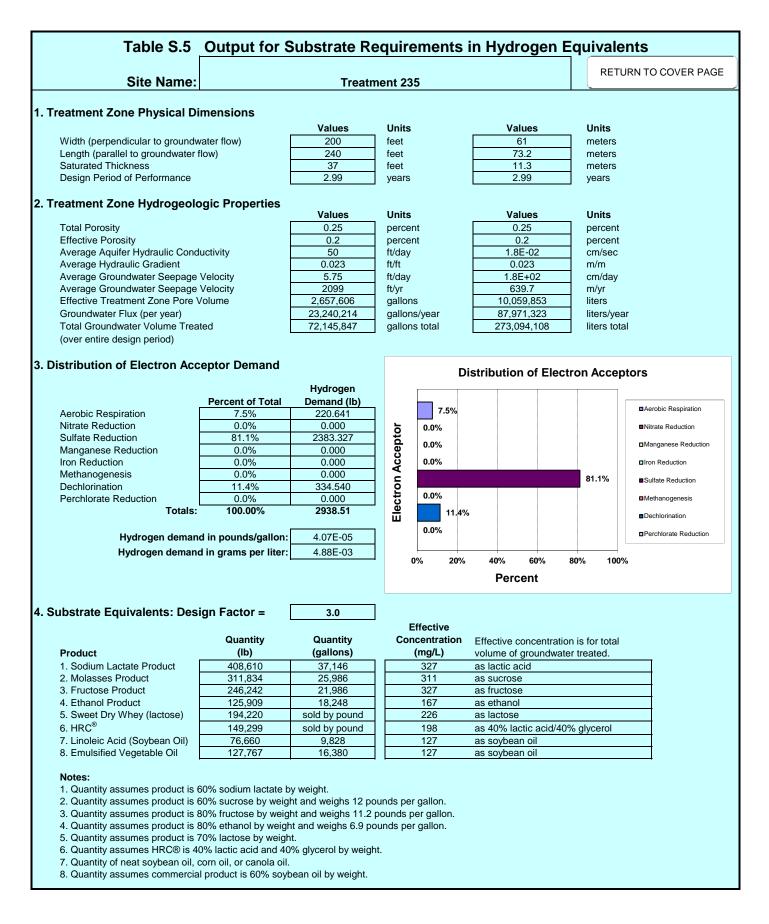
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 µs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



### Treatment 235

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		37146.36247
2. Molasses Product		25986.15649
3. Fructose Product		21985.90866
4. Ethanol Product		18247.64287
5. Sweet Dry Whey (lactose)	194219.5818	
6. HRC®	149299.2048	
7. Linoleic Acid (Soybean Oil)		9828.243639
8. Emulsified Vegetable Oil		16380.40607
9. Lactoil Product		7862.594912
10. Lactic Acid Product		0
11. Hydrogen Gas	2938.51	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for S		•			
Site Name:	Treatment 237			RETURN TO COVER PAGE	
NOTE: Unshaded boxes are user input.					
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes	
Width (Perpendicular to predominant groundwater flow direction)	80	1-10,000	feet		
Length (Parallel to predominant groundwater flow)	80	1-1,000	feet		
Saturated Thickness	11.5	1-100	feet		
Treatment Zone Cross Sectional Area	920		ft <sup>2</sup>		
Treatment Zone Volume	73,600		ft <sup>3</sup>		
Treatment Zone Total Pore Volume (total volume x total porosity)	137,669		gallons		
Treatment Zone Effective Pore Volume (total volume x effective porosity)	110,135		gallons		
Design Period of Performance	2.0	.5 to 5	year		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
2. Treatment Zone Hydrogeologic Properties					
Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity	20%	.05-50	percent	Default = 20%	
Average Aquifer Hydraulic Conductivity	2.4	.01-1000	ft/day		
Average Hydraulic Gradient	0.002	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	0.02		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	8.8		ft/yr		
Average Groundwater Discharge through the Treatment Zone	12,060		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors					
Oxygen	0.5	0.01 to 10	mg/L	Default = 5	
Nitrate	0.13	0.1 to- 20	mg/L	Default = 1	
Sulfate	803	10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced)     0.0     0.1 to 20     mg/L     Default = 0					
B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

### 4. Contaminant Electron Acceptors

0.968	mg/L	
0.185	mg/L	
0.068	mg/L	
0.001	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.000	mg/L	
0.050	mg/L	
	mg/L	
	mg/L	
	0.185 0.068 0.001	0.185          mg/L           0.068          mg/L           0.001          mg/L            mg/L             mg/L             mg/L             mg/L             mg/L            0.000          mg/L           0.050          mg/L            mg/L

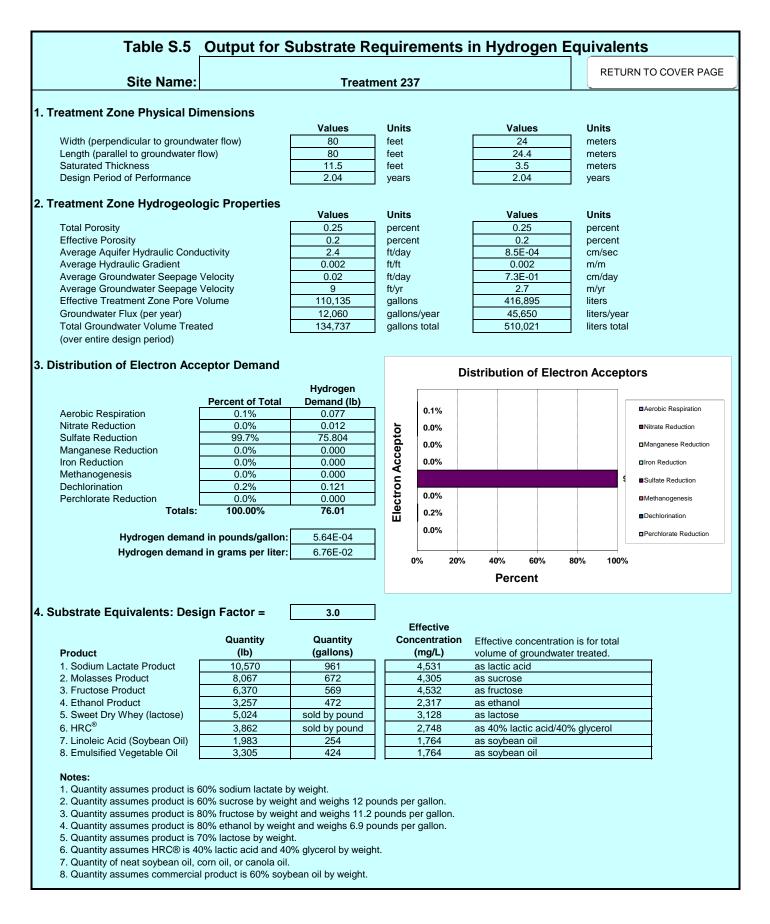
## 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



### Treatment 237

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		960.9157161
2. Molasses Product		672.2194183
3. Fructose Product		568.7395417
4. Ethanol Product		472.0367123
5. Sweet Dry Whey (lactose)	5024.143311	
6. HRC®	3862.126539	
7. Linoleic Acid (Soybean Oil)		254.2406079
8. Emulsified Vegetable Oil		423.7343465
9. Lactoil Product		203.3924863
10. Lactic Acid Product		0
11. Hydrogen Gas	76.01	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents						
Site Name:	Treatment 238			RETURN TO COVER PAGE		
	NOTE: Unshaded	boxes are use	r input.			
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	105	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	110	1-1,000	feet			
Saturated Thickness	11.5	1-100	feet			
Treatment Zone Cross Sectional Area	1207.5		ft <sup>2</sup>			
Treatment Zone Volume	132,825		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	248,449		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	198,759		gallons			
Design Period of Performance	1.4	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
2. Treatment Zone Hydrogeologic Properties	Γ					
Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	2.4	.01-1000	ft/day			
Average Hydraulic Gradient	0.002	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.02		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	8.8		ft/yr			
Average Groundwater Discharge through the Treatment Zone	15,828		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors						
Oxygen	1.8	0.01 to 10	mg/L	Default = 5		
Nitrate	0.12	0.1 to- 20	mg/L	Default = 1		
Sulfate	485	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors						
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

# 4. Contaminant Electron Acceptors

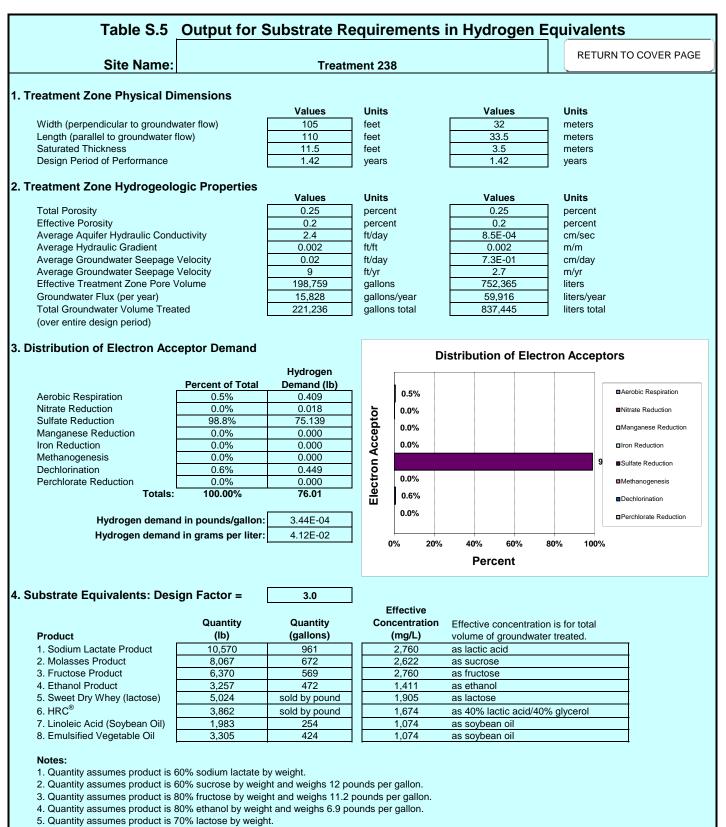
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

8. Quantity assumes commercial product is 60% soybean oil by weight.

# Treatment 238

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		960.9142187
2. Molasses Product		672.2183708
3. Fructose Product		568.7386554
4. Ethanol Product		472.0359767
5. Sweet Dry Whey (lactose)	5024.135482	
6. HRC®	3862.120521	
7. Linoleic Acid (Soybean Oil)		254.2402117
8. Emulsified Vegetable Oil		423.7336862
9. Lactoil Product		203.3921694
10. Lactic Acid Product		0
11. Hydrogen Gas	76.01	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 239			RETURN TO COVER PAGE		
	NOTE: Unshaded		•			
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	77	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	48	1-1,000	feet			
Saturated Thickness	11.5	1-100	feet			
Treatment Zone Cross Sectional Area	885.5		ft <sup>2</sup>			
Treatment Zone Volume	42,504		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	79,504		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	63,603		gallons			
Design Period of Performance	4.3	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
2. Treatment Zone Hydrogeologic Properties						
Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	2.4	.01-1000	ft/day			
Average Hydraulic Gradient	0.002	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	0.02		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	8.8		ft/yr			
Average Groundwater Discharge through the Treatment Zone	11,608		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors	Γ	1				
Oxygen	1.8	0.01 to 10	mg/L	Default = 5		
Nitrate	0.00	0.1 to- 20	mg/L	Default = 1		
Sulfate	138	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors						
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	1.417	mg/L	
Trichloroethene (TCE)	0.276	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.435	mg/L	
Vinyl Chloride (VC)	0.008	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)	0.000	mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)	0.040	mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

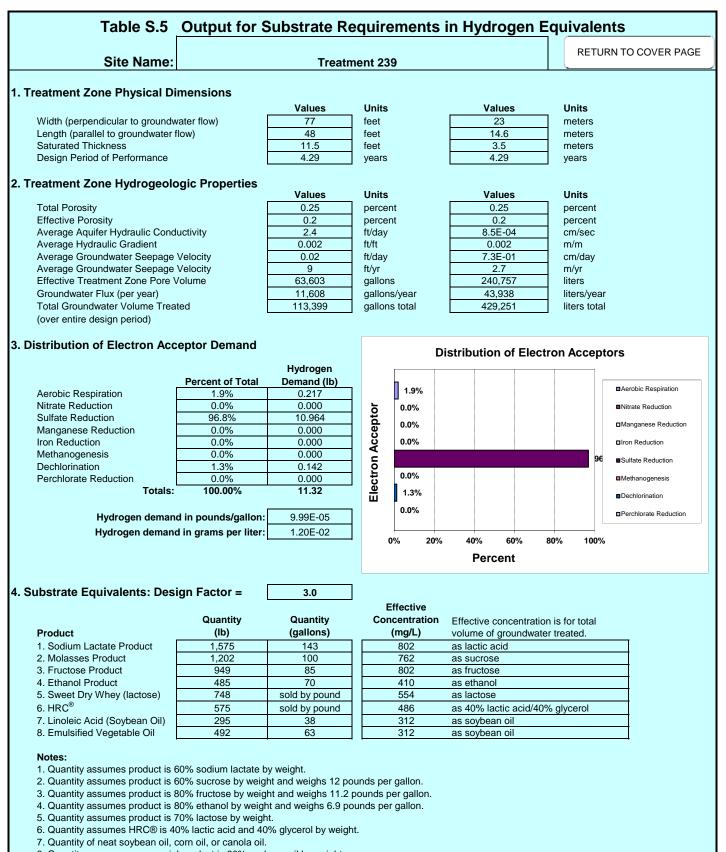
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 μs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as $CaCO_3$	Default = 10%



## Treatment 239

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		143.1374265
2. Molasses Product		100.1333998
3. Fructose Product		84.71909968
4. Ethanol Product		70.31430444
5. Sweet Dry Whey (lactose)	748.3933624	
6. HRC®	575.3000438	
7. Linoleic Acid (Soybean Oil)		37.87152787
8. Emulsified Vegetable Oil		63.11921311
9. Lactoil Product		30.29722229
10. Lactic Acid Product		0
11. Hydrogen Gas	11.32	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 241			RETURN TO COVER PAGE		
	NOTE: Unshaded	hoxes are use	r input			
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes		
Width (Perpendicular to predominant groundwater flow direction)	80	1-10,000	feet			
Length (Parallel to predominant groundwater flow)	130	1-1,000	feet			
Saturated Thickness	12	1-100	feet			
Treatment Zone Cross Sectional Area	960		ft <sup>2</sup>			
Treatment Zone Volume	124,800		ft <sup>3</sup>			
Treatment Zone Total Pore Volume (total volume x total porosity)	233,438		gallons			
Treatment Zone Effective Pore Volume (total volume x effective porosity)	186,751		gallons			
Design Period of Performance	2.9	.5 to 5	year			
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3		
2. Treatment Zone Hydrogeologic Properties		_				
Total Porosity	25%	.05-50	percent	Default = 25%		
Effective Porosity	20%	.05-50	percent	Default = 20%		
Average Aquifer Hydraulic Conductivity	72	.01-1000	ft/day			
Average Hydraulic Gradient	0.007	0.0001-0.1	ft/ft			
Average Groundwater Seepage Velocity through the Treatment Zone	2.52		ft/day			
Average Groundwater Seepage Velocity through the Treatment Zone	919.8		ft/yr			
Average Groundwater Discharge through the Treatment Zone	1,321,333		gallons/year			
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7		
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%		
3. Native Electron Acceptors						
A. Aqueous-Phase Native Electron Acceptors	r	1				
Oxygen	5.0	0.01 to 10	mg/L	Default = 5		
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1		
Sulfate	46	10 to 5,000	mg/L	Default = 50		
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0		
B. Solid-Phase Native Electron Acceptors						
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0		
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0		

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)         0.000          mg/L           Trichloroethene (TCE)         12.500          mg/L           Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)         1.010          mg/L           Vinyl Chloride (VC)         0.000          mg/L           Carbon Tetrachloride (CT)          mg/L           Trichloromethane ( or chloroform) (CF)          mg/L           Dichloromethane ( or methylene chloride) (MC)          mg/L           Chloromethane (1,1,1,2-PCA and 1,1,2,2-PCA)          mg/L           Trichloroethane (1,1,1-TCA and 1,1,2,2-PCA)          mg/L           Dichloroethane (1,1,1-DCA and 1,2-DCA)          mg/L           Dichloroethane (1,1-DCA and 1,2-DCA)          mg/L           Chloroethane (1,1-DCA and 1,2-DCA)          mg/L           Dichloroethane (1,1-DCA and 1,2-DCA)          mg/L           Chloroethane (1,1-DCA and 1,2-DCA)          mg/L           Chloroethane          mg/L				
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)1.010mg/LVinyl Chloride (VC)0.000mg/LCarbon Tetrachloride (CT)mg/LTrichloromethane ( or chloroform) (CF)mg/LDichloromethane (or methylene chloride) (MC)mg/LChloromethanemg/LTetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)mg/LTrichloroethane (1,1,1-TCA and 1,1,2-TCA)mg/LDichloroethane (1,1-DCA and 1,2-DCA)mg/LChloroethanemg/LChloroethanemg/LDichloroethanemg/LDichloroethanemg/LDichloroethanemg/LDichloroethanemg/L	Tetrachloroethene (PCE)	0.000	mg/L	
Vinyl Chloride (VC)0.000mg/LCarbon Tetrachloride (CT)mg/LTrichloromethane ( or chloroform) (CF)mg/LDichloromethane (or methylene chloride) (MC)mg/LChloromethanemg/LChloromethanemg/LTetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)mg/LTrichloroethane (1,1,1-TCA and 1,1,2-TCA)mg/LDichloroethane (1,1-DCA and 1,2-DCA)mg/LChloroethanemg/LChloroethanemg/LDichloroethanemg/LDichloroethanemg/LDichloroethanemg/L	Trichloroethene (TCE)	12.500	mg/L	
Carbon Tetrachloride (CT)mg/LTrichloromethane ( or chloroform) (CF)mg/LDichloromethane (or methylene chloride) (MC)mg/LChloromethanemg/LTetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)mg/LTrichloroethane (1,1,1-TCA and 1,1,2-TCA)mg/LDichloroethane (1,1-DCA and 1,2-DCA)mg/LChloroethanemg/LChloroethanemg/LDichloroethanemg/LDichloroethanemg/LDichloroethanemg/LDichloroethanemg/LChloroethanemg/L	Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	1.010	mg/L	
Trichloromethane ( or chloroform) (CF)mg/LDichloromethane (or methylene chloride) (MC)mg/LChloromethanemg/LTetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)mg/LTrichloroethane (1,1,1-TCA and 1,1,2-TCA)mg/LDichloroethane (1,1-DCA and 1,2-DCA)mg/LChloroethane (1,1-DCA and 1,2-DCA)mg/LChloroethanemg/LChloroethanemg/L	Vinyl Chloride (VC)	0.000	mg/L	
Dichloromethane (or methylene chloride) (MC)mg/LChloromethanemg/LTetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)mg/LTrichloroethane (1,1,1-TCA and 1,1,2-TCA)mg/LDichloroethane (1,1-DCA and 1,2-DCA)mg/LChloroethanemg/LChloroethanemg/L	Carbon Tetrachloride (CT)		mg/L	
Chloromethanemg/LTetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)mg/LTrichloroethane (1,1,1-TCA and 1,1,2-TCA)mg/LDichloroethane (1,1-DCA and 1,2-DCA)mg/LChloroethanemg/L	Trichloromethane ( or chloroform) (CF)		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)mg/LTrichloroethane (1,1,1-TCA and 1,1,2-TCA)mg/LDichloroethane (1,1-DCA and 1,2-DCA)mg/LChloroethanemg/L	Dichloromethane (or methylene chloride) (MC)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)      mg/L       Dichloroethane (1,1-DCA and 1,2-DCA)      mg/L       Chloroethane      mg/L	Chloromethane		mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)      mg/L       Chloroethane      mg/L	Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Chloroethane mg/L	Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L	
	Dichloroethane (1,1-DCA and 1,2-DCA)		mg/L	
	Chloroethane		mg/L	
Perchlorate mg/L	Perchlorate		mg/L	

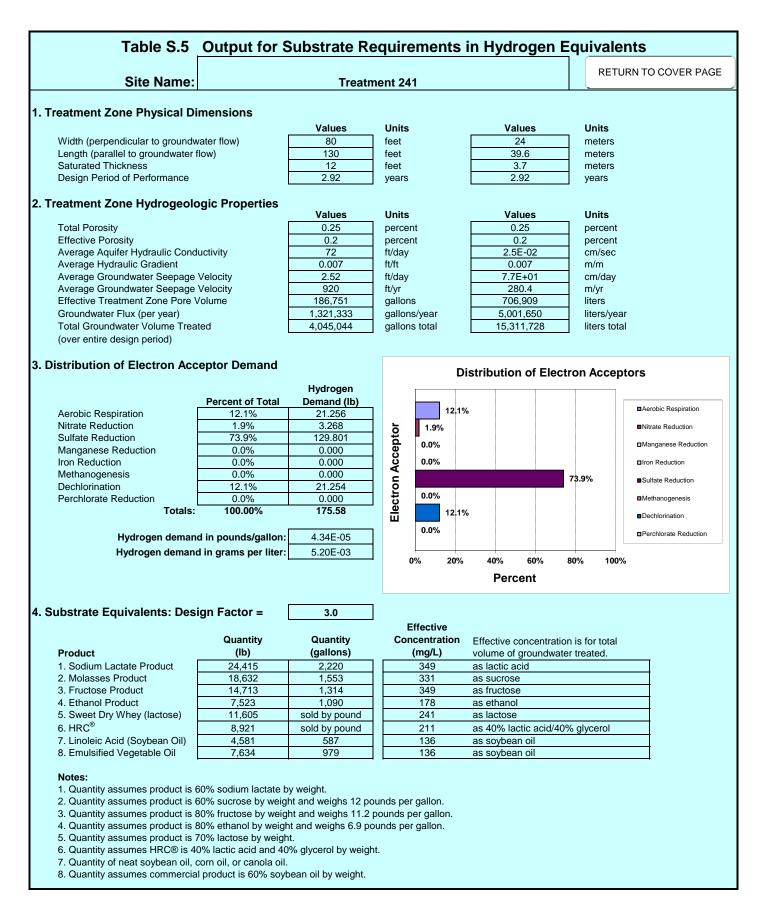
# 5. Aquifer Geochemistry (Optional Screening Parameters)

### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 µs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

#### **B. Aquifer Matrix**

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



# Treatment 241

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		2219.538179
2. Molasses Product		1552.702947
3. Fructose Product		1313.683506
4. Ethanol Product		1090.317795
5. Sweet Dry Whey (lactose)	11604.84495	i
6. HRC®	8920.800402	
7. Linoleic Acid (Soybean Oil)		587.2489402
8. Emulsified Vegetable Oil		978.7482337
9. Lactoil Product		469.7991522
10. Lactic Acid Product		0
11. Hydrogen Gas	175.58	;

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 243			RETURN TO COVER PAGE
Olic Maille.		<b>b</b> avaa ava avaa		
1. Treatment Zone Physical Dimensions	NOTE: Unshaded Values		Units	User Notes
		Range		User notes
Width (Perpendicular to predominant groundwater flow direction)	180	1-10,000	feet	
Length (Parallel to predominant groundwater flow)	25	1-1,000	feet	
Saturated Thickness	6.5	1-100	feet ft <sup>2</sup>	
Treatment Zone Cross Sectional Area	1170		π ft <sup>3</sup>	
Treatment Zone Volume	29,250			
Treatment Zone Total Pore Volume (total volume x total porosity)	54,712		gallons	
Treatment Zone Effective Pore Volume (total volume x effective porosity)	43,770		gallons	
Design Period of Performance	5.3	.5 to 5	year	
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3
2. Treatment Zone Hydrogeologic Properties				
Total Porosity	25%	.05-50	percent	Default = 25%
Effective Porosity	20%	.05-50	percent	Default = 20%
Average Aquifer Hydraulic Conductivity	1	.01-1000	ft/day	
Average Hydraulic Gradient	0.026	0.0001-0.1	ft/ft	
Average Groundwater Seepage Velocity through the Treatment Zone	0.13		ft/day	
Average Groundwater Seepage Velocity through the Treatment Zone	47.5		ft/yr	
Average Groundwater Discharge through the Treatment Zone	83,075		gallons/year	
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors				
Oxygen	2.1	0.01 to 10	mg/L	Default = 5
Nitrate	0.11	0.1 to- 20	mg/L	Default = 1
Sulfate	382	10 to 5,000	mg/L	Default = 50
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0
B. Solid-Phase Native Electron Acceptors				
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0

# 4. Contaminant Electron Acceptors

0.000	mg/L	
0.134	mg/L	
0.010	mg/L	
0.133	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.000	mg/L	
0.000	mg/L	
	mg/L	
	mg/L	
	0.134 0.010 0.133 0.000	0.134          mg/L           0.010          mg/L           0.133          mg/L           0.133          mg/L            mg/L             mg/L             mg/L             mg/L             mg/L            0.000          mg/L           0.000          mg/L           0.000          mg/L           0.000          mg/L

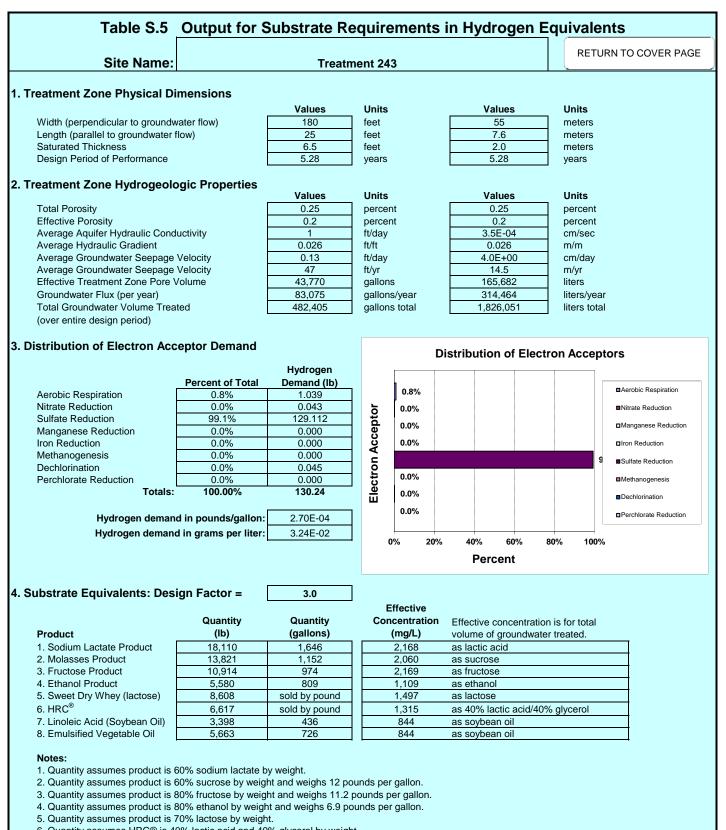
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

/ inducedo eccenteninou y	
Oxidation-Reduction Potential (ORP)	-400 to +500 mV
Temperature	5.0 to 30 °C
рН	4.0 to 10.0 su
Alkalinity	10 to 1,000 mg/L
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L
Specific Conductivity	100 to 10,000 µs/cm
Chloride	10 to 10,000 mg/L
Sulfide - Pre injection	0.1 to 100 mg/L
Sulfide - Post injection	0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

8. Quantity assumes commercial product is 60% soybean oil by weight.

## Treatment 243

Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		1646.375036
2. Molasses Product		1151.740211
3. Fructose Product		974.4440307
4. Ethanol Product		808.7592349
5. Sweet Dry Whey (lactose)	8608.064148	
6. HRC®	6617.134692	
7. Linoleic Acid (Soybean Oil)		435.6005246
8. Emulsified Vegetable Oil		726.0008743
9. Lactoil Product		348.4804197
10. Lactic Acid Product		0
11. Hydrogen Gas	130.24	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Table S.1 Input for Substrate Requirements in Hydrogen Equivalents					
Site Name:	Treatment 245			RETURN TO COVER PAGE	
	NOTE: Unshaded	boxes are use	r input.		
1. Treatment Zone Physical Dimensions	Values	Range	Units	User Notes	
Width (Perpendicular to predominant groundwater flow direction)	50	1-10,000	feet		
Length (Parallel to predominant groundwater flow)	50	1-1,000	feet		
Saturated Thickness	33	1-100	feet		
Treatment Zone Cross Sectional Area	1650		ft <sup>2</sup>		
Treatment Zone Volume	82,500		ft <sup>3</sup>		
Treatment Zone Total Pore Volume (total volume x total porosity)	154,316		gallons		
Treatment Zone Effective Pore Volume (total volume x effective porosity)	123,453		gallons		
Design Period of Performance	1.1	.5 to 5	year		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
2. Treatment Zone Hydrogeologic Properties					
Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity	20%	.05-50	percent	Default = 20%	
Average Aquifer Hydraulic Conductivity	6.9	.01-1000	ft/day		
Average Hydraulic Gradient	0.003	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	0.10		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	37.8		ft/yr		
Average Groundwater Discharge through the Treatment Zone	93,275		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
3. Native Electron Acceptors <u>A. Aqueous-Phase Native Electron Acceptors</u>	r				
Oxygen	5.0	0.01 to 10	mg/L	Default = 5	
Nitrate	1.00	0.1 to- 20	mg/L	Default = 1	
Sulfate	293	10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0	
B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

# 4. Contaminant Electron Acceptors

Tetrachloroethene (PCE)	0.000	mg/L	
Trichloroethene (TCE)	3.910	mg/L	
Dichloroethene (cis-DCE, trans-DCE, and 1,1-DCE)	0.013	mg/L	
Vinyl Chloride (VC)	0.000	mg/L	
Carbon Tetrachloride (CT)		mg/L	
Trichloromethane ( or chloroform) (CF)		mg/L	
Dichloromethane (or methylene chloride) (MC)		mg/L	
Chloromethane		mg/L	
Tetrachloroethane (1,1,1,2-PCA and 1,1,2,2-PCA)		mg/L	
Trichloroethane (1,1,1-TCA and 1,1,2-TCA)		mg/L	
Dichloroethane (1,1-DCA and 1,2-DCA)	0.000	mg/L	
Chloroethane		mg/L	
Perchlorate		mg/L	

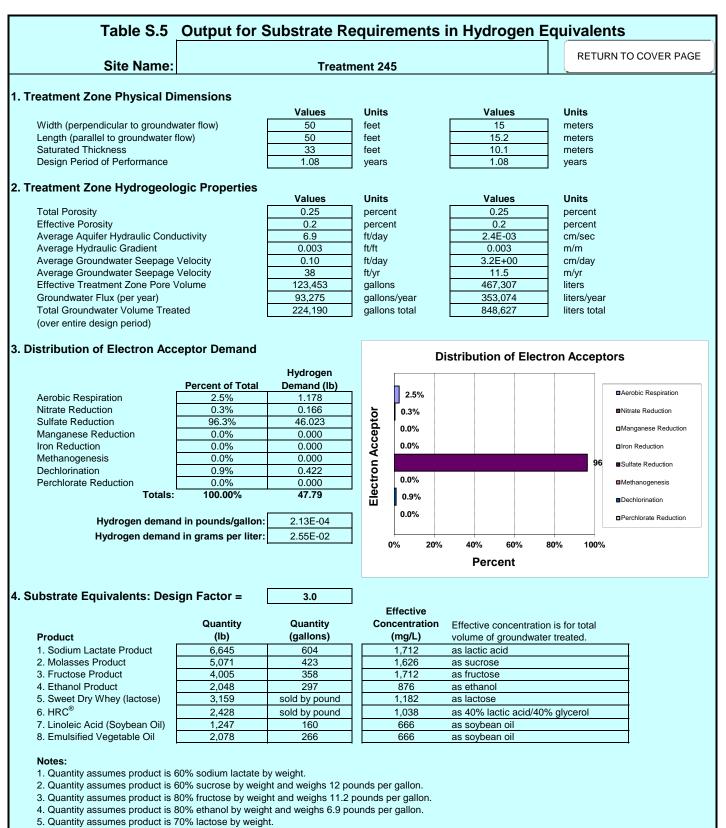
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

-400 to +500 mV
5.0 to 30 °C
4.0 to 10.0 su
10 to 1,000 mg/L
10 to 1,000 mg/L
100 to 10,000 µs/cm
10 to 10,000 mg/L
0.1 to 100 mg/L
0.1 to 100 mg/L

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.

7. Quantity of neat soybean oil, corn oil, or canola oil.

8. Quantity assumes commercial product is 60% soybean oil by weight.

## Treatment 245

-		
Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		604.1092497
2. Molasses Product		422.6114337
3. Fructose Product		357.5556233
4. Ethanol Product		296.7604123
5. Sweet Dry Whey (lactose)	3158.582376	
6. HRC®	2428.044756	
7. Linoleic Acid (Soybean Oil)		159.8361857
8. Emulsified Vegetable Oil		266.3936429
9. Lactoil Product		127.8689486
10. Lactic Acid Product		0
11. Hydrogen Gas	47.79	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.

Site Name:	Treatment 246			RETURN TO COVER PAGE	
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. Treatment Zone Physical Dimensions	Values	NOTE: Unshaded boxes are user input. Values Range Units		User Notes	
Width (Perpendicular to predominant groundwater flow direction)	80	1-10,000	feet		
Length (Parallel to predominant groundwater flow)	80	1-1,000	feet		
Saturated Thickness	10.5	1-1,000	feet		
Treatment Zone Cross Sectional Area	840		ft <sup>2</sup>		
Treatment Zone Volume	67,200		ft <sup>3</sup>		
Treatment Zone Volume (total volume x total porosity)	125,698		gallons		
Treatment Zone Effective Pore Volume (total volume x total porosity)	100,558		gallons		
Design Period of Performance	1.8	.5 to 5	vear		
Design Factor (times the electron acceptor hydrogen demand)	3.0	2 to 20	unitless	Default = 3	
	0.0	2 10 20	unitess	Deladit = 3	
. Treatment Zone Hydrogeologic Properties		_			
Total Porosity	25%	.05-50	percent	Default = 25%	
Effective Porosity	20%	.05-50	percent	Default = 20%	
Average Aquifer Hydraulic Conductivity	29.5	.01-1000	ft/day		
Average Hydraulic Gradient	0.002	0.0001-0.1	ft/ft		
Average Groundwater Seepage Velocity through the Treatment Zone	0.30		ft/day		
Average Groundwater Seepage Velocity through the Treatment Zone	107.7		ft/yr		
Average Groundwater Discharge through the Treatment Zone	135,345		gallons/year		
Soil Bulk Density	1.7	1.4-2.0	gm/cm <sup>3</sup>	Default = 1.7	
Soil Fraction Organic Carbon (foc)	0.05%	0.01-10	percent	Default = 0.05%	
3. Native Electron Acceptors A. Aqueous-Phase Native Electron Acceptors					
Oxygen	0.4	0.01 to 10	mg/L	Default = 5	
Nitrate	0.05	0.1 to- 20	mg/L	Default = 1	
Sulfate	524	10 to 5,000	mg/L	Default = 50	
Carbon Dioxide (estimated as the amount of Methane produced)	0.0	0.1 to 20	mg/L	Default = 0	
B. Solid-Phase Native Electron Acceptors					
Manganese (IV) (estimated as the amount of Mn (II) produced)		0.1 to 20	mg/L	Default = 0	
Iron (III) (estimated as the amount of Fe (II) produced)		0.1 to 20	mg/L	Default = 0	

# 4. Contaminant Electron Acceptors

0.002	mg/L	
0.109	mg/L	
0.010	mg/L	
0.000	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
	mg/L	
0.000	mg/L	
0.000	mg/L	
	mg/L	
	mg/L	
	0.109 0.010 0.000	0.109          mg/L           0.010          mg/L           0.000          mg/L            mg/L             mg/L             mg/L             mg/L             mg/L            0.000          mg/L           0.000          mg/L           0.000          mg/L           0.000          mg/L

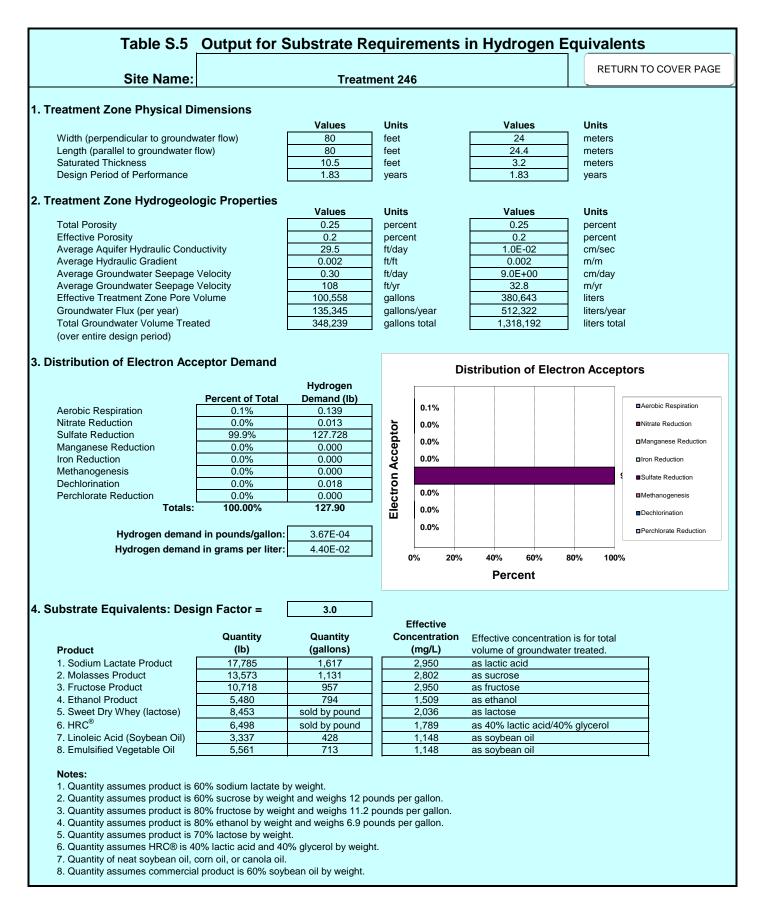
# 5. Aquifer Geochemistry (Optional Screening Parameters)

#### A. Aqueous Geochemistry

Oxidation-Reduction Potential (ORP)	-400 to +500 mV	
Temperature	5.0 to 30 °C	
рН	4.0 to 10.0 su	
Alkalinity	10 to 1,000 mg/L	
Total Dissolved Solids (TDS, or salinity)	10 to 1,000 mg/L	
Specific Conductivity	100 to 10,000 µs/cm	
Chloride	10 to 10,000 mg/L	
Sulfide - Pre injection	0.1 to 100 mg/L	
Sulfide - Post injection	0.1 to 100 mg/L	

### B. Aquifer Matrix

Total Iron	10000	200 to 20,000 mg/kg	Default = 10,000
Cation Exchange Capacity	NA	1.0 to 10 meq/100 g	
Neutralization Potential	10.0%	1.0 to 100 Percent as CaCO <sub>3</sub>	Default = 10%



## Treatment 246

-		
Substrate	Quantity(lb)	Quantity (gallons)
1. Sodium Lactate Product		1616.794802
2. Molasses Product		1131.047025
3. Fructose Product		956.9363051
4. Ethanol Product		794.2283492
5. Sweet Dry Whey (lactose)	8453.404034	
6. HRC®	6498.245382	
7. Linoleic Acid (Soybean Oil)		427.7741393
8. Emulsified Vegetable Oil		712.9568988
9. Lactoil Product		342.2193114
10. Lactic Acid Product		0
11. Hydrogen Gas	127.90	

- 1. Quantity assumes product is 60% sodium lactate by weight.
- 2. Quantity assumes product is 60% sucrose by weight and weighs 12 pounds per gallon.
- 3. Quantity assumes product is 80% fructose by weight and weighs 11.2 pounds per gallon.
- 4. Quantity assumes product is 80% ethanol by weight and weighs 6.9 pounds per gallon.
- 5. Quantity assumes product is 70% lactose by weight.
- 6. Quantity assumes HRC® is 40% lactic acid and 40% glycerol by weight.
- 7. Quantity of neat soybean oil, corn oil, or canola oil.
- 8. Quantity assumes commercial product is 60% soybean oil by weight.
- 9. Quantity assumes commercial product is 80% soybean oil by weight.
- 10. Quantity assumes commercial product is 88% lactic acid by weight.
- 11. Quantity assumes hydrogen gas substrate is 100% by weight.